Final Report To

COPPER RIVER AHTNA INTERTRIBAL NATURAL RESOURCE CONSERVATION DISTRICT

for

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Executive Summary

To be added post review and comments

Introduction

Purpose

Copper River Ahtna Intertribal natural resource conservation district (CRITR) was awarded a Conservation Innovation Grant (CIG) from the USDA Natural Resources Conservation Service to develop an ecologically based resource assessment across all of Ahtna Lands. CRITR was established to link the two land-owning corporations (Ahtna, Inc., and Chitina Native Corporation) with the Ahtna Tribes to promote Tribal stewardship of subsistence resources including an integrated approach to food production through habitat enhancement, biomass energy production, and wildfire protection. The Ahtna lands lack many site-specific tools that have been developed and applied in many other locations such as widely applied ecological site descriptions for supporting ecologically-based implementation of treatments. Landscape-scale resource assessments that can help support management decisions for subsistence food and biomass energy production are lacking. For these reasons, CRITR initiated a project with the objective of conducting a landscape assessment to provide the foundation for better planning and landscape-level tools to facilitate ecologically sustainable subsistence food and biomass energy production. In addition, the project reviewed fire planning at landscape scales. Wildfire has been aggressively suppressed from the landscape in the Copper River Basin for the past 40 years and has reduced the historically diverse vegetation mosaic produced by fires. Such changes can result in reduced moose habitat quality. CRITR is working to develop treatments for habitat enhancement to produce sustainable subsistence moose harvests, maintain caribou habitat, generate biomass for energy in the form of wood chips and firewood, and develop landscape-scale fire plans. These initiatives will produce an integrated ecological approach to securing local food, energy, wildlife benefits, and fire safe communities for the region.

A planned output of the CIG project is to develop 10-year plans for vegetation treatments focusing on the 8 tribal communities and surrounding Ahtna lands within the Ahtna Traditional Use Territory. These plans will build from the landscape assessment and will integrate the objectives of expanding the role of wildfire in desired outlying areas, improving moose habitat, producing biomass for use by the communities, and protecting high value caribou habitat.

Objectives

The objectives of the 10-year management plan are:

- Improve moose habitat through mechanical treatments
- Evaluate and recommend an expanded use of prescribed burning or let-burn wildfire areas,
- Increase opportunities for moose harvest through selection of habitat improvement areas to attract moose into accessible sites,
- Produce biomass through mechanical treatments for use as a local fuel,
- Protect caribou habitat quality and berry production areas, and
- Maintain ecosystem integrity within the project area.

Project Area

Boundaries and Ownership

The Ahtna Traditional Use Territory consists of 26,589,244 acres or 41,500 square miles which encompasses the Copper River, Upper Susitna River, Upper Matanuska River, Upper White River and the headwaters of several watershed flowing north from the Alaska Range into the Tanana River. The Traditional Use Territory contains the 18,639,897 acre Ahtna Regional Corporation boundary established under the Alaska Native Claims Settlement Act (ANCSA) of 1971. Figure 1 shows both boundaries and their location in Alaska relative to the major river basins and population centers.

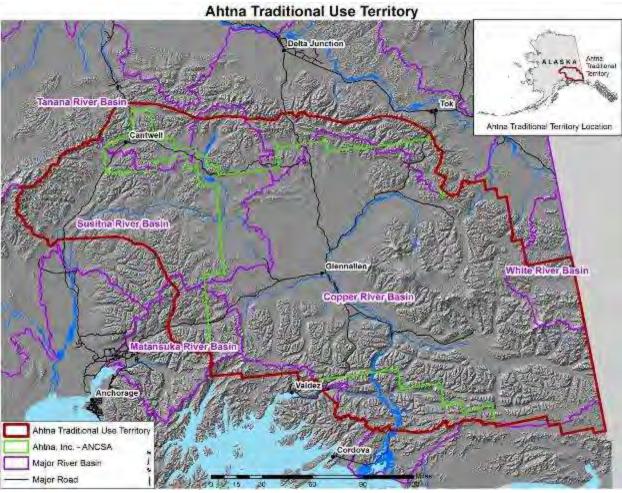


Figure 1. Ahtna Traditional Use Territory and Ahtna Regional Corporation boundary in Alaska.

Surface land ownership in the area is divided among several Federal agencies, the State of Alaska, Ahtna, Inc., Chitina Native Corporation, other native corporations, municipal government, and individual, private landowners. Figure 2 shows surface ownership in the Ahtna Traditional Use Territory. Table 1 displays surface ownership by acreage for each landowner.

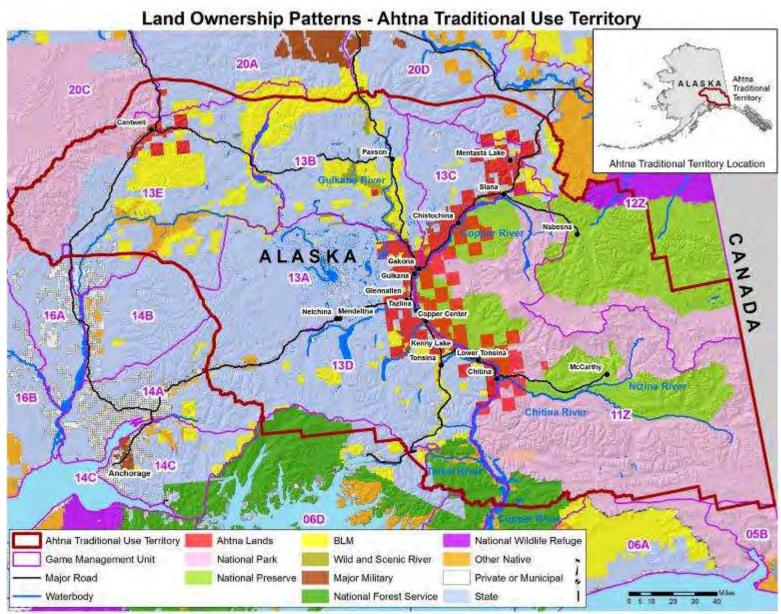


Figure 2. Surface ownership in the Ahtna Traditional Use Territory.

Table 1. Surface ownership by acreage for each landowner in the Ahtna Traditional Use Territory.

Surface Owner	Acres
Ahtna, Inc.	1,470,422
Ahtna, Inc Selected Lands	224,278
Chitina Native Corporation	105,782
Other Native Corporation	402,917
Native Allotments	32,988
Bureau of Land Management	2,924,640
United States Forest Service	817,447
National Park Service	12,504,014
United States Fish and Wildlife Service	143,663
Municipal Land	262
State of Alaska	11,012,926
Private	375,887

The National Park Service is the largest surface land owner in the project area, closely followed by the state of Alaska. Ahtna, Inc. and Chitina own a combined 1,800,482 acres when including lands that have been selected for transfer from Federal ownership to Ahtna, Inc. ownership.

Land management and planning varies based on the missions, needs, and goals of each land owner. Federal land managers within the project area include BLM, USFS, USFWS, and NPS. In addition, the NRCS provides guidelines for land management on both private and public land.

The mission of the BLM is to sustain the health, diversity, and productivity of public lands for the use and enjoyment of present and future generations. This is accomplished through a multiple-use management style that strives to strike the balance between healthy, sustainable ecosystems, the protection of natural, cultural, and historical resource values, and a wide range of public values and uses. The balancing of these factors is the lens used to evaluate proposed activities on BLM land.

The mission of the NPS (specifically Wrangell St. Elias and Denali National Parks and Preserves) is to ensure that these lands are properly administered for the enjoyment and education of the people, to protect their natural environment, and to assist state and local governments and citizen groups in the development of park areas. The specific mission of Wrangell St. Elias is, "preserve and protect ecological integrity and heritage resources of a vast ecosystem in south-central Alaska, while providing for public use in a wilderness setting. Wrangell St. Elias, at 13.2 million acres, was specifically designated to encompass an area large enough to include a diverse range of scenery, high latitude biomes, and landscape level processes where man is considered an integral part of the ecosystem. Ecosystem integrity and carefully planned public use is essential so there is opportunity for the continuation of subsistence lifestyles, future scientific investigations, interpretation of natural forces, and the inspiration and solitude of wilderness experience for present and future generations. Compatible public uses and

increased access, where appropriate, will be promoted to the extent that the quality of the experience and the natural and cultural resources are maintained." The specific mission of Denali is, "protect intact, the globally significant Denali ecosystems, including their cultural, aesthetic, and wilderness values, and ensure opportunities for inspiration, education, research, recreation and subsistence for this and future generations."

The mission of the USFWS (specifically Tetlin Wildlife Refuge) "is to conserve fish and wildlife populations and habitats in their natural diversity, to provide interpretation and environmental education to the public and to provide subsistence hunting opportunities to rural inhabitants." The Tetlin Refuge is located in the northeast corner of the planning area and represents a small (<1%) portion of the overall planning region.

The mission of the USFS (Chugach National Forest), "is to sustain the health, diversity, and productivity of the Nation's forests to meet the needs of present and future generations. The USDA Forest Service provides leadership in the protection, management, and use of the Nation's forest, rangeland, and aquatic ecosystems. Our ecosystem approach to management integrates ecological, economic, and social factors to maintain and enhance the quality of the environment to meet current and future needs. Through implementation of land and resource management plans, the agency ensures sustainable ecosystems by restoring and maintaining species diversity and ecological productivity that helps provide recreation, water, timber, minerals, fish, wildlife, wilderness, and aesthetic values for current and future generations of people."

While not a land owner, the NRCS provides an important advisory role in land management. The mission of the NRCS is to help people help the land. The NRCS states that they, "improve the health of our Nation's natural resources while sustaining and enhancing the productivity of American agriculture. We achieve this by providing voluntary assistance through strong partnerships with private landowners, managers, and communities to conserve, protect, restore, and enhance the lands and waters upon which people and the environment depend." The NRCS is an important partner for public and private entities in implementing conservation practices and other on the ground management.

Land owned by the state of Alaska is primarily managed by the Division of Forestry. The mission of the Division of Forestry is to develop, conserve, and enhance Alaska's forests to provide a sustainable supply of forest resources for Alaskans. This is done by, "protecting water quality, fish and wildlife habitat, and other forest values through appropriate forest practices and administration of the Forest Resources and Practices Act; managing a wildland fire program on public, private, and municipal lands; encouraging development of the timber industry and forest products markets; conducting timber sales for personal and commercial use and for fuel-wood; administering the Community Forestry, Conservation Education, Forest Health, and Stewardship programs; and giving technical assistance to forest landowners.

Geology

The geology of the Ahtna Traditional Use Territory was described in part in the <u>Copper River Basin Soil Survey</u>. Rocks in the area consist of schist, greenstone, graywacke, shale, and sandstone and andesite bedrock of Pleistocene age occurs in the southeastern part of the area. During Pleistocene glaciations (35,000 to 9,000 BP) glaciers covered the entire basin floor. During much of the glaciation period, ice dammed the channel of the Copper River through the Chugach Mountains forming a large proglacial lake in the central basin. "Lacustrine sediments deposited in the lake partially buried older glacial features. Over time, the lake level fluctuated widely, and eventually drained completely approximately 9,000 years ago (Ferrians, Nichols, and Williams 1983)." Following retreat of the glaciers and drainage of

the lake, permafrost formed in many lacustrine and glacial deposits. In addition, rivers incised canyons in the lacustrine and glacial sediments, and loess began to accumulate. Tarr and Martin (Tarr and Martin 1913) provided a detailed description of the geology of the Copper River Basin, noting that nearly all of the basin was derived from glacial deposits that are 500-700 ft. deep and in some places more than 1000 ft. deep. They reported some locations containing clay deposits likely deposited from glacial lakes, while other areas supported sand dunes deposited from windblown sands. Some areas supporting loess or eolian silt contain imbedded vegetation indicating that deglaciation occurred at least 700-1000 years ago.

Climate

The climate of the area was described in the Copper River Basin Soil Survey. This report stated: "The climate of the Copper River basin is subarctic continental characterized by long cold winters and short warm summers. Mean January temperature is -10 °F (-23 °C); daily low temperatures of -50 °F (-46 °C) or less occur frequently during the winter and may last for two or more weeks. Mean July temperature is 56 °F (13 °C); daily high temperatures on occasion exceed 85 °F (30 °C). Although the daily minimum temperature in summer averages in the forties, freezing temperatures have been recorded in every month.... the length of the growing season varies greatly from year to year. Mean annual precipitation across the basin ranges from 8 to 17 inches (23 to 41 cm). Of this, about 38 percent is received as rain during the growing season, which lasts from early June through the end of August. Thunderstorm activity is common during the early summer. During many years, a lack of precipitation in May and June results in a soil moisture deficit during the period of plant emergence. Average annual snowfall is 47 inches (119 cm) at Old Edgerton Farms in the Kenny Lake area and 49 inches (124 cm) at Glennallen. Although snowfall varies greatly from year to year, at least 1 inch (2.5 cm) of snow is on the ground an average of 180 days per year. Continuous sunlight and twilight occur from early June through mid-July. Day length at the winter solstice is less than 5 hours long. Prevailing wind at Gulkana airfield is from the southeast at 6.8 miles per hour (10.9 km per hour)."

Soils

Limited mapping has been done for soils in the Copper Data. Figure 3 displays both existing soil mapping and projected soil mapping to be completed by NRCS in coming years. Additional information about soil texture and soil drainage can be found in the individual village planning sections.

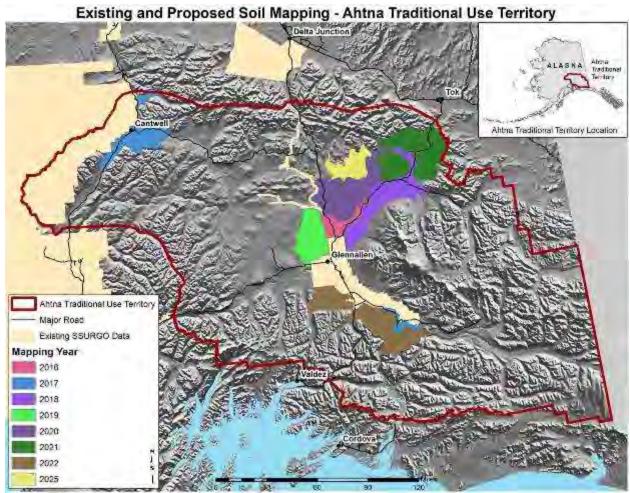


Figure 3. Existing and proposed soil mapping in the Ahtna Traditional Use Territory. Data from NRCS, Palmer Office.

Permafrost

A significant factor influencing the vegetation in the landscape is the occurrence of permafrost under some of the project area. The Copper River Basin Soil Survey described the role of permafrost as: "Permafrost, or perennially frozen ground, underlies most of the Copper River basin. The depth at which it occurs and its ice content varies widely. Permafrost characteristically occurs as ice crystals disseminated throughout the soil. Although not extensive near the soil surface, massive ice wedges and lenses do occur in the subsoil in some areas. A perched water table and saturated conditions are common above the permafrost during the summer due to restricted drainage. The fire history of the site and the thickness of the insulating organic layer on the soil surface control depth to permafrost and water table, in part. Disturbance of the organic layer usually results in increased soil temperatures and a lowering of the permafrost level. As permafrost thaws, a large volume of water is released. Variation in the ice content of the permafrost and the rate of thawing results in differential subsidence of the soil surface and slumping on steeper slopes. The occurrence of permafrost requires special consideration when selecting lands for clearing and agriculture and during construction of roads and buildings." Permafrost considerations should be evaluated in other management decisions including selection of areas for moose habitat improvements.

Vegetation Description

The Copper River Basin Soil Survey provided a general description of the vegetation occurring in the Tazlina project area. It stated: "The vegetation of the survey area is boreal forest, similar to elsewhere in Interior Alaska. Boreal forest consists of a mosaic of vegetation types reflecting the combined effects of landform, topographic position, soil type, and the occurrence of past fires. The Copper River basin has a long history of frequent wildfires. Between 1900 and 1950, an average of 10,000 acres burned annually, although this average has been reduced with improved fire protection measures (Barney 1969) (Figure 4). High-intensity crown fires that typically kill entire stands characterize the natural fire regime (Viereck and Schandlemeier 1980). Stands are then replaced through natural regeneration. Forest types on productive well-drained sites include white spruce, mixed white spruce-aspen, mixed white spruce-balsam poplar, aspen, and, in the southern end of the survey area, mixed white spruce/paper birch. Stunted black spruce and white spruce forests of low productivity occur on north facing slopes and cold, wet sites with shallow permafrost. Following forest fires, willow shrub dominates most sites until eventually replaced by forest vegetation. Where topographic and soil conditions inhibit tree growth, shrub and herbaceous vegetation develop. Seasonally flooded riverwash on the floodplains of major rivers supports dense alder shrub. Willow and ericaceous shrub occupy bogs, fens, and narrow drainages. Wet sedge meadows are common on the margins of lakes and ponds. Steppe vegetation, characteristic of semi-arid areas elsewhere in northeastern Asia and northwestern North America (Murry et al. 1983), is found on steep south-facing terrace escarpments."

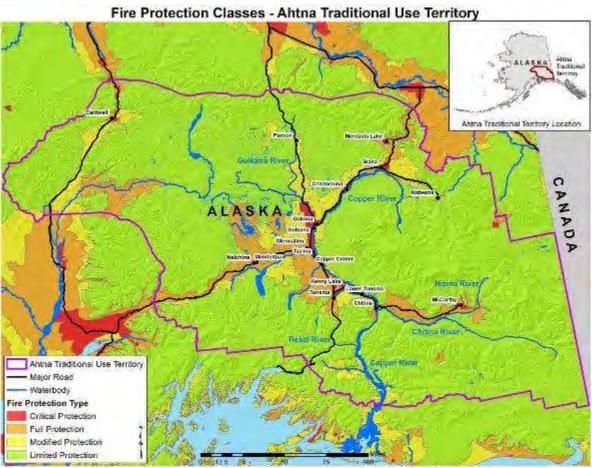


Figure 4. Current fire protection classes in Ahtna Traditional Use Territory. Data from Alaska Interagency Coordination Center.

Disturbance Factors

As mentioned in reference to permafrost and vegetation, fire is a disturbance factor influencing the vegetation ecology in the project area. Although the level of fire occurring in this southcentral Alaska landscape is substantially less than that occurring in more interior areas of Alaska north of the Alaska Range, fire is still a significant disturbance when it occurs. Lynch et al. (2004) estimated mean fire return intervals for the Copper River Basin as between 150-210 years, substantially longer than for many other areas of Alaska. Fire serves to set back succession. It can also burn off the organic material at the ground surface, including peat that can occur on many sites. This can influence the thermal layer protecting the underlying permafrost on some sites, causing the permafrost to melt (thermokarst) and changing the site conditions through this process. Riparian areas are also influenced by flooding and ice events. These serve to set back succession of vegetation in riparian areas, and can even shift site conditions, particularly in the case of significant flooding events. Insects and disease are another type of disturbance that influence ecosystems in the Copper Basin. Figure 5 displays areas that have been disturbed by fire or insects.

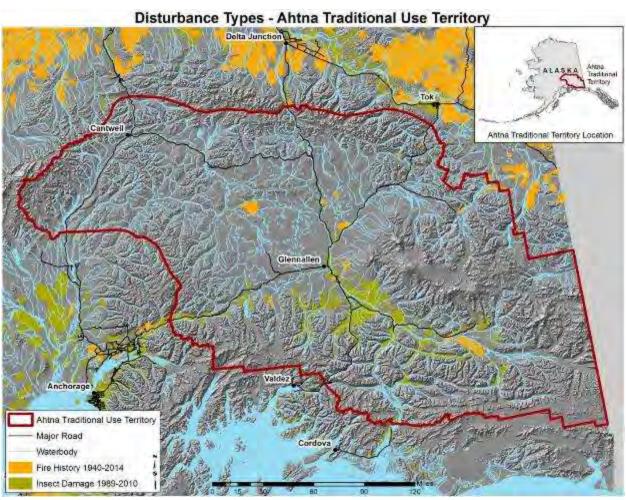


Figure 5. Disturbance factors (wildland fire and insect damage) in the Ahtna Traditional Use Territory.

Methods

Landscape Assessment Methods

As a component of the CIG project, the project team developed an ecosystem-based landscape classification system and mapped this classification in a GIS for use by CRITR and Ahtna. An ecosystem is considered a specific plant community defined by its composition, structure, and abiotic setting, and is thus a very specific description of a repeating vegetation community and its associated abiotic environment. Classifying ecosystems includes identifying a delineation of the abiotic environment that sets boundaries on the types of plant communities that can occur in each specifically identified abiotic setting as well as the specific plant communities that will occur in each abiotic setting in response to disturbance processes. The landscape in the Ahtna Traditional Use Territory was classified and mapped based on different ecological sites (abiotic differences in environmental conditions) and the plant communities occurring on each of these ecological sites in response to natural disturbances. The resulting ecosystem diversity framework provides the classification system that can then be used to map species habitat, biomass production areas, and other ecosystem services.

Ecological Sites

Several different types of classification systems were considered that could be used to identify and map the abiotic environment. We selected the biophysical setting classification used in <u>LANDFIRE</u> as the classification system to use as it could be applied across the entire project area which included 26.5 million acres of the Ahtna Traditional Use in Southcentral Alaska. <u>LANDFIRE</u> described each biophysical setting (BpS) within delineated ecoregions and then developed coarse maps of the locations of these BpS's. These maps had a number of inaccuracies in BpS designations. We made corrections where we could identify obvious errors, and produced an improved map of BpS locations within the Ahtna Traditional Use Territory, where possible. The BpS classification was stratified by both LANDFIRE zone and NRCS MLRA as shown in Figure 6.

Ecoregion Delineation

We divided the Ahtna Traditional Use Territory into discrete ecoregions using NRCS's Major Land Resource Areas (MLRA's) (Figure 6). Ecosystem classifications were described separately for each of these MLRA's. Thus, while we used the BpS classifications from <u>LANDFIRE</u>, we characterized the overall ecosystem diversity within each MLRA and used data specific to that MLRA in describing characteristics of each BpS and other classification categories.

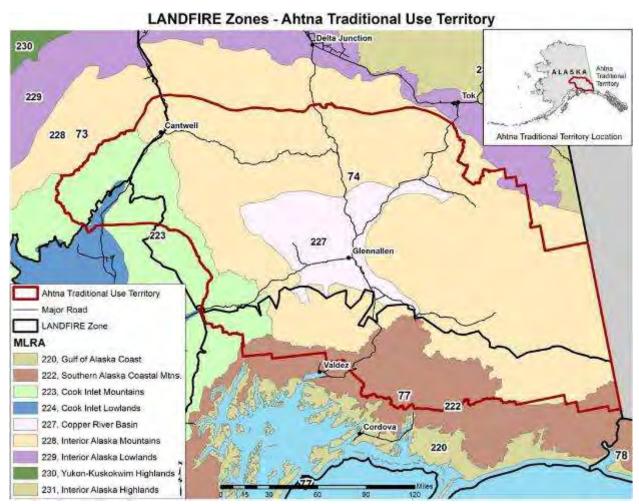


Figure 6. LANDFIRE zones and NRCS Major Land Resource Areas in the Ahtna Traditional Use Territory.

Disturbance Class

Once the classification of ecological sites (BpS) was selected and mapped within each MLRA, we then focused on classifying the disturbance processes and resulting plant communities (ecosystems) that could occur across each BpS. We relied on state and transition models to describe disturbance processes and the transitions among each specific ecosystem that resulted from either disturbances or succession.

We used different frameworks for state and transition models for forested ecosystems and grass and shrub ecosystems. The full ecosystem diversity framework for the Ahtna Traditional Use Territory and surrounding landscape including the disturbance class identified for each BpS type are shown in the ecosystem diversity matrices included in the following report sections.

Once we identified the classification of disturbance class, we mapped these to the extent possible with existing remotely-sensed information. While LANDFIRE has mapped disturbance classes, its accuracy is limited. Instead, we used existing vegetation mapping developed by the Wrangell-Saint Elias National Park for the eastern portion of the project area and existing vegetation mapping developed by Michael Fleming and hosted by the Geographic Information Network of Alaska for the western portion (Figure 7), and created a decision tree in Microsoft Access to crosswalk these classifications of vegetation to

disturbance classes. An example of the crosswalk decision tree for an upland forested biophysical setting is shown in Figure 8.

Moose and Caribou Models

Mapping ecosystem diversity seamlessly across a landscapes allows the development of ancillary models such as moose and caribou habitat suitability indices. These models use a habitat suitability index to depict the quality of moose and caribou habitat within the Ahtna Traditional Use Territory. A full description of the moose model including methods and results for the entire project area can be found in Appendix A. A full description of the caribou model including methods and results for the entire project area can be found in Appendix B. In addition, maps of moose and caribou results particular to each village planning region can be found in the relevant portion of each village plan.

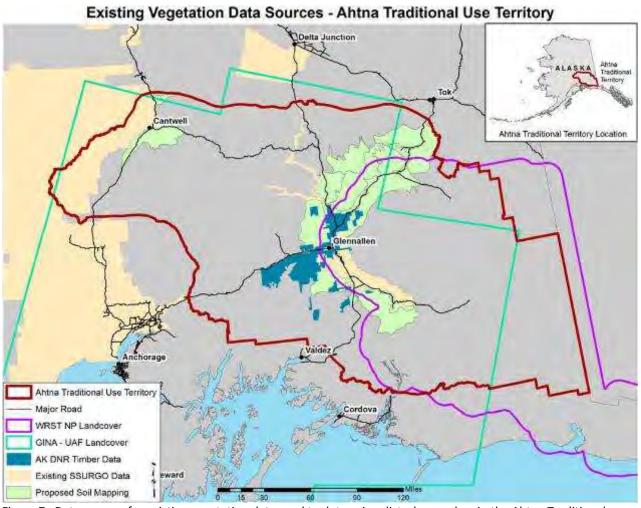


Figure 7. Data sources for existing vegetation data used to determine disturbance class in the Ahtna Traditional Use Territory.

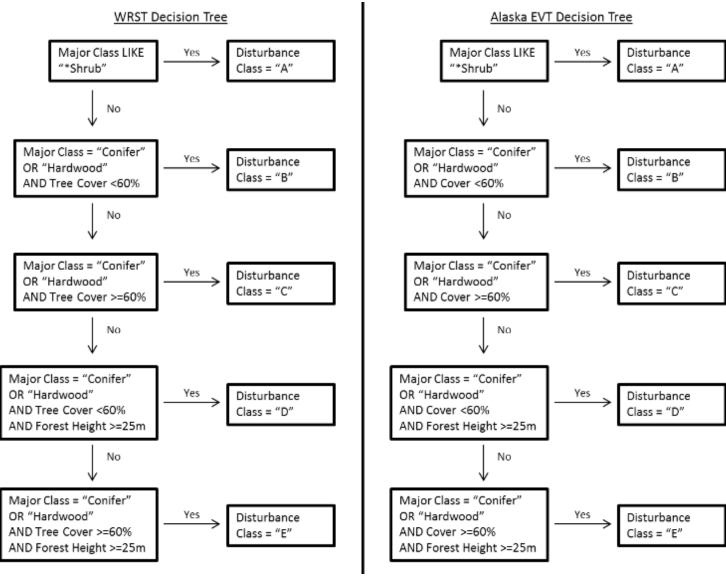


Figure 8. Example decision tree for BpS 16790 (White Spruce-Hardwood – SubBoreal). The primary difference between WRST and Alaska EVT is that tree cover was estimated from plot data in WRST and only total cover was available from LANDFIRE for Alaska EVT.

Ecosystem Diversity

Upland Forested Ecosystem Diversity

Upland forested ecosystems include BpS's that have greater than 10 percent tree cover under climax conditions and have vegetation that is not influenced by the presence of surface or subsurface water. Within the Ahtna Traditional Use Territory there are 5.65 million acres that are classified as upland forested BpS's. Approximately half of these acres, 2.84 million, are in successional states that contain mature trees and the other half, 2.82 million acres, consists of succession states dominated by shrubs, seedlings, and saplings.

Successional states within a given system are defined by the disturbance processes, the size and cover of the vegetation, and the plant species present. In upland forest BpS's the primary disturbances are wildfire, insects, and disease. Disturbance events can occur simultaneously as insect or disease outbreaks can lead to increased frequency and intensity of wildfire. In general, disturbance returns an upland forested system to grass/shrub successional states. Ecosystem diversity matrices that display successional states for upland forested systems in the Ahtna Traditional Use Territory are displayed for MLRA 222 in figure 9, MLRA 223 in figure 10, MLRA 227 in figure 11, and MLRA 228 in figure 12.

Treeline White Spruce Woodland - Boreal (16011)

This BpS covers an estimated 199,970 acres of the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 15,813 acres. This BpS is primarily found north of the Alaska Range, but occurs in pockets throughout the Copper River Basin. This type occurs at the elevational limit of tree growth and can occur as a fairly thin band in the transition zone between forested and grass/shrub types. Fire is the primary disturbance to this BpS, with a mean fire return interval estimated at 100 years (LANDFIRE).

For approximately the first 24 years following disturbance this type falls within an herbaceous and shrub vegetation class. Shrubs typically will resprout following fire and white spruce (*Picea glauca*) begins to reestablish from seeds that come from adjacent stands or remaining trees. The shrub layer is dominated by *Betula nana*, with *Vaccinium uliginosum*, *Ledum groenlandicum* and *Salix pulchra* being common. In some stands *Alnus viridis* may be the dominant shrub. The dominant ground cover is usually feathermoss or *Cladina* spp (<u>LANDFIRE</u>). This class was estimated to have historically occurred on 10% of this BpS, however this amount may be lower in the Copper River Basin with the lower amounts of fire occurring in this landscape compared to other areas in Alaska supporting this BpS such as north of the Alaska Range.

For the period of 25-69 years post disturbance there are two possible successional paths for this type. The first path (occurring in 4% of stands) is dominated by a hardwood or white spruce-hardwood forest. In this class, *Betula papyrifera* and *Populus tremuloides* gain canopy dominance over the shrubs. In some cases canopy dominance is shared with white spruce. Forest canopy cover is generally between 10-25%. Eventually hardwoods begin to die out and white spruce gains canopy dominance (<u>LANDFIRE</u>). The hardwood class historically occurred on 15% of this BpS.

MAJOR LAND RESOURCE AREA 222 - UPLAND FOREST SYSTEMS ECOSYSTEM DIVERSITY FRAMEWORK

						LAN	IDFIRE - BIOPHYSICAL SETTIN	vGS .				
	Mesic Black Spruce Canopy Cover ^a		Canopy Cover Open Closed		White Spruce-Hardwood Cahopy Cover Open Closed		Treeline White Spruce	Sitka Spruce	Mountain Hemlock Canopy Cover Open Closed		Periglacial Woodland- Shrubland Canopy Cover	
TREE							Canopy Cover Open Closed	Canopy Gover Open Closed				
	Open	DS1 (10%) SCLASS A	DS1 (5%) SCLASS A		DS 1 (5%) SCLASS A		DS 1 (9%) SCLASS A	DS1 (100%) SCLASS A	Open DS1 (5% SCLASS	a 1:	Open DS1 (15% SCLASS A	Closed) S B
GFS/SEEDLING- APLING (DBH<5")	BENA, LEDUM, VAUL, VAVI		TSHE, PISI, VAO	, MEFE	CACA4, EQAR, CHA	N9, MEFE	BENA, VAUL. LEGR. SAPU15		VAOV, MEFE,	RUSP	DRDR, EQVA, CHLA13, SASI2, SAB	
	DSZ - NO HARD SCLASS E			DS2 (15%) SCLASS B	DSZ (15%) SGLASS B	DS3 (10%) SCLASS C	DS2 (5%) SCLASS B			DS2 (??%)	DS2 (25% SCLASS (1
our i	PIMA, PIGL, BENA					DIOL BERG						
POLE (DBH 5-9")	DS3 - HARD (5%) SCLASS C			TSHE, PISI		PIGL, BEPA, POBA2, POTR5	PIGL, BEPA, BENA	PISI, OPHO, VAOV, SARA2		TSME, PILU, MEFE, ALVIS	POBA2, PISI, ALV	IS, SALIX
	BEPA, POTR5, PIMA, PIGL											
	DS4 (50%) SCLASS D	DS5 (25%) SCLASS E		DS3 (20%) SCLASS C	DS4 (65%) SCLASS D	DS5 (5%) SCLASS E	DS3 (90%) SCLASS C		DS3 (35%) SCLASS B	DS4 (60%) SCLASS C		DS3 (60%) SCLASS D
MEDIUM (DBH 9-20")	PIMA, PIGL, BENA, LEDUM	PIMA, PIGL, BENA, LEDUM		TSHE, PISI, VAOV, MEFE			PIGL, BENA, VAUL. CLADIS		TSME, PILU, MEFE, ALVIS	TSME, PILU, MEFE, ALVIS		PISI, RUSP SARA2, OPHO
				DS4 (60%) SCLASS D	PIGL, BEPA, POBA2, POTR5	PIGL, BEPA, POBA2, POTR5						
LARGE (DBH >20")				TSHE, PISI, VAOV, MEFE								
ACRES	110,86		74,69 5		55,120 16790		39,105 16012	16,683 16440	4,35 5		2,502	

Figure 9. Ecosystem diversity matrix for MLRA 222 – Upland forest ecosystems. The percentages in parenthesis are the historical estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. ^aOpen canopy cover has a value from 10-59% canopy cover and closed canopy cover has a value from 60-100%.

MAJOR LAND RESOURCE AREA 223 - UPLAND FOREST SYSTEMS ECOSYSTEM DIVERSITY FRAMEWORK

						LANDFIRE - BIOF	HYSICAL SETTINGS						
	White Spruce-Ha Subbore		White Spruce-Hardwood- Boreal	Mesic Black S Subbore		Treeline White Spruce - Subboreal	Treeline White Spr Boreal	ruce -	Mesic Birch-A	Aspen	Dry Aspen-Steppe Bluff	Mountain Hemio	ock North
TREE SIZECLASS	Canopy Cov		Canopy Cover			Canopy Cover	Carropy Cover		Canopy Cover		Canopy Cover	Canopy Cover	
and control to be	Open DS 1 (5%)		Open Closted D81 (20%)	Ореп	DS1 (18%)	Open Closed DS 1 (5%)	Open DS 1 (10%)	Gissed	Open DS1 (10%)		<u>Cpsn Glosed</u> 081 (20%)	Open DS1 (5%	
GFS/SEEDLING-	SCLASS A. CACA4, EQAR, CHANS, MEFE		SCLASS A & B CHANA2, CACA4, EQSY, EQAR, ARUV, VAVI, BENA, LEPAD		SCLASS A BENA LEDUM	SCLASS A BENA, VAUL, LEGR, SAPU18	SCLASS A BENA, VAUL, LEGR, SAPU	118, ALVIS	SCLASS A & B CHANS, CACA4, EQUIS, MEPA, ROAC, VIED, LEDUM, ALNUS		SCLASS A & B CAPU, PSSP6, ARFR4, JUCO6, POTR5.	SGLASS VAOV, MEFE,	2.
SAPLING (DBH<5")	100000000000000000000000000000000000000		VAVI. BENA, LEPAD		VAUL VAVI		1 1 1 1 1 1 1 1		VIED, LEDUM, A	LNUS	ROAC, ARALS	2000	
	DSZ (16%) SCLASS B	DS3 (10%) SCLASS C	DS2 - HARD (3D%) SCLASS C	DS2 - NO HARI SCLASS		DS2 (5%) SGLASS B	DS 2 (15%) SCLASS B			DS3 (15%) SCLASS C	DS2 (70%) SCLASS C		bs2 (??%)
DOLE.	POW POWN COM	PIGL, BEPA.	BEPA, POTRS, PIGL, ROAC	PIMA, PIGL, BENA	377374					BEPA.			
POLE (DBH 5-9")	PIGL. BEPA, POBA2. POTR5	POBAZ POTRS	DS3 - NO HARD (10%) SCLASS D	DS3 - HARD SCLASS	C	PIGL, BEPA, BENA	BEPA, POTRB, PIGL, PIMA			POTRS, ROAC, VIED	POTRS, ROAG, JUCOS, ARALS		TSME, PILU, MEFE, ALVIS
			PIGL, BENA, ARRU, VAVI	BEPA, POTR5, PI	IMA, PIGL								
	DS4 (65%) SCLASS D	DS5 (5%) SCLASS E	DS 4 (40%) SCLASS E	DS4 (50%) SCLASS D	DS5 (25%) SCLASS E	DS3 (90%) SCLASS C	DS 3 (75%) SCLASS C		DS4 (60%) SCLASS E	D55 (15%) SCLASS D	DS3 (10%) SCLASS D	DS3 (35%) SCLASS B	DS4 (60%) SCLASS C
MEDIUM (DBH 9-20")			PIGL, ROAC, VIED, BENA	PIMA, PIGL, BENA, LEDUM	PIMA, PIGL, BENA, LEDUM	PIGL. BENA, VAUL. CLADIS	PIGL; PIMA, BENA, CLADIS		BEPA, POTRS, ALNUS, LEDUM	BEPA, POTR5, ROAC, VIED	POTRS, PIGL, ROAC, JUCOS	TSME, PILU, MEFE, ALVIS	TSME, PILU. MEFE, ALVIS
	PIGL BEPA, POBA2. POTR5	PIGL, BEPA. POBA2, POTR6											
LARGE (DBH >20")													
ACRES	85,79 4		258,974 16030	41,10		37,224 16012	7137		27,320 16050	(-	2656 16061	2815 16481	

Figure 10. Ecosystem diversity matrix for MLRA 223 – Upland forest ecosystems. The percentages in parenthesis are the historical estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. Open canopy cover has a value from 10-59% canopy cover and closed canopy cover has a value from 60-100%.

MAJOR LAND RESOURCE AREA 227 - UPLAND FOREST SYSTEMS ECOSYSTEM DIVERSITY FRAMEWORK

						LANDFI	RE - BIOPH	IYSICAL SETTINGS						
	White Spruce-Hardwood - Boreal	White Spruce-Har SubBorea		Mesic Black Spruce - B	Boreal	Mesic Black Sp SubBorea		Treeline White Sp Boreal	pruce -	Treeline White Sp SubBoreal	ruce -	Mesic Birch-	Aspen	Dry Aspen-Steppe Bluff
TREE	Canopy Cover ^a	Canopy Cov	er	Canopy Cover		Canopy Cov	er er	Canopy Cover		Canopy Cover		Canopy Cover		Canopy Cover
SIZECLASS	Open Closed	Open	Closed	Open Cl	losed	Open	Closed	Open Closed		Open Closed		Open Closed		Open Closed
	DS1(20%) SCLASS A & B	DS 1 (5%) SCLASS A		DS1 (20%) SCLASS A & B			DS1 (10%) SCLASS A	DS 1 (10%) SCLASS A		DS 1 (5%) SCLASS A		DS1 (10%) SCLASS A 8		DS1 (20%) SCLASS A & B
GFS/SEEDLING- SAPLING (DBH<5")	CHANA2, CACA4, EQSY, EQAR, ARUV, VAVI, BENA, LEPAD	CACA4, EQAR, CHA	N9, MEFE	CACA4, CHAN9, EQUIS, S BENA, LEDUM, ROA	4, CHAN9, EQUIS, SALIX,		BENA, LEDUM, VAUL, VAVI		BENA, VAUL, LEGR, SAPU15, ALVI5		SA P U15	CHAN9, CACA4, EQUIS, MEPA, ROAC, VIED, LEDUM, ALNUS		CAPU, PSSP6, ARFR4, JUCO6, POTR5, ROAC, ARAL5
	DS2 - HARD (30%) SCLASS C	DS2 (15%) SCLASS B	DS3 (10%) SCLASS C	DS2 - HARD (30%) SCLASS C		DS2 - NO HARD SCLASS E		DS 2 (15%) SCLASS B		DS2 (5%) SCLASS B			DS3 (15%) SCLASS C	DS2 (70%) SCLASS C
POLE (DBH 5-9")	BEPA, POTR5, PIGL, ROAC DS3 - NO HARD (10%) SCLASS D PIGL, BENA, ARRU, VAVI	PIGL, BEPA, POBA2, POTR5	PIGL, BEPA, POBA2, POTR5	BEPA, POTR5, PIMA, F DS3 - NO HARD (30% SCLASS D PIMA, PIGL, BENA, LED	6)	DS3 - HARD SCLASS (BEPA, POTR5, PII	(5%)	BEPA, POTR5, PIGL, PIMA		PIGL, BEPA, BENA		BEP/ POTR ROA! VIEC		POTR5, ROAC, JUCO6, ARAL5
	DS 4 (40%) SCLASS E	DS4 (65%) SCLASS D	DS5 (5%) SCLASS E	DS4 (20%) SCLASS E		DS4 (50%) SCLASS D	DS5 (25%) SCLASS E	DS 3 (75%) SCLASS C		DS3 (90%) SCLASS C		DS4 (60%) SCLASS E	DS5 (15%) SCLASS D	DS3 (10%) SCLASS D
MEDIUM (DBH 9-20")	PIGL, ROAC, VIED, BENA		PIGL,	PIM A, PIGL, BENA, LEDUM		PIMA, PIGL, BENA, LEDUM	PIM A, PIGL, BENA, LEDUM	PIGL, PIMA, BENA, CLADI3		PIGL, BENA, VAUL, CLADI3		BEPA, POTR5, ALNUS, LEDUM	BEPA, POTR5, ROAC, VIED	POTR5, PIGL, ROAC, JUCO6
		PIGL, BEPA, POBA2, POTR5	BEPA, POBA2,											
LARGE (DBH >20")			POTR5											
ACRES	941,726	6217	l.	358,657		5012	•	106,193	ı	9947		67,835	•	10,378
	16030	16790		16041		16042		16011		16012		16050		16061

Figure 11. Ecosystem diversity matrix for MLRA 227 – Upland forest ecosystems. The percentages in parenthesis are the historical estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. Open canopy cover has a value from 10-59% canopy cover and closed canopy cover has a value from 60-100%.

MAJOR LAND RESOURCE AREA 228 - UPLAND FOREST ECOSYSTEM DIVERSITY FRAMEWORK

				LAN	DFIRE - BIOPHYSICAL SETTI	NGS			
	White Spruce-Hardwood - Boreal	White Spruce-Hardwood - SubBoreal	Mesic Black Spruce - Boreal	Mesic Black Spruce - SubBoreal	Treeline White Spruce - Boreal	Treeline White Spruce - SubBoreal	Mesic Birch-Aspen	Subalpine Balsam Poplar- Aspen	Dry Aspen-Steppe Bluff
TREE	Canopy Cover ^a	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover
SIZECLASS	Open Closed	Open Closed	Open Closed	Open Closed	Open Closed	Open Closed	Open Closed	Open Moderate Closed	Open Moderate Closed
	DS1 (20%) SCLASS A & B	DS 1 (5%) SCLASS A	DS1 (20%) SCLASS A & B	DS1 (10%) SCLASS A	DS 1 (10%) SCLASS A	DS 1 (5%) SCLASS A	DS1 (10%) SCLASS A & B	DS1(25%) SCLASS A	DS1(20%) SCLASS A & B
GFS/SEEDLING- SAPLING (DBH<5")	CHANA2, CACA4, EQSY, EQAR, ARUV, VAVI, BENA, LEPAD	CACA4, EQAR, CHAN9, MEFE	CACA4, CHAN9, EQUIS, SALIX, BENA, LEDUM, ROAC	BENA, LEDUM, VAUL, VAVI	BENA, VAUL, LEGR, SAPU15, ALVI5	BENA, VAUL, LEGR, SAPU15	CHAN9, CACA4, EQUIS, MEPA, ROAC, VIED, LEDUM, ALNUS	POBA2, POTR5, VIED, ROAC	CAPU, PSSP6, ARFR4, JUCO6, POTR5, ROAC, ARAL5
	DS2 - HARD (30%) SCLASS C	DS2 (15%) DS3 (10%) SCLASS B SCLASS C	DS2 - HARD (30%) SCLASS C	DS2 - NO HARD (10%) SCLASS B	DS 2 (15%) SCLASS B	DS2 (5%) SCLASS B	DS3 (15%) SCLASS C	DS2 (75%) SCLASS B	DS2 (70%) SCLASS C
POLE (DBH 5-9")	BEPA, POTR5, PIGL, ROAC DS3 - NO HARD (10%) SCLASS D PIGL, BENA, ARRU, VAVI	PIGL, PIGL, BEPA, POBA2, POTR5 POBA2, POTR5	BEPA, POTR5, PIMA, PIGL DS3 - NO HARD (30%) SCLASS D PIMA, PIGL, BENA, LEDUM	PIM A, PIGL, BENA, LEDUM DS3 - HARD (5%) SCLASS C BEPA, POTR5, PIM A, PIGL	BEPA, POTR5, PIGL, PIMA	PIGL, BEPA, BENA	BEPA, POTR5, ROAC, VIED	POBA2, POTR5, VIED, ROAC	POTR5, ROAC, JUCO6, ARAL5
	DS 4 (40%) SCLASS E	DS4 (65%) DS5 (5%) SCLASS D SCLASS E	DS4 (20%) SCLASS E	DS4 (50%) DS5 (25%) SCLASS D SCLASS E	DS 3 (75%) SCLASS C	DS3 (90%) SCLASS C	DS4 (60%) DS5 (15%) SCLASS E SCLASS D		DS3 (10%) SCLASS D
MEDIUM (DBH 9-20")	PIGL, ROAC, VIED, BENA		PIMA, PIGL, BENA, LEDUM	PIM A, PIGL, BENA, PIGL, BENA, LEDUM	PIGL, PIMA, BENA, CLADI3	PIGL, BENA, VAUL, CLADI3	BEPA, POTR5, POTR5, ALNUS, LEDUM ROAC, VIED		POTR5, PIGL, ROAC, JUCO6
LARGE (DBH >20")		PIGL, BEPA, BEPA, POBA2, POTR5							
ACRES	2,167,144	169,029	221,238	262,276	87,387	303,473	246,672	39,798	6,686
BpS CODE	16030	16790	16041	16042	16011	16012	16050	16070	16061

Figure 12. Ecosystem diversity matrix for MLRA 228 – Upland forest ecosystems. The percentages in parenthesis are the historical estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. ^aOpen canopy cover has a value from 10-59% canopy cover and closed canopy cover has a value from 60-100%.

The second successional path from the herbaceous and shrub class is directly to a white spruce dominated class. This is the most common successional path for this BpS and is the climax vegetation class for this type. This class is dominated by white spruce with canopy cover from 10-25%. The understory includes a variety of low shrubs, herbs, and mosses. As the stand ages, lichens (primarily *Cladina* spp.) become more prevalent (<u>LANDFIRE</u>). This state was estimated to occur across 75% of this BpS under historical fire regimes.

Vegetation plots for the Treeline White Spruce Woodland-Boreal BpS were sampled by Ahtna in the planning landscape. The results of this sampling are presented in Table 2. Both the hardwood and white spruce dominated classes have a total carbon availability of 91.2 tons/acre and an annual production of 0.064 tons/acre. There are approximately 44 tons of biomass available per acre. Photos depicting different vegetation stands in the Treeline White Spruce Woodland-Boreal BpS are found in Figures 13-15.

Table 2. Vegetation characteristics from plots sampled in the Treeline White Spruce Woodland-Boreal BpS from the planning landscape. GFS stands for grass, forb and seedling size class, Seed/Sap stands for seedling-sapling size class, Pole stands for pole size class (5-9" DBH trees), medium refers to the medium size class (9-20" DBH), Large refers to large trees (>20" DBH). Dwarf Shrub is shrub cover less than 1 m tall, medium shrub cover is shrubs 1-3 m in height and tall shrubs are >3m in height. SS refers to seedlings and saplings. BA refers to basal area of trees. TPA refers to trees per acre. Values are mean values with the standard deviation in parentheses.

CODE	SIZE-CLASS	DWARF SHRUB	MEDIUM SHRUB	TALL SHRUB	LICHEN	SEED/ SAPLING	POLE	MEDIUM	LARGE	SEED/ SAPLING	POLE	MEDIUM	LARGE	
			% Cove	(StDev)			TPA (StDev)		Basal Area (StDev)				
16011-A	GFS	11.45 (13.44)	30.5 (11.57)	21.14 (11.84)	0.14 (0.65)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	
16011-B	POLE-HWD	0 (0)	5.0 (0)	50.5 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	



Figure 13. Example of BpS 16011-A stand in Ahtna Traditional Use Territory (LANDFIRE Photo).



Figure 14. Example of BpS 16011-C stand in Ahtna Traditional Use Territory (LANDFIRE Photo).



Figure 15. Example of BpS 16030-B stand in Ahtna Traditional Use Territory.

Treeline White Spruce Woodland – Sub-boreal (16012)

This BpS covers an estimated 386,651 acres of the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 28,297 acres. This BpS is primarily found south of the Alaska Range, and occurs throughout the Copper River Basin. This type occurs at the elevational limit of tree growth and can occur as a fairly thin band in the transition zone between forested and grass/shrub types. Fire is the primary disturbance to this BpS, with a mean fire return interval estimated at 300 years (LANDFIRE). Spruce bark beetles are also a disturbance factor, especially where fires are more restricted. Spruce bark beetle outbreaks have been reported and mapped in the Copper River Basin.

For approximately the first 24 years following disturbance this type falls within an herbaceous and shrub vegetation class. Shrubs typically will resprout following fire and white spruce (*Picea glauca*) begins to reestablish from seeds that come from adjacent stands or remaining trees. The shrub layer is dominated by *Betula nana*, with *Vaccinium uliginosum*, *Ledum groenlandicum* and *Salix pulchra* being common. In some stands *Alnus viridis* may be the dominant shrub. The dominant ground cover is usually feathermoss or *Cladina* spp (<u>LANDFIRE</u>). This class was estimated to have historically occurred on 5% of this BpS.

For the period of 25-69 years post disturbance there are two possible successional paths for this type. The first path (occurring in 4% of stands) is dominated by a hardwood or white spruce-hardwood forest. In this class, *Betula papyrifera* and *Populus tremuloides* gain canopy dominance over the shrubs. In some cases canopy dominance is shared with white spruce (<u>LANDFIRE</u>). Forest canopy cover is generally between 10-25%. Eventually hardwoods begin to die out and white spruce gains canopy dominance. The hardwood historically occurred on 5% of this BpS.

The second successional path from the herbaceous and shrub class is directly to a white spruce dominated class. This is the most common successional path for this BpS and is the climax vegetation class for this type. This class is dominated by white spruce with canopy cover from 10-25%. The understory includes a variety of low shrubs, herbs, and mosses. As the stand ages, lichens (primarily *Cladina* spp.) become more prevalent (<u>LANDFIRE</u>). This state was estimated to occur across 90% of this BpS under historical fire regimes.

Both the hardwood and white spruce dominated classes have a total carbon availability of 91.2 tons/acre and an annual production of 0.064 tons/acre. There are approximately 44 tons of biomass available per acre.

White Spruce Hardwood - Boreal (16030)

This BpS covers an estimated 3,346,867 acres of the Ahtna Traditional Use Territory, and is the dominant biophysical setting in the project area. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 482,283 acres. Fire is the primary disturbance to this BpS, with a mean fire return interval estimated at 150 years, but with considerable variation. Spruce beetle may also be a disturbance factor, especially where fires are more restricted. Spruce beetle outbreaks have been reported and mapped in the Copper River Basin.

Following fire, an herbaceous disturbance class will occur for approximately 5 years depending on fire severity. Common species include: *Chamerion angustifolium, Calamagrostis canadensis, Equisetum sylvaticum, E. arvense, Geocaulon lividum, Mertensia paniculata and Pyrola* ssp. (Viereck et al. 1992). This disturbance class was estimated to have historically occurred on 5% of this BpS, however this amount may be lower in the Copper River Basin with the lower amounts of fire occurring in this landscape compared to other areas in Alaska supporting this BpS such as north of the Alaska Range.

A shrub and sapling disturbance class will typically occur from 5-29 years post-fire. Common shrubs include *Rosa acicularis, Viburnum edule, Betula nana, Ledum palustre ssp. Decumbens, L. groenlandicum, Vaccinium vitis-idaea, V. uliginosum, Empetrum nigrum,* and also trembling aspen saplings (Viereck et al. 1992, LANDFIRE). With low severity fire, plants may regenerate from underground propagules, while with high severity fire establishment from seeding will occur. Sites with high amounts of regenerating aspen will be high quality foraging areas for moose. This disturbance class is estimated to have historically occurred on 15% of this BpS, but again may be lower in the Copper River Basin due to lower amounts of fire. Peters et al. (2005) reported that white spruce regeneration in mixed hardwood sites was influenced by whether a site burned in a year with high amounts of masting by white spruce. They found that when a site burned concurrent with a masting year, substantially more white spruce regenerated on the site than if a fire occurred 1-3 years prior to a masting year. However, Peters et al. (2005) studied initial versus delayed regeneration of white spruce and found little relationship between fire and masting as a major influence on whether a site had initial regeneration or delayed regeneration of white spruce. They suggested that fire severity played an important role in addition to other finer scale site factors.

An intermediate disturbance class occurring from 30-129 years post-fire is a mixed hardwood and spruce type. It has a mix of white spruce, black spruce, and aspen, with spruce increasing in dominance as it overtakes aspen which will be getting more decadent. With senescence of the hardwoods, spruce will dominate, with 25-70% canopy cover (LANDFIRE). This state is estimated to have occurred on up to 30% of the BpS. When this state is dominated by hardwoods the total carbon availability is 83.6 tons/acre with an annual carbon production of 0.76 tons/acre. There are approximately 41 tons of biomass available per acre. When this state is dominated by spruce the total carbon availability is 91.2 tons/acre with an annual carbon production of 0.064 tons/acre. There are approximately 44 tons of biomass available per acre.

A mature forest state generally occurs >130 years post-fire. This state is characterized by stands of spruce, primarily white spruce but can be mixed with black spruce. The understory includes *Rosa acicularis, Viburnum edule, Shepherdia canadensis, Vaccinium vitis-idaea, Arctostaphylos spp., Linnaea borealis, Chamerion angustifolium and Geocaulon lividum* (LANDFIRE). On some sites, increasing cover of lichens will occur, specifically various *Cladina* species, which can provide good foraging habitat for caribou. Feathermoss may occur on some sites, particularly following low severity fire, and may keep lichen abundance at lower amounts, but it is less characteristic on this setting than in wetter biophysical settings. This state was estimated to have occurred across 40% of this BpS under historical fire regimes. This state has a total carbon availability of 141.6 tons/acre with an annual carbon production of 0.34 tons/acre. There are approximately 68.3 tons of biomass available per acre.

Vegetation plots for the White Spruce Hardwood-Boreal BpS were sampled by Ahtna in the planning landscape. The results of this sampling are presented in Table 3. Photos depicting different vegetation stands in the White Spruce Hardwood-Boreal BpS are found in Figures 16 and 17.

Table 3. Vegetation characteristics from plots sampled in the White Spruce Hardwood-Boreal BpS from the planning landscape. GFS stands for grass, forb and seedling size class, Seed/Sap stands for seedling-sapling size class, Pole stands for pole size class (5-9" DBH trees), medium refers to the medium size class (9-20" DBH), Large refers to large trees (>20" DBH). Dwarf Shrub is shrub cover less than 1 m tall, medium shrub cover is shrubs 1-3 m in height and tall shrubs are >3m in height. SS refers to seedlings and saplings. BA refers to basal area of trees. TPA refers to trees per acre. Values are mean values with the standard deviation in parentheses.

CODE	SIZE-CLASS	DWARF SHRUB	MEDIUM SHRUB	TALL SHRUB	LICHEN	SEED/ SAPLING	POLE	MEDIUM	LARGE	SEED/ SAPLING	POLE	MEDIUM	LARGE
			% Cove	r (StDev)			TPA (StDe	Basal Area (StDev)					
16030-A	GFS	6.6 (6.9)	24.2 (27.0)	11.5 (14.4)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)
16030-B	SEED/SAP	9.1 (9.8)	30.9 (20.7)	19.0 (18.4)	0 (0)	1601.3 (1019.5)	0 (0)	0 (0)	0 (0)	31 (9.8)	0 (0)	0 (0)	0 (0)
16030-C	POLE-HWD	8.2 (12.4)	10.8 (12.5)	9.2 (7.4)	0 (0)	620 (0)	280 (0)	0 (0)	0 (0)	95.1 (0)	74.0 (0)	0 (0)	0 (0)
16030-D	POLE-CON	5.3 (9.8)	26.2 (26.3)	11.2 (13.2)	0 (0)	636.9 (562.8)	175 (111.5)	6 (0)	0 (0)	57.8 (32.2)	41.6 (24.9)	4.0 (1.0)	0 (0)
16030-E	MEDIUM	17.4 (23.3)	39.4 (29.0)	13.8 (16.2)	0.4 (1.0)	462.1 (365.3)	219.1 (120.2)	77.2 (51.6)	17.5 (10.6)	102.6 (47.1)	93.7 (45.5)	57.0 (42.1)	93.1 (91.3)



Figure 16. Example of BpS 16030-C stand in Ahtna Traditional Use Territory.



Figure 17. Example of BpS 16030-E stand in Ahtna Traditional Use Territory.

Mesic Black Spruce- Boreal (16041)

This BpS occurs across an estimated 579,483 acres of the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 125,045 acres. It is characterized by black spruce as the dominant overstory species, but with white spruce also occurring on many sites (Nature Serve 2008). The shrub component is typically *Rosa acicularis, Betula nana, Ledum* spp., *V. uliginosum, Vaccinium vitis-idaea* and *Empetrum nigrum* while the predominant herbaceous species include *Calamagrostis canadensis, Chamerion angustifolium* and *Equisetum* spp. (Nature Serve 2008). Fire is the primary disturbance factor typically resetting the successional process. This BpS may be difficult to distinguish from the White Spruce Hardwood Boreal Forest type when white spruce occurs mixed with black spruce on mesic black spruce sites.

Early disturbance classes and successional processes are very similar to those of the White Spruce Hardwood Boreal Forest BpS. Where aspen occurs in younger stands, good foraging areas for moose may be provided. These disturbance classes were estimated to have historically occurred on 20% of this BpS. Mid successional stages (30-119 years) are dominated by either black spruce which may be mixed with some white spruce with feathermoss occurring in the understory, or occurring as a mixed hardwood and black spruce forest (LANDFIRE). Tree cover typically ranges from 50-70%. These two states were estimated to have each historically comprised 30% of this BpS. When this state is dominated by hardwoods the total carbon availability is 83.6 tons/acre with an annual carbon production of 0.76 tons/acre. There are approximately 41 tons of biomass available per acre. When this state is dominated by spruce the total carbon availability is 91.2 tons/acre with an annual carbon production of 0.76 tons/acre. There are approximately 44 tons of biomass available per acre.

Late successional conditions (>120 years old) contain open, old black spruce with tree cover generally less than 60%, with some sites mixed with white spruce. Understories vary from tall or short shrubs, herbaceous species, or mosses and lichens (LANDFIRE). On some sites, where feathermoss has not predominated in the understories, lichens can increase over time. These sites may become high quality sites for caribou forage. Late successional conditions were estimated to historically occur on 20% of this BpS. This state has a total carbon availability of 131.1 tons/acre with an annual carbon production of 0.46 tons/acre. There are approximately 63 tons of biomass available per acre.

Vegetation characteristics of the Mesic Black Spruce-Boreal BpS for the planning landscape are listed in Table 4. Photos depicting different vegetation stands in the Mesic Black Spruce-Boreal BpS are found in Figures 18-21.

Table 4. Vegetation characteristics from plots sampled in the Mesic Black Spruce-Boreal BpS for the planning landscape. GFS stands for grass, forb and seedling size class, Seed/Sap stands for seedling-sapling size class, Pole stands for pole size class (5-9" DBH trees), medium refers to the medium size class (9-20" DBH). Dwarf Shrub is shrub cover less than 1 m tall, medium shrub cover is shrubs 1-3 m in height and tall shrubs are >3m in height. SS refers to seedlings and saplings. BA refers to basal area of trees. TPA refers to trees per acre. Values are mean values with the standard deviation in parentheses.

C	ODE	SIZE-CLASS	DWARF SHRUB	MEDIUM SHRUB	TALL SHRUB	LICHEN	SEED/ SAPLING	POLE	MEDIUM	SEED/ SAPLING	POLE	MEDIUM	
				% Cover	(StDev)			TPA (StDev)		Basal Area (StDev)			
160	041-B	SEED/SAP	12.9 (15.0)	23.4 (19.9)	11.5 (10.5)	6.5 (16.0)	971.8 (478.9)	10 (0)	0 (0)	25.1 (2.5)	2.06 (0)	0 (0)	
160	041-D	POLE	14 (18.5)	57.1 (31.9)	14.6 (14.7)	2.2 (5.7)	929 (719.7)	135.4 (78.3)	0 (0)	56.1 (34.9)	29 (18.9)	0 (0)	
160	041-E	MEDIUM	22.1 (26.7)	58.8 (31.8)	12.2 (15.2)	1.6 (3.9)	632.6 (351.9)	212.3 (110.6)	47.2 (31.7)	84.5 (39.7)	68.51 (40.1)	29.1 (20.9)	



Figure 18. Example of BpS 16041-B stand in Ahtna Traditional Use Territory (LANDFIRE Photo).



Figure 19. Example of BpS 16041-C stand in Ahtna Traditional Use Territory.



Figure 20. Example of BpS 16041-D stand in Ahtna Traditional Use Territory (LANDFIRE Photo).



Figure 21. Example of BpS 16041-E stand in Ahtna Traditional Use Territory.

Mesic Black Spruce- Sub-boreal (16042)

This BpS occurs across an estimated 410,832 acres of the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 54,545 acres. This type is typically found south of the Alaska Range. Both black spruce and white spruce are share canopy dominance in mature stands (Nature Serve 2008). The shrub component is typically *Betula nana*, *Ledum spp.*, *V. uliginosum*, *Vaccinium vitis-idaea* and *Empetrum nigrum* (Nature Serve 2008). Fire is the primary disturbance factor typically resetting the successional process. The fire return interval is estimated to be around 170 years with a longer interval than boreal sites due to the reduced prevalence of lightning strikes in this area (LANDFIRE). This BpS may be difficult to distinguish from the White Spruce Hardwood Boreal Forest type when white spruce occurs mixed with black spruce on mesic black spruce sites.

Early disturbance classes and successional processes are very similar to those of the White Spruce Hardwood Boreal Forest BpS. Where aspen occurs in younger stands, good foraging areas for moose may be provided. Early successional stages occur from 0-14 years following disturbance and occurred on 10% of the historical landscape.

Mid successional stages (15-75 years) are dominated by either black spruce which may be mixed with some white spruce, or occurring as a mixed hardwood and black spruce forest. Tree cover typically ranges around 60% cover (LANDFIRE). These two states were estimated to have each historically comprised 30% of this BpS. When this state is dominated by hardwoods the total carbon availability is 83.6 tons/acre with an annual carbon production of 0.76 tons/acre. There are approximately 41 tons of biomass available per acre. When this state is dominated by spruce the total carbon availability is 91.2 tons/acre with an annual carbon production of 0.064 tons/acre. There are approximately 44 tons of biomass available per acre.

Late successional conditions (>75 years old) also result in two different stand types. The first consists of open, spruce with tree cover generally less than 60%, with some sites mixed with white spruce. Understories vary from tall or short shrubs, herbaceous species, or mosses and lichens (LANDFIRE). These sites may become high quality sites for caribou forage with the spread of *Cladina* lichen species. This type comprised 50% of historical landscapes in this BpS. This state has a total carbon availability of 83.6 tons/acre with an annual carbon production of 0.76 tons/acre. There are approximately 41 tons of biomass available per acre.

The second stand type is a closed mature spruce forest. Canopy cover ranges between 60% and 70%. The understory includes various tall shrubs, low shrubs, herbs, mosses, and lichens. This type comprised 25% of historical landscapes in this BpS. This state has a total carbon availability of 131.1 tons/acre with an annual carbon production of 0.46 tons/acre. There are approximately 63 tons of biomass available per acre.

Mesic Birch Aspen Forest- Boreal (16050)

This BPS occurs on an estimated 338,288 acres of rolling hills and side slopes of the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 84,159 acres. Soils are well drained glacial till, loess, and colluvium (Nature Serve 2008). These sites tend to be warmer and drier than white spruce dominated BpS sites, and are dominated by aspen in the Copper River Basin with balsam poplar an associated species. Canopy cover ranges from 25-90% (LANDFIRE). Understory species include *Alnus* spp., *Ledum* spp., *Vaccinium vitisidaea*, *Betula nana*, *Rosa acicularis*, *Viburnum edule* and *Equisetum* spp. with feathermoss common as well. Fire is a primary disturbance

factor, but tends to occur at longer fire return intervals than spruce stands. This BpS can serve as a fire break under certain conditions. Leaf miner may be an additional disturbance factor to aspen stands. Seral stages may be difficult to distinguish from those of the White-Spruce Hardwood- Boreal BpS (LANDFIRE).

Following fire, herbaceous species including *Chamerion angustifolium, Calamagrostis canadensis, Equisetum sylvaticum, E. arvense, Mertensia paniculata* and *Geocaulon lividum* can occur along with aspen that is propagating from suckers (Viereck and Schandelmeier 1980). This state, lasting approximately 5 years, is estimated to occur on 5% of the BpS, but again may be less in the Copper River Basin due to the reduced amounts of fire in this landscape. Shrub cover then tends to dominate from 5-14 years post-fire and historically occurred on 5% of this BpS. Shrubs can include *Alnus* spp., *Ledum* spp., *Vaccinium vitis-idaea, Betula nana, Rosa acicularis, Shepherdia canadensis* and *Viburnum edule* along with aspen (LANDFIRE). This state may provide good forage habitat for moose.

From 15-49 years, hardwoods, principally aspen but sometimes with balsam poplar, will become the overstory with the shrub species maintaining in the understory. This state was estimated to have historically occupied 15% of the BpS. Hardwoods mature from 50-99 years, with this state estimated to have occurred on 15% of the BpS. This state will still maintain a shrub and feathermoss understory. Stands >100 years post-fire historically occurred on 60% of the BpS and contain old and dying aspen, with resprouting of aspen around dead trees (LANDFIRE). For all states within this BpS the estimated available carbon is 129.2 tons per acre with annual carbon production of 1.82 tons per acre. Available biomass is approximately 64 tons per acre.

Vegetation characteristics of the Mesic Birch Aspen Forest-Boreal BpS for the planning landscape are listed in Table 5. Photos depicting different vegetation stands in the Mesic Birch Aspen Forest-Boreal BpS are found in Figures 22-24.

Table 5. Vegetation characteristics from plots sampled in the Mesic Birch Aspen Forest-Boreal BpS from the planning landscape. Seed/Sap stands for seedling-sapling size class, Pole stands for pole size class (5-9" DBH trees), medium open refers to the medium size class (9-20" DBH) with <60% canopy cover, medium closed refers to the medium size class with >60% canopy cover. Dwarf Shrub is shrub cover less than 1 m tall, medium shrub cover is shrubs 1-3 m in height and tall shrubs are >3m in height. SS refers to seedlings and saplings. BA refers to basal area of trees. TPA refers to trees per acre. Values are mean values with the standard deviation in parentheses.

CODE	SIZE CLASS	DWARF SHRUB	MEDIUM SHRUB	TALL SHRUB	LICHEN	SEED/ SAPLING	POLE	MEDIUM	LARGE	SEED/ SAPLING	POLE	MEDIUM	LARGE
		% Cover (StDev)				TPA (StDev)				Basal Area (StDev)			
16050-B	SEED/SAP	2.6 (4.1)	9.1 (14.0)	15.8 (18.1)	0 (0)	250 (0)	0 (0)	0 (0)	0 (0)	12.27 (0)	0 (0)	0 (0)	0 (0)
16050-C	POLE	1.6 (2.9)	38.6 (29.7)	7 (5.7)	0 (0)	712 (314.4)	152 (49)	7 (0)	0 (0)	54.1 (13.4)	29.9 (8.9)	4.28 (0)	0 (0)
16050-D	MEDIUM-OPEN	0 (0)	60 (0)	0 (0)	0 (0)	2615 (0)	115 (0)	65 (0)	0 (0)	151.8 (0)	100 (0)	80 (0)	0 (0)
16050-E	MEDIUM-CLOSED	19.3 (30.1)	55.5 (31.9)	18 (30.0)	0 (0)	274.3 (289.2)	131 (84.5)	59.4 (35.3)	6 (4.2)	61.8 (32.5)	60.4 (31.8)	42.5 (24.5)	15.1 (6.9)



Figure 22. Example of BpS 16050-B stand in Ahtna Traditional Use Territory.



Figure 23. Example of BpS 16050-C stand in Ahtna Traditional Use Territory (LANDFIRE Photo).



Figure 24. Example of BpS 16050-D stand in Ahtna Traditional Use Territory.

White Spruce Hardwood – Sub-boreal (16790)

This BpS covers an estimated 392,000 acres of the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 26,289 acres. Fire is rare in this BpS, with a mean fire return interval estimated at 600 years, but with considerable variation. Spruce bark beetles are a major disturbance factor as well. Spruce bark beetle outbreaks occur every 50 years on average and result in a thinning of the overstory spruce canopy.

Following fire, an herbaceous and shrub disturbance class will dominate from years 0 to approximately year 29. Common herbaceous species include *Calamagrostis canadensis*, *Equisetum arvense*, *Dryopteris expansa* and *Gymnocarpium dryopteris*. Common shrub species include *Menziesia ferruginea*, *Alnus viridis* ssp. *sinuata*, *Vaccinium ovalifolium*, *Oplopanax horridus*, *Vaccinium vitis-idaea* and *Linnaea borealis* (Viereck et al. 1992, <u>LANDFIRE</u>). Some white spruce, aspen, and birch seedlings may be present depending on fire severity. Sites with high amounts of regenerating aspen will be high quality foraging areas for moose. This disturbance class was estimated to have historically occurred on 5% of the BpS.

An intermediate disturbance class occurring from 30-129 years post-fire is a mixed hardwood and spruce type. There are two pathways for this time frame with one resulting in a more closed canopy (60-100% cover) and the other a more open canopy (25-60% cover). Both classes contain a mix of white spruce, aspen, birch, and poplar with spruce increasing in dominance as it overtakes the hardwoods which become more decadent over time. Common understory species include *Rosa acicularis*, *Equisetum* spp. and *Linnaea borealis* (LANDFIRE). The closed class is estimated to have occurred on 10% of the BpS historically, and the open class occurred on 15% of the BpS. Both stand types contain an

estimated 83.58 tons of carbon per acre with an annual carbon production of 0.76 tons per acre. Biomass averages approximately 41 tons per acre.

A mature forest state generally occurs >130 years post-fire. This disturbance state also includes two pathways consisting of open and closed stands. Both pathways are characterized by stands of white spruce, with occasional remnant hardwoods. The understories are composed of a variety of tall shrubs, low shrubs, herbs, mosses, and lichens (LANDFIRE). The open state has tree canopy cover <60% and occurred on 65% of the BpS historically. The closed state has canopy cover >60% and occurred across 5% of this BpS under historical fire regimes. Both stand types contain an estimated 109.46 tons of carbon per acre with an annual carbon production of 1.28 tons per acre. Biomass averages approximately 54 tons per acre.

Reference Conditions and Current Conditions

The landscape modeling done by the LANDFIRE team makes it possible to calculate the estimated acres in each BpS successional state under historical reference conditions. These are the conditions that were expected to have occurred prior to human intervention of fire regimes. While the level of control of fires in South Central Alaska may be limited, in locations near to human development effects of fire control may cause significant shifts in amounts of disturbance states that would be present without this control. The percentages listed for each disturbance class that were estimated by the models can be compared between estimated historical reference conditions and existing conditions. Table 6 shows percentages of the landscape and acres of upland forested types for the entire project area and Table 7 shows percentages and acres for upland forested types on Ahtna lands within the project area. Significant differences in these Tables are apparent. It is unlikely that these differences are present, but rather, that inaccuracies in both the estimated historical amounts and in the amounts of current disturbance states are the cause. Thus, both sets of numbers should be viewed with caution, but the underlying principles should be recognized and considered in future planning.

Table 6. Acres and percentages for each disturbance state in upland forested types for historical reference conditions and existing conditions in the Ahtna Traditional Use Territory.

				E	stimate	d Historical Re	ference Condition	ons - Upland For	est Types	<u> </u>	<u> </u>			
	Distu	rbance	State	(%)				Disturband	ce State (acres)					
BpS	Α	В	С	D	E	BpS	Α	В	С	D	E			
16011	10	15	75	-	-	16011	19,996.98	29,995.47	149,977.35	N/A	N/A			
16012	5	5	90	-	-	16012	19,332.54	19,332.54	347,985.75	N/A	N/A			
16030	5	15	30	10	40	16030	167,343.34	502,030.01	1,004,060.03	334,686.68	1,338,746.70			
16041	5	15	30	30	20	16041	28,974.17	86,922.50	173,844.99	173,844.99	115,896.66			
16042	10	10	5	50	25	16042	41,083.18	41,083.18	20,541.59	205,415.92	102,707.96			
16050	5	5	15	15	60	16050	16,914.38	16,914.38	50,743.15	50,743.15	202,972.62			
16790	5	10	15	65	5	16790	19,600.02	39,200.03	58,800.05	254,800.22	19,600.02			
						Existing Co	nditions - Upland	Forest Types						
	Distu	rbance	State	(%)			Disturbance State (acres)							
BpS	Α	В	С	D	E	BpS	Α	В	С	D	E			
16011	17.8	82.1	0.1	-	-	16011	35,598.32	164,235.81	135.66	N/A	N/A			
16012	28.7	71.2	0.1	-	-	16012	110,953.76	275,261.84	435.23	N/A	N/A			
16030	2.5	30.4	67.0	0.0	0.2	16030	83,434.82	1,016,782.38	2,241,061.00	0	5,588.56			
16041	2.0	12.5	8.8	76.6	0.1	16041	11,645.49	72,473.86	50,920.23	444,124.81	318.91			
16042	25.4	56.8	17.6	0.2	0.0	16042	104,316.82	233,369.31	72,136.48	970.98	38.25			
16050	51.6	47.9	0.0	0.2	0.3	16050	174,610.98	161,932.47	0	730.57	1,013.68			
16790	45.7	49.2	4.8	0.3	0.0	16790	179,005.08	192,759.98	18,846.20	1,273.66	115.42			

Table 7. Acres and percentages for each disturbance state in upland forested types for historical reference conditions and existing conditions on lands owned by Ahtna, Inc.

					Estimate	ed Historical R	eference Condit	ions - Upland Fo	orest Types		
	Distu	ırbance	State	(%)				<u>Disturba</u>	nce State (acres)		
BpS	Α	В	C	D	E	BpS	Α	В	С	D	E
16011	10	15	75	-	-	16011	1,581.32	2,371.98	11,859.88	N/A	N/A
16012	5	5	90	-	-	16012	1,414.83	1,414.83	25,467.03	N/A	N/A
16030	5	15	30	10	40	16030	24,114.17	72,342.50	144,685.01	48,228.34	192,913.34
16041	5	15	30	30	20	16041	6,252.25	18,756.74	37,513.47	37,513.47	25,008.98
16042	10	10	5	50	25	16042	5,454.46	5,454.46	2,727.23	27,272.32	13,636.16
16050	5	5	15	15	60	16050	4,207.94	4,207.94	12,623.83	12,623.83	50,495.34
16790	5	10	15	65	5	16790	1,314.34	2,628.69	3,943.03	17,086.45	1,314.34
						Existing Co	nditions - Uplar	nd Forest Types			
	Distu	ırbance	State	(%)				Disturba	nce State (acres)		
BpS	Α	В	C	D	E	BpS	Α	В	С	D	E
16011	21.0	79.0	0.0	-	=	16011	3313.68	12486.37	2.45	N/A	N/A
16012	21.4	78.4	0.1	-	=	16012	6069.38	22182.84	17.57	N/A	N/A
16030	2.8	19.7	77.3	0.0	0.2	16030	13478.69	94989.15	372812.33	0.00	904.93
16041	3.2	10.8	6.1	80.1	0.0	16041	3958.41	13514.27	7675.07	100105.79	49.59
16042	20.5	58.1	20.7	0.2	0.0	16042	11185.13	31673.53	11274.34	106.75	3.56
16050	66.6	32.8	0.0	0.2	0.2	16050	56084.89	27577.41	0.00	173.25	207.49
16790	55.5	39.7	4.2	0.3	0.0	16790	14581.32	10447.00	1111.75	80.51	11.56

Upland Grass and Shrub Ecoystems

Upland grass and shrub BpS's have less than 10 percent tree cover under climax conditions and have vegetation that is not influenced by the presence of surface or subsurface water. These sites typically occur in areas that are not capable of supporting trees due to elevation, soil depth, climate, or frequent disturbance. Within the Ahtna Traditional Use Territory there are 5.3 million acres that are classified as upland grass and shrub BpS's.

Successional states within a given BpS are defined by the disturbance processes, the size and cover of the vegetation, and the plant species present. In upland grass and shrub BpS's the primary disturbance types are wildfire, avalanches, and wind. In grass and shrub BpS's certain disturbances such as avalanches or wind keep taller vegetation such as large shrub and trees from establishing. Disturbances such as wildfire return a system from a shrub to a grass state. Ecosystem diversity matrices that display successional states for upland grass and shrub systems in the Ahtna Traditional Use Territory are displayed for MLRA 222 in figure 25, MLRA 223 in figure 26, MLRA 227 in figure 27, and MLRA 228 in figure 28.

MAJOR LAND RESOURCE AREA 222 - UPLAND GRASS-SHRUB SYSTEMS ECOSYSTEM DIVERSITY FRAMEWORK

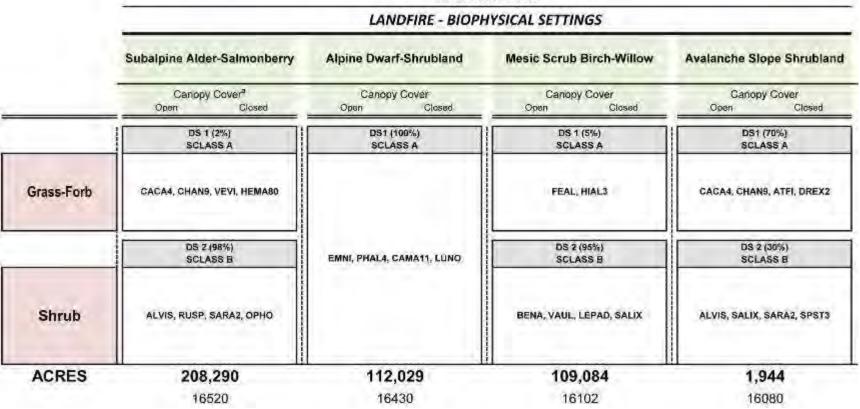


Figure 25. Ecosystem diversity matrix for MLRA 222 – Upland grass-shrub ecosystems. The percentages in parenthesis are the estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. ^aOpen canopy cover has a value from 10-59% and closed canopy cover has a value from 60-100%.

MAJOR LAND RESOURCE AREA 223 - UPLAND GRASS-SHRUB SYSTEMS ECOSYSTEM DIVERSITY FRAMEWORK

					LANDF	IRE - BIOPHYSICAL SI	TTINGS				
	Alpine Ericaceous Dwarf-Shrubland	Masic Scrub Birch- Willow	Mesic Subalpine Alder	Alpine Dwarf-Shrub Summit	Low Shrub-Tussock Tundra	Alpine Dwarf Shrubland	Mesic Bluejoint Meadow	Alpine Mesic Herbaceous Meadow	Dry Grassland	Tussock Tundra	Avalanche Slope Shrubland
	Canopy Cover ^a Open Closed	Canopy Cover Open Closed	Canopy Cover Open Siosed	Canopy Cover Open Closed	Canapy Cover Open Closed	Canopy Cover Open Cosed	Canopy Cover Open Skeed	Canopy Cover Open Gosed	Canopy Cover Open Closed	Canopy Cover Open Closed	Canopy Cover Open Cosed
	DS 1 (100%) SCLASS A	DS 1 (5%) SCLASS A	DS 1 (5%) SCLASS A	DS 1 (100%) SCLASS A	DS1 (10%) SCLASS A	DS1 (100%) SCLASS A	DS1 (100%) SCLASS A	DS1 (100%) SCLASS A	091 (100%) SCLASS A	DS1 (85%) SCLASS A	DS1 (30%) SCLASS A
Grass-Forb		FEAL, HIAL3	CACA4, CHANS, ATFI, DREX2		ERVA4, CABI5, CACA4, ARCTA					ERVA4, CABI6	CACA4, CHAN9, ATFL DREX2
	CATE11, EMNI, VAUL, DRINA	DS 2 (95%) SCLASS B	DS 2 (95%) SGLASS B	VAUL, EMNI, VAVI, DRYAS	DS 2 (15%) DS 3 (76%) SCLASS C SCLASS B	EMNI, PHAL4, GAMA11, LUNO	CACA4, CHANA2, HEMASO, ATFI	CAMA11, GEER2, SACA14, VASI	FEAL, FERUZ, CAPU, BRINA	DS 2 (15%) SCLASS B	DS 2 (70%) SGLASS B
Shrub		BENA, VAUL, LEPAD, SALIX	ALVIS, SALIX, SARA2, SPST3		BENA, SALIX, BENA, SALIX, ERVA4, ERVA4, PECA80 GABIS					CHCA2, VAOX, ERVA4, CABI5	ALVIS, SALIX, SARA2, SPST3
ACRES	312,775 16351	220,823 16102	126,769 16090	105,327 16310	83,884 16280	50,432 16430	20,077 16110	17,377 16450	15,009 16120	4,533 16290	3,415 16800

Figure 26. Ecosystem diversity matrix for MLRA 223 – Upland grass-shrub ecosystems. The percentages in parenthesis are the estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. ^aOpen canopy cover has a value from 10-59% and closed canopy cover has a value from 60-100%.

MAJOR LAND RESOURCE AREA 227 - UPLAND GRASS-SHRUB SYSTEMS ECOSYSTEM DIVERSITY FRAMEWORK

			LANDF	IRE - BIOPHYSICAL SET	TINGS			
	Mesic Scrub Birch- Willow	Low Shrub-Tussock Tundra	Alpine Ericaceous Dwarf-Shrubland	Tussock Tundra	Alpine Mesic Herbaceous Meadow	Mesic Bluejoint Meadow	Mesic Subalpine Alder	
- <u></u>	Canopy Cover ^a Open Closed	Canopy Cover Open Closed	Canopy Cover Open Closed	Canopy Cover Open Closed	Canopy Cover Open Closed	Canopy Cover Open Closed	Canopy Cover Open Closed	
	DS 1 (5%) SCLASS A	DS1 (10%) SCLASS A	DS 1 (100%) SCLASS A	DS1 (85%) SCLASS A	DS1 (100%) SCLASS A	DS1 (100%) SCLASS A	DS 1 (5%) SCLASS A	
Grass-Forb	FEAL, HIAL3	ERVA4, CABI5, CACA4, ARCTA		ERVA4, CABI5			CACA4, CHAN9, ATFI, DREX2	
	DS 2 (95%) SCLASS B	DS 2 (15%) DS 3 (75%) SCLASS C SCLASS B	CATE11, EM NI, VAUL, DR IN 4	DS 2 (15%) SCLASS B	CABI5, BENA, ARAL2, EM NI	CACA4, CHANA2, HEMA80, ATFI	DS 2 (95%) SCLASS B	
Shrub	BENA, VAUL, LEPAD, SALIX	BENA, BENA, SALIX, SALIX, ERVA4, ERVA4, PECA60 CABI5		CHCA2, VAOX, ERVA4, CABI5			ALVIS, SALIX, SARA2, SPST3	
ACRES	89,588	73,551	9,602	7,890	4,299	1,952	1,938	
	16102	16280	16351	16290	16330	16110	16090	

Figure 27. Ecosystem diversity matrix for MLRA 227 – Upland grass-shrub ecosystems. The percentages in parenthesis are the estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. ^aOpen canopy cover has a value from 10-59% and closed canopy cover has a value from 60-100%.

MAJOR LAND RESOURCE AREA 228 - UPLAND GRASS-SHRUB SYSTEMS ECOSYSTEM DIVERSITY FRAMEWORK

- -					L	ANDFIRE - BIOPHYS	SICAL SETTINGS					
	Mesic Scrub Birch/Willow	Alpine Ericaceous Dwarf-Shrubland	Alpine Dwarf Shrubland	Low Shrub-Tussock Tundra	Mesic Subalpine Alder	Alpine Dwarf- Shrub Summit	Avalanche Slope Shrubland	Tussock Tundra	Subalpine Alder- Salmonberry	Alpine Mesic Herbaceous Meadow	Mesic Bluejoint Meadow	Dry Grassland
	Canopy Cover ^a	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover
	Open Closed	Open Closed	Open Closed	Open Closed	Open Closed	Open Closed	Open Closed	Open Closed	Open Closed	Open Closed	Open Closed	Open Closed
[DS 1 (5%) SCLASS A	DS 1 (100%) SCLASS A	DS1 (100%) SCLASS A	DS1 (10%) SCLASS A	DS 1 (5%) SCLASS A	DS 1 (100%) SCLASS A	DS1 (70%) SCLASS A	DS1(85%) SCLASS A	DS 1 (2%) SCLASS A	DS1 (100%) SCLASS A	DS1 (100%) SCLASS A	DS1 (100%) SCLASS A
Grass- Forb	FEAL, HIAL3			ERVA4, CABI5, CACA4, ARCTA	CACA4, CHAN9, ATFI, DREX2		CACA4, CHAN9, ATFI, DREX2	ERVA4, CABI5	CACA4, CHAN9, VEVI, HEM A80			
	DS 2 (95%) SCLASS B	CATE11, EMNI, VAUL, DRIN4	EM NI, PHAL4, CAM A 11, LUNO	DS 2 (15%) DS 3 (75%) SCLASS B	DS 2 (95%) SCLASS B	VAUL, EMNI, VAVI, DRYAS	DS 2 (30%) SCLASS B	DS 2 (15%) SCLASS B	DS 2 (98%) SCLASS B	CABI5, BENA, ARAL2, EMNI	CACA4, CHANA2, HEMA80, ATFI	FEAL, FERU2, CAPU, BRINA
Shrub	BENA, VAUL, LEPAD, SALIX			BENA, BENA, SALIX, SALIX, ERVA4, ERVA4, PECA60 CABI5	ALVIS, SALIX, SARA2, SPST3		ALVIS, SALIX, SARA2, SPST3	CHCA2, VAOX, ERVA4, CABI5	ALVIS, RUSP, SARA2, OPHO			
ACRES	2,344,241	944,326	487,028	359,659	212,862	142,722	40,101	28,100	27,776	15,155	4,615	4,541
	16102	16351	16430	16280	16090	16310	16080	16290	16520	16330	16110	16120

Figure 28. Ecosystem diversity matrix for MLRA 228 – Upland grass-shrub ecosystems. The percentages in parenthesis are the estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. Open canopy cover has a value from 10-59% and closed canopy cover has a value from 60-100%.

Boreal Mesic Scrub Birch-Willow Shrubland (16102)

This BpS occurs across an estimated 2,712,003 acres of the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 128,705 acres. The system occurs on well-drained sites and often is found in the subalpine. The shrub layer is usually dominated by *Betula nana* with *Vaccinium uliginosum*, *Ledum decumbens*, *Salix pulchra*, or *Salix barclayi* may be common or dominant. Dwarf shrubs like *Empetrum nigrum* and *Vaccinium vitis-idea* are often found in the low shrub layer. Herbaceous species are often scarce but occasionally include *Festuca altaica* and *Hierochloe alpina*. These vegetation types occur on mesic sites on mid to upper slopes, above tree line and on terraces and sideslopes. The soils are mineral with a well-decomposed organic layer 5-30cm thick (Viereck et al. 1992, NatureServe 2008). Fire is the primary disturbance with rapid resprouting of shrubs following fire. In some cases, woodlands near timberline may be converted to this type following fire (Pegau 1972). Fire return intervals are likely >100 years.

Fire severity determines the successional pathway for these stands. High severity fire results in herbaceous dominated stands, typically consisting of *Festuca altaica* and *Hierochloe alpina*. This state persists for longer than 5 years with high severity fire and may dominate for up to 4 years with low to moderate severity fire (LANDFIRE). This state can occur on up to 5% of the BpS, but is likely less common in the Copper Basin due to reduced fire frequency.

The shrub state is by far the most common, occurring 95% of the time in this BpS. Due to relatively low fuels this state can persist >300 years before fire would return it to the herbaceous class. Sites are typically dominated by *Betula nana*. *Betula glandulosa*, *Vaccinium uliginosum*, *Ledum decumbens*, *Salix pulchra*, or *Salix barclayi* may be common or dominant (Viereck 1979, Viereck et al. 1992). Dwarf shrubs like *Empetrum nigrum* and *Vaccinium vitis-idea* are often found in the low shrub layer. Trees may begin to encroach the shrubland given enough time post-disturbance.

Vegetation characteristics of the Boreal Mesic Scrub Birch-Willow Shrubland BpS for plots sampled in the planning landscape are listed in Table 8. A photo depicting a late successional vegetation stand in the Boreal Mesic Scrub Birch-Willow Shrubland BpS is found in Figure 29.

Table 8. Vegetation characteristics from plots sampled in the Boreal Mesic Scrub Birch-Willow BpS for the planning landscape. GFS stands for grass, forb and seedling size class. Dwarf Shrub is shrub cover less than 1 m tall, medium shrub cover is shrubs 1-3 m in height and tall shrubs are >3m in height. Values are mean values with the standard deviation in parentheses.

CODE	SIZE-CLASS	DWARF SHRUB	MEDIUM SHRUB	TALL SHRUB	LICHEN
			% Cover	(StDev)	
16102-B	GFS	6.67 (12.01)	44.73 (21.02)	14.85 (12.92)	1.15 (6.44)



Figure 29. Example of BpS 16102-B stand in Ahtna Traditional Use Territory (LANDFIRE Photo).

Boreal Low Shrub Tussock Tundra (16280)

This BpS occurs across an estimated 508,039 acres of the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 37,200 acres. These sites usually have permafrost present and contain a surface peat layer 10-40 cm thick (Viereck et al. 1992). Fire is the primary disturbance with return intervals varying widely between sites (Viereck and Schandelmeier 1980). As a generalization the mean fire return interval is considered 200 years.

Following fire, an herbaceous disturbance class will dominate from years 0 to approximately year 14. *Eriophorum* spp. and *Carex* spp. regrow from rhizomes and shrubs begin to resprout from rootstock. In some stands *Calamagrostis* spp. and *Acrtagrostis* spp. dominate following fire (<u>LANDFIRE</u>). Under historic disturbance regimes this state occupied 10% of the BpS.

An intermediate disturbance class occurring from 15-80 years post-fire is a low shrub and tussock type. The tussocks are dominated by *Eriophorum* spp. and *Carex* spp. with common shrubs including *Betula nana*, *Salix* spp. and *Vaccinium uliginosum*. Lichens also begin to reestablish in this state, but have cover <25% (LANDFIRE). This state represented 75% of the BpS under historical conditions. A photo depicting this succession stage is found in Figure 30.

The climax class in this BpS is distinguished by the cover of lichen species. In most stands this occurs 80+ years post-fire. Lichen species have total cover >25%. Herbaceous and shrub species present in earlier states remain common (<u>LANDFIRE</u>). This state represented 15% of the BpS under historical conditions. A photo depicting this succession stage is found in Figure 31.



Figure 30. Example of BpS 16280-B stand in Ahtna Traditional Use Territory (LANDFIRE Photo).



Figure 31. Example of BpS 16280-C stand in Ahtna Traditional Use Territory (S. Yeats Photo).

Boreal Alpine Dwarf-Shrub Summit (16310)

This BpS occurs across an estimated 243,966 acres in the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 16,485 acres. This shrub system occurs on exposed, windswept summits and ridges in alpine and subalpine areas of Alaska (Viereck et al. 1992). The sites have sparse cover due to the exposed nature. Common shrubs include *Dryas integrifolia*, *Vaccinium uliginosum*, *Empetrum nigrum*, *Vaccinium vitis-idaea*, *Diapensia lapponica*, *Loiseuria procumbens*, *Salix arctica*, *Salix rotundifolia*, *Salix reticulata*, *Arctostaphylos rubra* and *Arcostaphylos* alpina. Exposed rock and lichens can be abundant. Herbaceous species include *Hierochloe alpina*, *Polygonum bistorta*, *Anemone* spp., *Festuca* spp. and *Luzula* spp (Viereck et al. 1992). There is little disturbance in this system with the shrub communities representing a stable climax and specific plant distribution being controlled by wind desiccation and a short growing season.

Due to the relatively stable nature of this BpS there is only one state. This is a shrub state with typical shrub cover <50% and interspersed with exposed rocks and lichen. The dominant species are listed in the previous paragraph.

Vegetation characteristics of the Boreal Alpine Dwarf-Shrub Summit BpS for plots sampled in the planning landscape are shown in Table 9. A photo depicting a vegetation stand in the Boreal Alpine Dwarf-Shrub Summit BpS is found in Figure 32.

Table 9. Vegetation characteristics from plots sampled in the Boreal Alpine Dwarf-Shrub Summit BpS for the planning landscape. GFS stands for grass, forb and seedling size class. Dwarf Shrub is shrub cover less than 1 m tall, medium shrub cover is shrubs 1-3 m in height and tall shrubs are >3m in height. Values are mean values with the standard deviation in parentheses.

CODE	SIZE-CLASS	DWARF SHRUB	TALL SHRUB LICHEN		
			% Cover	(StDev)	
16310-A	GFS	4.65 (5.34)	0.44 (0.95)	0.63 (1.4)	0.26 (0.96)



Figure 32. Example of BpS 16310-A stand in Ahtna Traditional Use Territory (LANDFIRE Photo).

Boreal Alpine Ericaceous Dwarf Shrubland - Complex (16351)

This BpS occurs across an estimated 1,219,920 acres in the Ahtna Traditional Use Territory. This is a dwarf shrub system that occurs from subalpine to alpine locations throughout Alaska. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 71,357 acres. This BpS typically occurs in alpine valleys and sideslopes, as well as low summits and ridges. Sites are usually well-drained and mesic to somewhat dry. These sites often retain snow late into the growing season which greatly influences moisture availability along with growing season length.

This BpS supports a wide range of dwarf shrub species, but *Ericaceous* or *Dryas* (especially *Dryas integrifolia* and/or *Dryas octopetala*) typically dominate. Other dwarf shrubs that commonly occur include *Empetrum nigrum, Vaccinium uliginosum, Harrimanella stellariana*, and *Arctostaphylos* spp. Other shrubs that may be common include *Betula nana, Diapensia lapponica, Dryas octopetala, Ledum palustre* ssp. *decumbens, Vaccinium vitis-idaea, Salix reticulata, Salix phlebophylla, Salix rotundifolia, Salix arctica* and *Oxytropis nigrescens*. Common herbaceous species include *Hierochloe alpina, Arnica lessingii, Carex bigelowii, Carex microchaeta, Senecio lugens, Minuartia arctica, Anemone parviflora, Ligusticum mutellinoides* ssp. *alpinum, Castilleja elegans, Poa arctica, Trisetum spicatum, Silene acaulis, Saxifraga* spp., *Campanula lasiocarpa, Anemone parviflora, Senecio lugens* and *Polygonum bistorta*. *Cassiope* and *Harimanella* tundra sites occur on terrain that is well protected by snow in the winter, and often remains snow covered until the middle of the growing season (Viereck et al. 1992).

These sites are not typically impacted by fire. The primary form of disturbance is continual wind, resulting in sites that are relatively stable over time. A site is categorized into this BpS due to having >20% cover of dryas dwarf shrubs and >25% total lichen cover (<u>LANDFIRE</u>). There is only one state in this BpS due to the vegetative stability.

Vegetation characteristics of the Boreal Alpine Ericaceous Dwarf-Shrubland – Complex BpS for plots sampled in the planning landscape are shown in Table 10. A photo depicting a vegetation stand in the Boreal Alpine Ericaceous Dwarf-Shrubland – Complex BpS is found in Figure 33.

Table 10. Vegetation characteristics from plots sampled in the Boreal Alpine Ericaceous Dwarf-Shrubland - Complex BpS for the planning landscape. GFS stands for grass, forb and seedling size class. Dwarf Shrub is shrub cover less than 1 m tall, medium shrub cover is shrubs 1-3 m in height and tall shrubs are >3m in height. Values are mean values with the standard deviation in parentheses.

CODE	SIZE-CLASS	DWARF SHRUB	MEDIUM SHRUB	TALL SHRUB	LICHEN
			% Cover (StDev)	
16351-A	GFS	50.03 (35.34)	34.63 (36.94)	12.49 (19.09)	3.17 (9.55)



Figure 33. Example of BpS 16351-A stand in Ahtna Traditional Use Territory (LANDFIRE Photo).

Alpine Dwarf Shrubland (16430)

This BpS occurs across an estimated 645,781 acres in the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 31,778 acres. This herbaceous and shrub BpS occurs on a variety of sites in subalpine and alpine habitats. Within the alpine zone snow cover can persist nearly year round, resulting in sites with a high proportion of exposed rock and soil. This BpS represents several existing vegetation types so species composition is highly variable. Shrub species may include: *Artemisia arctica*, *Cassiope mertensiana*, *Empetrum nigrum*, *Harrimanella*

stelleriana, Luetkea pectinata, Loiseleuria procumbens, Phyllodoce aleutica, Phyllodoce glanduliflora, Salix arctica, Salix reticulata, Salix rotundifolia, Sibbaldia procumbens, Vaccinium uliginosum and Vaccinium vitisidaea. Herbceous species may include: Aconitum delphiniifolium, Anemone narcissiflora, Astragalus alpinus, Athyrium filix-femina, Carex macrochaeta, Castilleja unalaschcensis, Chamerion angustifolium, Chamerion latifolium, Calamagrostis canadensis, Geranium erianthum, Lupinus nootkatensis, Minuartia arctica, Nephrophyllidium crista-galli, Saxifraga bracteata, Saxifraga bronchialis, Silene acaulis, Sanguisorba canadensis, Senecio triangularis, Valeriana sitchensis, Veratrum viride and Viola spp (LANDFIRE).

The lack of vegetation at these sites due to soil disturbances, snow avalanches, and wind disturbance results in a scarcity of fire and only a single described state. There can be variation among the vegetative species listed in the preceding paragraph.

Vegetation characteristics of the Alpine Dwarf-Shrubland BpS for plots sampled in the planning landscape are shown in Table 11. A photo depicting a vegetation stand in the Alpine Dwarf-Shrubland BpS is found in Figure 34.

Table 11. Vegetation characteristics from plots sampled in Alpine Dwarf-Shrubland BpS for the planning landscape. GFS stands for grass, forb and seedling size class. Dwarf Shrub is shrub cover less than 1 m tall, medium shrub cover is shrubs 1-3 m in height and tall shrubs are >3m in height. Values are mean values with the standard deviation in parentheses.

CODE	SIZE-CLASS	DWARF SHRUB	TALL SHRUB	LICHEN	
			% Cover (StDev)	
16430-A	GFS	74.61 (43.56)	20.68 (24.15)	10.17 (14.06)	1.35 (5.88)



Figure 34. Example of BpS 16430-A stand in Ahtna Traditional Use Territory (LANDFIRE Photo).

Reference Conditions and Current Conditions

The landscape modeling done by the LANDFIRE team also makes it possible to calculate the acres in each BpS under historical reference conditions. In addition, the percentages within each disturbance class that were estimated by the models makes it possible to compare historical reference conditions and existing conditions. It should be noted that caution should be used when comparing historical and existing conditions acres. For both conditions the acreages are only estimates based on modeling and remote sensing. At this time an accuracy assessment has not been completed for either data source. Table 12 shows percentages of the landscape and acres of upland grass/shrub types for the entire project area and Table 13 shows percentages and acres for upland grass/shrub types on Ahtna lands within the project area.

Table 12. Acres and percentages for each disturbance state in upland grass/shrub types for historical reference conditions and existing conditions in the Ahtna Traditional Use Territory.

Estimated Historical Reference Conditions - Upland Grass/Shrub Types

	<u>D</u>	<u> Disturba</u>	ance St	<u>ate (%)</u>	<u>l</u>		<u>Disturbance State (acres)</u>							
BpS	Α	В	C	D	E	BpS	Α	В	С	D	E			
16102	5	95	-	-	-	16102	135,600.15	2,576,402.79	N/A	N/A	N/A			
16280	10	75	15	-	-	16280	50,803.94	381,029.52	76,205.90	N/A	N/A			
16310	100	-	-	-	-	16310	243,965.98	N/A	N/A	N/A	N/A			
16351	100	-	-	-	-	16351	1,219,920.23	N/A	N/A	N/A	N/A			
16430	100	-	-	-	-	16430	645,781.73	N/A	N/A	N/A	N/A			

Existing Conditions - Upland Grass/Shrub Types

	<u>C</u>	<u> Disturba</u>	ance Sta	ate (%)	<u> </u>		<u>Disturbance State (acres)</u>						
BpS	Α	В	С	D	E	BpS	Α	В	С	D	E		
16102	12.4	87.6	-	-	-	16102	336,463.19	2,375,539.75	N/A	N/A	N/A		
16280	36.6	51.7	11.6	-	-	16280	186,194.66	262,882.00	58,962.70	N/A	N/A		
16310	100.0	-	-	-	-	16310	243,965.98	N/A	N/A	N/A	N/A		
16351	100.0	-	-	-	-	16351	1,219,920.23	N/A	N/A	N/A	N/A		
16430	100.0	-	-	-	-	16430	645,781.73	N/A	N/A	N/A	N/A		

Table 13. Acres and percentages for each disturbance state in upland grass/shrub types for historical reference conditions and existing conditions on lands owned by Ahtna, Inc.

	<u>C</u>	<u> Disturba</u>	ance Sta	ate (%)			<u>Disturbance State (acres)</u>						
BpS	Α	В	C	D	E	BpS	Α	В	С	D	E		
16102	5	95	-	-	-	16102	6,435.24	122,269.48	N/A	N/A	N/A		
16280	10	75	15	-	-	16280	3,720.04	27,900.32	5,580.06	N/A	N/A		
16310	100	-	-	-	-	16310	16,485.24	N/A	N/A	N/A	N/A		
16351	100	-	-	-	-	16351	71,357.01	N/A	N/A	N/A	N/A		
16430	100	-	-	-	-	16430	31,777.70	N/A	N/A	N/A	N/A		

Existing Conditions - Upland Grass/Shrub Types

	<u>C</u>	<u> Disturba</u>	ance Sta	ate (%)			<u>Disturbance State (acres)</u>						
BpS	Α	В	C	D	E	BpS	Α	В	С	D	E		
16102	23.1	76.6	-	-	-	16102	29,747.32	98,546.87	N/A	N/A	N/A		
16280	63.5	27.5	8.9	-	-	16280	23,621.01	10,228.61	3,326.14	N/A	N/A		
16310	100.0	-	-	-	-	16310	16,485.46	N/A	N/A	N/A	N/A		
16351	100.0	-	-	-	-	16351	71,283.20	N/A	N/A	N/A	N/A		
16430	100.0	-	-	-	-	16430	31,777.70	N/A	N/A	N/A	N/A		

Riparian and Wetland Ecosystems

Riparian and wetland BpS's have vegetation that is influenced by the presence of surface or subsurface water either year-round or seasonally. These sites occur along rivers and lakes and over sites with shallow permafrost. Within the Ahtna Traditional Use Territory there are 2.5 million acres that are classified as riparian and wetland BpS's.

Successional states within a given BpS are defined by the disturbance processes, the size and cover of the vegetation, and the plant species present. In riparian and wetland systems the primary disturbance types are flooding, thermokarst, and wildfire. Thermokarst occurs in areas underlain with permafrost and is a result of heaving caused by freezing and thawing. Wildfire occurs less frequently in riparian and wetland systems but will occur in stands of dwarf black spruce. In general, disturbance returns a forested successional state to a grass/shrub state. Ecosystem diversity matrices that display successional states for riparian and wetland BpS's in the Ahtna Traditional Use Territory are displayed for MLRA 222 in figures 35 and 36, MLRA 223 in figures 37 and 38, MLRA 227 in figures 39 and 40, and MLRA 228 in figures 41 and 42.

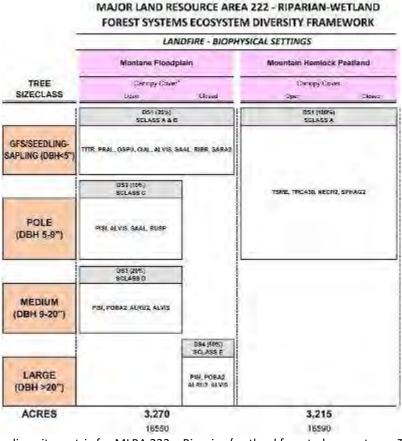


Figure 35. Ecosystem diversity matrix for MLRA 222 – Riparian/wetland forested ecosystems. The percentages in parenthesis are the estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. ^aOpen canopy cover has a value from 10-59% and closed canopy cover has a value from 60-100%.



Figure 36. Ecosystem diversity matrix for MLRA 222 – Riparian/wetland grass-shrub ecosystems. The percentages in parenthesis are the estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. ^aOpen canopy cover has a value from 10-59% and closed canopy cover has a value from 60-100%.

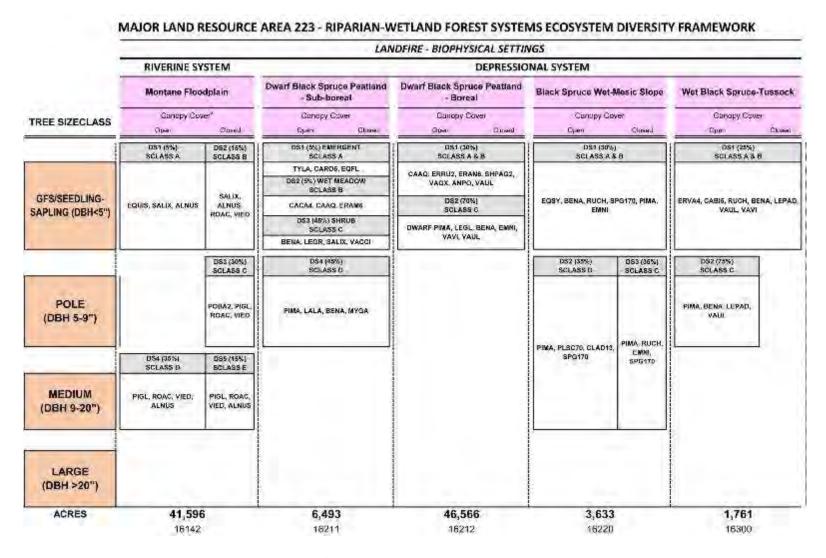


Figure 37. Ecosystem diversity matrix for MLRA 223 – Riparian/wetland forested ecosystems. The percentages in parenthesis are the estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. ^aOpen canopy cover has a value from 10-59% and closed canopy cover has a value from 60-100%.

MAJOR LAND RESOURCE AREA 223 - RIPARIAN-WETLAND GRASS-SHRUB SYSTEMS ECOSYSTEM DIVERSITY FRAMEWORK

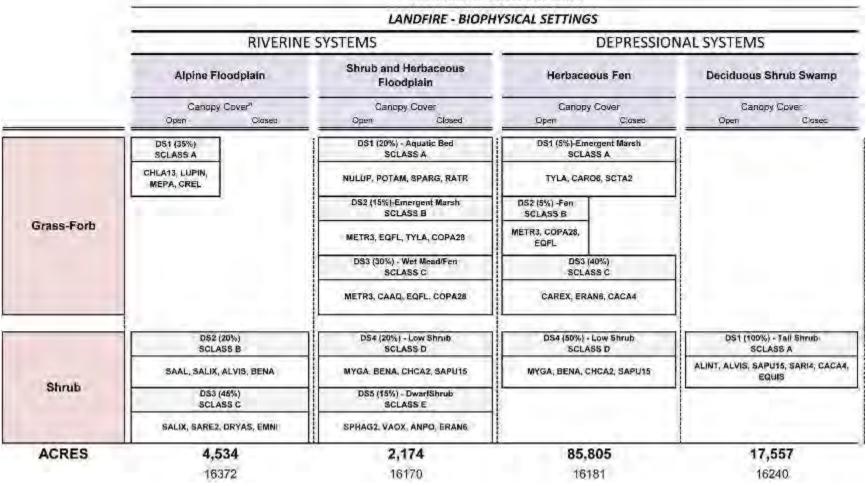


Figure 38. Ecosystem diversity matrix for MLRA 223 – Riparian/wetland grass-shrub ecosystems. The percentages in parenthesis are the estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. ^aOpen canopy cover has a value from 10-

59% and closed canopy cover has a value from 60-100%.

MAJOR LAND RESOURCE AREA 227 - RIPARIAN-WETLAND FOREST SYSTEMS ECOSYSTEM DIVERSITY FRAMEWORK

		LANDFIRE - BIOPHYSICAL SETTINGS											
			RIVERINE SY:	STEMS			DEPRESSIONAL	SYSTEMS					
	Montane Floo	dplain	Large River Flo	odplain	Riparian Stringer	Dwarf Black Spruce Peatland - Boreal	Black Spruce Wet-Mesic Slope		Wet Black Spruce	-Tussock			
TREE SIZECLASS	Canopy Cov		Canopy Cov		Canopy Cover	Canopy Cover	Canopy Co		Canopy Cov				
	Open	Closed	Open	Closed	Open Closed	Open Closed	Open	Closed	Open	Closed			
	DS1 (5%) SCLASS A	DS2 (20%) SCLASS B	DS1(5%) SCLASS A	DS2 (10%) SCLASS B	DS1 (10%) SCLASS A	DS1 (5%) EMERGENT SCLASS A	DS1 (30% SCLASS A		DS1 (25%) SCLASS A 8				
GFS/SEEDLING- SAPLING (DBH<5")	EQUIS, SALIX, ALNUS	SALIX, ALNUS, ROAC, VIED	EQUIS, SALIX, POBA2, LUP IN	SALIX, ALNUS, POBA2, ROAC	SALIX, ALNUS, CAREX, CACA4	TYLA, CARO6, EQFL DS2 (5%) WET MEADOW SCLASS B CACA4. CAAQ. ERAM6 DS3 (45%) SHRUB SCLASS C BENA, LEGR, SALIX, VACCI	EQSY, BENA, RUCI PIMA, EM		ERVA4, CABI5, RU LEPAD, VAUL,				
		DS3 (40%) SCLASS C	DS3 - HARD ((50%) C	DS2 (40%) SCLASS B	DS4 (45%) SCLASS D	DS2 (35%) SCLASS D	DS3 (35%) SCLASS C	DS2 (75%) SCLASS C				
POLE (DBH 5-9")		POBA2, PIGL, ROAC, VIED	POBA2, PIGL, RO DS4 - NO HA PIGL, ROAC, VIEE	ARD	BEPA, PIGL, PIMA, ALNUS	PIM A, LALA, BENA, MYGA	PIMA, PLSC70,	PIMA, RUCH,	PIMA, BENA, LEPAD, VAUL				
	DS4 (25%) SCLASS D			DS6 (10%) SCLASS E	DS3 (40%) SCLASS C		CLAD13, SPG170 EM NI, SPG170						
MEDIUM (DBH 9-20")	PIGL, ROAC, VIED, ALNUS	PIGL, ROAC, VIED, ALNUS	PIGL, ROAC, VIED, ALNUS	PIGL, ROAC, VIED, ALNUS	PIGL, BEPA, POBA2, PIMA								
LARGE (DBH >20")													
ACRES	259,88	2	3,491		2,497	673,384	8,062		18,384				
	16141		16150		16160	16211	16211 16220			16300			

Figure 39. Ecosystem diversity matrix for MLRA 227 – Riparian/wetland forested ecosystems. The percentages in parenthesis are the estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. ^aOpen canopy cover has a value from 10-59% and closed canopy cover has a value from 60-100%.

MAJOR LAND RESOURCE AREA 227 - RIPARIAN-WETLAND GRASS-SHRUB SYSTEMS ECOSYSTEM DIVERSITY FRAMEWORK

	LAI	NDFIRE - BIOPHYSICAL SETTII	vgs
	RIVERINE SYSTEMS	DEPRESSIO	NAL SYSTEMS
	Shrub and Herbaceous Floodplain	Deciduous Shrub Swamp	Herbaceous Fen
	Canopy Cover ^a Open Closed	Canopy Cover Open Closed	Canopy Cover Open Closed
	DS1 (20%) - A quatic Bed SCLASS A		DS1 (5%)-Emergent Marsh SCLASS A
	NULUP, POTAM, SPARG, RATR		TYLA, CARO6, SCTA2
	DS2 (15%)-Emergent Marsh SCLASS B		DS2 (5%) -Fen SCLASS B
Grass-Forb	METR3, EQFL, TYLA, COPA28		METR3, COPA28, EQFL
	DS3 (30%) - Wet Mead/Fen SCLASS C		DS3 (40%) SCLASS C
	METR3, CAAQ, EQFL, COPA28		CAREX, ERANG, CACA4
	DS4 (20%) - Low Shrub SCLASS D	DS1 (100%) - Tall Shrub SCLASS A	DS4 (50%) - Low Shrub SCLASS D
	MYGA, BENA, CHCA2, SAPU15	ALINT, ALVIS, SAPU15, SARI4, CACA4, EQUIS	M YGA, BENA, CHCA2, SAPU15
Shrub	DS5 (15%) - DwarfShrub SCLASS E		
	SPHAG2, VAOX, ANPO, ERAN6		
ACRES	8,549	4,301	1,127
	16170	16240	16181

Figure 40. Ecosystem diversity matrix for MLRA 227 – Riparian/wetland grass-shrub ecosystems. The percentages in parenthesis are the estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. ^aOpen canopy cover has a value from 10-59% and closed canopy cover has a value from 60-100%.

MAJOR LAND RESOURCE AREA 228- RIPARIAN-WETLAND FOREST SYSTEMS ECOSYSTEM DIVERSITY FRAMEWORK

		LANDFIRE - BIOPHYSICAL SETTINGS												
				RIVERINE	SYSTEMS				DEPRESSIOI	NAL SYSTEMS				
	Montane Floodpl	ain - Boreal	Montane Floor	-	Large River Flo	odplain	Riparian Stringer	Dwarf Black Spruce Dwarf Black Spruce Peatland - Boreal Peatland - Subboreal		Black Spruce Wet-Mesic Slope	Wet Black Spruce-Tussock			
TREE	Canopy Co	over ^a	Canopy Co	ver	Canopy Co	ver	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover	Canopy Cover			
SIZECLASS	Open	Closed	Open	Closed	Open	Closed	Open Closed	Open Closed	Open Closed	Open Closed	Open Closed			
	DS1 (5%) SCLASS A	DS2 (20%) SCLASS B	DS1(5%) SCLASS A	DS2 (15%) SCLASS B	DS1(5%) SCLASS A	DS2 (10%) SCLASS B	DS1 (10%) SCLASS A	DS1 (5%) EMERGENT SCLASS A	DS1 (30%) SCLASS A & B	DS1 (30%) SCLASS A & B	DS1 (25%) SCLASS A & B			
GFS/SEEDLING-		SALIX,		SALIX,		SALIX,		TYLA, CARO6, EQFL DS2 (5%) WET MEADOW SCLASS B	CAAQ, ERRU2, ERAN6, SHPAG2, VAOX, ANPO, VAUL		ERVA4, CABI5, RUCH, BENA, LEPAD, VAUL, VAVI			
SAPLING	EQUIS, SALIX, ALNUS	ALNUS, ROAC,	EQUIS, SALIX, ALNUS	ALNUS, ROAC,	EQUIS, SALIX, POBA2, LUPIN	ALNUS, POBA2,	SALIX, ALNUS, CAREX, CACA4	CACA4. CAAQ. ERAM 6	DS2 (70%) SCLASS C	EQSY, BENA, RUCH, SPG170, PIMA, EMNI				
(DBH<5")		VIED		VIED		ROAC		DS3 (45%) SHRUB SCLASS C	DWARF PIMA, LEGL, BENA,	·				
								BENA, LEGR, SALIX, VACCI	EM NI, VAVI, VAUL					
		DS3 (40%) SCLASS C		DS3 (30%) SCLASS C	DS3 - HARD SCLASS		DS2 (40%) SCLASS B	DS4 (45%) SCLASS D		DS2 (35%) DS3 (35%) SCLASS D SCLASS C	DS2 (75%) SCLASS C			
	POBA2,	POBA2,		AC, VIED										
POLE (DBH 5-9")		POBA2, PIGL, ROAC, VIED	PIGL, ROAC, VIED	DS4 - NO HARD		BEPA, PIGL, PIMA, ALNUS	PIMA, LALA, BENA, MYGA			PIMA, BENA, LEPAD, VAUL				
					PIGL, ROAC, VIED, ALNUS					PIMA, PLSC70, RUCH,				
_	DS4 (25%) SCLASS D	DS5 (10%) SCLASS E	DS4 (35%) SCLASS D	DS5 (15%) SCLASS E	DS5 (25%) SCLASS D	DS6 (10%) SCLASS E	DS3 (40%) SCLASS C			CLAD 13, SPG 170 EM NI, SPG 170				
MEDIUM (DBH 9-20")	PIGL, ROAC, VIED ALNUS	PIGL, ROAC, VIED, ALNUS	PIGL, ROAC, VIED, ALNUS	PIGL, ROAC, VIED, ALNUS	PIGL, ROAC, VIED, ALNUS	PIGL, ROAC, VIED, ALNUS	PIGL, BEPA, POBA2, PIMA							
LARGE	_													
(DBH >20")														
	248,1		169,77		23,427		3,106	515,641	173,703	83,545	28,868			
	16141	1	16142		16150		16160	16211	16212	16220	16300			

Figure 41. Ecosystem diversity matrix for MLRA 228 – Riparian/wetland forested ecosystems. The percentages in parenthesis are the estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. ^aOpen canopy cover has a value from 10-59% and closed canopy cover has a value from 60-100%.

MAJOR LAND RESOURCE AREA 228 - RIPARIAN-WETLAND GRASS-SHRUB SYSTEMS ECOSYSTEM DIVERSITY FRAMEWORK

		ECOSTSTEIN DIV	LIGHT HIN HALL		
		LANDFIRE - BIO	PHYSICAL SETTINGS		
	RIVERIN	E SYSTEMS	DEPRESSIC	NAL SYSTEMS	
	Shrub and Herbaceous Floodplain	Alpine Floodplain	Deciduous Shrub Swamp	Herbaceous Fen	
	Canopy Cover ^a Open Closed	Canopy Cover Open Closed	Canopy Cover Open Closed	Canopy Cover Open Closed	
	DS1 (20%) - Aquatic Bed SCLASS A	DS1 (35%) SCLASS A		DS1 (5%)-Emergent Marsh SCLASS A	
	NULUP, POTAM, SPARG, RATR	CHLA 13, LUPIN, MEPA, CREL		TYLA, CARO6, SCTA2	
	DS2 (15%)-Emergent Marsh SCLASS B			DS2 (5%) -Fen SCLASS B	
Grass-Forb	METR3, EQFL, TYLA, COPA28			METR3, COPA28, EQFL	
	DS3 (30%) - Wet Mead/Fen SCLASS C			DS3 (40%) SCLASS C	
	METR3, CAAQ, EQFL, COPA28			CAREX, ERAN6, CACA4	
	DS4 (20%) - Low Shrub SCLASS D	DS2 (20%) SCLASS B	DS1 (100%) - Tall Shrub SCLASS A	DS4 (50%) - Low Shrub SCLASS D	
	MYGA, BENA, CHCA2, SAPU15	SAAL, SALIX, ALVIS, BENA	ALINT, ALVIS, SAPU15, SARI4, CACA4, EQUIS	MYGA, BENA, CHCA2, SAPU1	
Shrub	DS5 (15%) - DwarfShrub SCLASS E	DS3 (45%) SCLASS C			
	SPHAG2, VAOX, ANPO, ERAN6	SALIX, SARE2, DRYAS, EM NI			
ACRES	3,182	2,040	119,134	230,095	
	16170	16372	16240	16181	

Figure 42. Ecosystem diversity matrix for MLRA 228 – Riparian/wetland grass-shrub ecosystems. The percentages in parenthesis are the estimates of the percentage of each BpS that occurred in each specific successional state. The bold number at the bottom of each column represents the total acreage for a biophysical setting within the MLRA. The number below that is the biophysical setting code assigned by LANDFIRE. ^aOpen canopy cover has a value from 10-59% and closed canopy cover has a value from 60-100%.

Montane Floodplain Forest and Shrubland-Boreal (16141)

This BpS is estimated to occur on approximately 498,246 acres of the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 73,647 acres. It typically occurs on well drained sand or cobble without permafrost (LANDFIRE). Floods are the primary disturbance to this site, depositing alluvium that is then invaded by various species. Early seral species include balsam poplar and white spruce along with *Alnus viridi*, *Alnus incana*, *Salix barclayi* and *Salix alaxensis* (Boggs 2000, Shephard 1995, Thilenius 1990). This site may burn, but fire return intervals are typically >300 years, with flooding being the primary disturbance unless oxbows or other changes have isolated this type away from an active riverine system.

Following alluvial deposition, herbaceous species emerge from seeds including *Lupinus* spp., *Hedysarum* spp., *Equisetum* spp. and *Salix* spp. By year 5 willows and balsam poplar are typically well established and shrub cover may be up to 40%. From year 5-29, willows and alders will predominate along with balsam poplar saplings and white spruce as an understory species. This state may comprise 20% of the BpS, as flooding continues to provide disturbance to floodplain settings (LANDFIRE). This state can provide valuable foraging sites for moose.

The next state, occurring from 30-149 years historically represented 40% of the BpS and is characterized by closed canopies of mature balsam poplar mixed with white spruce. At later ages, the balsam popular may start to die increasing shrub understories that may become restricted to shade tolerant species under the younger, dense canopy (LANDFIRE). This state has a total carbon availability of 109.46 tons/acre with an annual carbon production of 1.28 tons/acre. There are approximately 54 tons of biomass available per acre.

Late seral conditions (>150 years post flood establishment) historically comprised 20% of the BpS. This state is comprised of increasing dominance of white spruce that will occur in open canopies. On some sites, a closed canopy of white spruce may develop, favoring feathermoss in the understory. This state has a total carbon availability of 141.6 tons/acre with an annual carbon production of 0.34 tons/acre. There are approximately 68 tons of biomass available per acre.

Vegetation characteristics of the Montane Floodplain Forest and Shrubland-Boreal BpS for the planning landscape are listed in Table 14. Photos depicting different vegetation stands in the Montane Floodplain Forest and Shrubland-Boreal BpS are found in Figures 43 and 44.

Table 14. Vegetation characteristics from plots sampled in the Montane Floodplain Forest and Shrubland-Boreal BpS for the planning landscape. Seed/Sap stands for seedling-sapling size class with <60% canopy cover, Seed/Sap closed refers to the seedling-sapling size class with >60% cover, Pole stands for pole size class (5-9" DBH trees), medium open refers to the medium size class (9-20" DBH) with <60% canopy cover, medium closed refers to the medium size class with >60% canopy cover, Large refers to large trees (>20" DBH). Dwarf Shrub is shrub cover less than 1 m tall, medium shrub cover is shrubs 1-3 m in height and tall shrubs are >3m in height. SS refers to seedlings and saplings. BA refers to basal area of trees. TPA refers to trees per acre. Values are mean values with the standard deviation in parentheses.

CODE	SIZE CLASS	DWARF SHRUB	MEDIUM SHRUB	TALL SHRUB	LICHEN	SEED/ SAPLING	POLE	MEDIUM	LARGE	SEED/ SAPLING	POLE	MEDIUM	LARGE
			% Cove	(StDev)		TPA (StDev)				Basal Area (StDev)			
16141-A	SEED/SAP-OPEN	18 (0)	21 (0)	18 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)
16141-B	SEED/SAP-CLOSED	2.8 (3.2)	3.3(3.2)	58 (2.8)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)
16141-D	MEDIUM-OPEN	10 (0)	6.7 (0)	0 (0)	0 (0)	210 (0)	170 (0)	60 (0)	0 (0)	91.6 (0)	88.9 (0)	57.2 (0)	0 (0)
16141-E	MEDIUM-CLOSED	13.8 (8.8)	42.5 (3.5)	28.8 (12.4)	0 (0)	410 (410.1)	90 (14)	25 (21.2)	10 (0)	61.9 (15.3)	48.0 (0.7)	31.0 (2.7)	32.9 (0)



Figure 43. Example of BpS 16141-C stand in Ahtna Traditional Use Territory.



Figure 44. Example of BpS 16141-D stand in Ahtna Traditional Use Territory.

Montane Floodplain Forest and Shrubland- Sub-boreal (16142)

This BpS is estimated to occur on approximately 204,372 acres of the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 18,898 acres. It typically occurs on well drained sand or cobble without permafrost (LANDFIRE). Floods are the primary disturbance to this site, depositing alluvium that is then invaded by various species. Early seral species include balsam poplar and white spruce along with *Alnus viridis* ssp. *sinuata*, *Alnus incana* ssp. *Tenuifolia*, *Salix barclayi and Salix alaxensis* (Boggs 2000, Scott 1974, Shephard 1995, Thilenius 1990, Viereck 1966). This site may burn, but fire return intervals are typically >300 years, with flooding being the primary disturbance unless oxbows or other changes have isolated this type away from an active riverine system.

Following alluvial deposition, herbaceous species emerge from seeds including *Lupinus* spp., *Hedysarum* spp., *Equisetum* spp. and *Salix* spp. By year 5 willows and balsam poplar are typically well established and shrub cover may be up to 40%. This state historically comprised 5% of the BpS.

From year 5-29, willows and alders will predominate along with balsam poplar saplings and white spruce as an understory species. This state may comprise 15% of the BpS, as flooding continues to provide disturbance to floodplain settings. This state can provide valuable foraging sites for moose. Common shrubs include *Alnus viridis* ssp. *sinuata*, *Alnus incana* ssp. *tenuifolia*, *Salix barclayi* and *Salix alaxensis*. On dryer sites *Shepherdia canadensis*, *Dryas octopetala*, *D. integrifolia* and fruticose lichens (*Steroucaulon* spp.) may be more dominant (<u>LANDFIRE</u>).

The next state, occurring from 30-149 years historically represented 30% of the BpS and is characterized by closed canopies of mature balsam popular mixed with white spruce. At later ages, the balsam popular may start to die leading to increasing shrub understories that may become restricted to shade tolerant species under the younger, dense canopy. These shrub species include *Rosa acicularis* and *Viburnum edule*. An alternative mid-seral stage is comprised of closed stands of white spruce. These stands occur in the absence of balsam popular regeneration following disturbance. Feather moss is often dominant in the understory (<u>LANDFIRE</u>). Approximately 15% of the BpS was in this state historically. Both of these states have a total carbon availability of 109.46 tons/acre with an annual carbon production of 1.28 tons/acre. There are approximately 54 tons of biomass available per acre.

Late seral conditions (>150 years post flood establishment) historically comprised 35% of the BpS. These conditions occur as white spruce gains dominance over balsam popular. This results in a mixed-age spruce stand with a relatively open canopy between 25% and 60% cover (LANDFIRE). This state has a total carbon availability of 141.6 tons/acre with an annual carbon production of 0.34 tons/acre. There are approximately 68 tons of biomass available per acre.

Boreal Herbaceous Fen – Sub-boreal (16181)

This BpS occurs across an estimated 311,353 acres in the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 22,287 acres. This wetland system occurs in permafrost free areas throughout Alaska. It is typically found in drainages and along pond margins. In early successional states sites may be semi-permanently flooded, and seasonal flooding is common in wetter states. In later successional states, soils are saturated for a portion of the growing season. Due to the wet nature of these sites fire is extremely rare. The primary factor driving succession is hydrology and the associated changes in water level or frequency. Major distributions to a site's hydrology can also cause a site to transition to a different BpS.

The earliest state in this BpS is a freshwater marsh, dominated by *Typha latifolia* and *Carex rostrata*. Other species that may dominate include *Carex utriculata*, *Schoenoplectus tabernaemontani*, *Arctophila fulva*, *Eleocharis* palustris, *Myriophyllum spicatum*, *Menyanthes trifoliata*, *Comarum palustris*, *Hippuris vulgaris*, and *Equistem fluviatile* (Jorgenson 1999). The state usually only lasts for a few years before transitioning to the next state.

The second state, which commonly occurs in years 3 to 5, is herbaceous fen. Commons species include *Menyanthes trifoliata, Comarum palustre, Equisetum fluviatile, Potentilla palustris,* and *Carex aquatilis*. The state is steady with frequent flooding, while increased water levels will return site to earlier state, and drying will move it to a later state (LANDFIRE).

The third state is a wet meadow and occurs from 6 to 25 years. Dominant species may include *Carex aquatilis*, *Carex livida*, *Carex chordorrhiza*, *Carex lasiocarpa*, *Eriophorum angustifolium*, *Calamagrostis canadensis*, *Comarum palustre*, *Menyanthes trifoliata*, *Equisetum fluviatile*, and *Equisetum palustre*. Shrubs can be a minor component of this state and include *Myrica gale*, *Alnus incana* spp. *tenuifolia* and *Salix* spp (NatureServe 2008). As conditions continue to become drier this site will move to a later state, while wetter conditions may return it to an earlier state.

The climax state is a low shrub peatland and occurs beyond 26 years. Common species include *Ledum* palustre, *Ledum groenlandicum*, *Betula nana*, *Rubus chamaemorus*, *Oxycoccus microcarpus*, *Myrica gale*, *Calamagrostis canadensis*, *Carex aquatilis*, *Comarum palustre*, *Salix fuscescens*, *Salix pulchra*, *Empetrum nigrum* and *Chamaedaphne calyculata*. *Sphagnum* spp. and/or brown mosses may be common in the ground layer (DeVelice et al. 1999, Jorgenson et al 2003). This state is stable in the absence of disturbance. Wetter conditions will cause it to transition back to an earlier state.

Vegetation characteristics of the Boreal Herbaceous Fen – Alaska Sub-Boreal Complex BpS for plots sampled in the planning landscape are shown in Table 15. Figure 45 is an example of a stand in the Boreal Herbaceous Fen – Alaska Sub-Boreal Complex BpS.

Table 15. Vegetation characteristics from plots sampled in the Boreal Herbaceous Fen BpS for the planning landscape. GFS stands for grass, forb and seedling size class. Dwarf Shrub is shrub cover less than 1 m tall, medium shrub cover is shrubs 1-3 m in height and tall shrubs are >3m in height. Values are mean values with the standard deviation in parentheses.

CODE	SIZE-CLASS	SIZE-CLASS DWARF MEDIUM SHRUB SHRUB		TALL SHRUB	LICHEN			
		% Cover (StDev)						
16181-A	WET MEADOW	0.46 (1.59)	1.16 (2.13)	0.9 (2.03)	0 (0)			
16181-D	GFS	2.24 (2.6)	24.16 (26.76)	1.53 (2.79)	0 (0)			



Figure 45. Example of BpS 16181-B stand in Ahtna Traditional Use Territory (LANDFIRE Photo).

Dwarf Black Spruce Peatland- Boreal (16211)

This BpS covers an estimated 1,175,753 acres of the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 315,538 acres. It occurs in valley bottoms and poorly drained floodplains on acidic soils with well-developed peat (LANDFIRE 2009, Nature Serve 2008).

Van Cleve et al. (1983) described taiga forests of interior Alaska, and stated that the role of fire was the predominant influence on vegetation patterns and compositions. Hollingsworth et al. (Hollingsworth et al. 2006) reported that "species composition of mature black spruce forests in interior Alaska results from the complex interaction of landscape and fire history, soil pH, paludification, permafrost, and topographic position." Van Cleve et al. (1983) stated that after severe fires where the forest floor was consumed, deciduous vegetation, specifically aspen and birches would take over previous black spruce sites, while black spruce could maintain its composition following lower severity fires. They reported that the productivity of black spruce sites generally increased for at least 10-20 years following fire as soil temperatures were increased with reductions in organic layers.

Barrett et al. (2011) reported that high severity fire in black spruce that left <10 cm of an organic layer increased the likelihood of a stand shifting from being spruce dominated to hardwood dominated. They found that this also increased the likelihood of the melting of permafrost under the site. They estimated that of the black spruce areas that burned in 2004, 39% could experience either a reduction in permafrost or a shift to deciduous vegetation. Similarly, Bernhardt et al. (2011) found that fire severity was the most important contributor to post-fire plant communities, being more important than preburn site moisture or soil pH.

Hollingsworth et al. (2013) found that fire severity accounted for the greatest percentage of post-fire vegetation variability in black spruce stands followed by soil pH and then an environmental gradient associated with elevation and assumed climate differences. Johnstone et al. (2009) found that black spruce could reestablish on sites that experienced low severity fires while deciduous species established on sites with higher fire severities where the organic layer was largely removed. Johnstone et al. (2011) clarified that it is the influence of fire on the organic layer as opposed to the influence on the overstory vegetation that largely drives the post-fire response of the plant community in black spruce forests. Thus, fire severity in black spruce should be considered the degree to which the organic layer on the forest floor is impacted by the fire, not just the impacts on the overstory plant community.

Shenoy et al. (2011) found that the effects of fire on the organic layer persisted for at least the first two decades following high severity fires, with deciduous trees maintaining their dominance on severely burned sites while black spruce gained increasing dominance on sites where the organic layer was not as disturbed. Turetsky et al. (2010) reported that mosses displayed a unimodal response following fire in boreal forests, peaking in amounts 30-70 years post-fire. Mann et al. (2012) speculated that given the shift in species composition from black spruce dominated stands to deciduous stands following severe fire, that major changes to black spruce forests and associated ecosystem services such as carbon storage are likely with predicted climate change. Similarly, Scheffer et al. (2012) reported that climate change may cause massive shifts in boreal biomes as a consequence of changing fire frequencies and severities.

The Dwarf Black Spruce Peatland- Boreal BpS typically occurs on permafrost. This is maintained by the thick layer of peat on the soil surface. This layer of peat keeps the site cool, reduces decomposition, and insulates the permafrost. Viereck et al. (1983) reported that these black spruce sites are nutrient poor and unproductive with these characteristics caused by the low soil temperature and high soil moisture. If severe fire occurs removing the layer of peat, the site may warm and melt the permafrost (thermokarst). This can cause the site to lower, changing it to a wetland-marsh system. On some sites, over time, the peat will again rebuild, and can lead to the permafrost reforming. With the rise of the site with freezing, the Black Spruce site may be reestablished, however, there are numerous possible pathways for these marsh systems to follow (LANDFIRE 2009).

Fire return intervals are estimated to be >150 years in the Copper River Basin. Where peatland persists, the site will first be a low shrub peatland with *Ledum groenlandicum*, *L. palustre*, *Andromeda polifolia*, *Vaccinium uliginosum*, *Salix* spp., *Betula nana*, and *Empetrum nigrum* as primary species. This can persist up to 74 years and historically occupied 45% of the BpS. Sites >75 years, dominated by dwarf black spruce, were estimated to historically occur on >45% of this BpS. This BpS has very minimal carbon with stands containing up to 22 tons per acre and annual carbon production of 0.01 tons per acre. Estimates of biomass have not been done for this BpS.

Vegetation characteristics for the Dwarf Black Spruce Peatland-Boreal BpS are listed in Table 16. Figure 46 is an example of a stand in the Dwarf Black Spruce Peatland-Boreal BpS.

Table 16. Vegetation characteristics from plots sampled in the Dwarf Black Spruce-Boreal BpS from the planning landscape. GFS stands for grass, forb and seedling size class, Seed/Sap stands for seedling-sapling size class, Pole stands for pole size class (5-9" DBH trees), medium refers to the medium size class (9-20" DBH). Dwarf Shrub is shrub cover less than 1 m tall, medium shrub cover is shrubs 1-3 m in height and tall shrubs are >3m in height. SS refers to seedlings and saplings. BA refers to basal area of trees. TPA refers to trees per acre. Values are mean values with the standard deviation in parentheses.

CODE	STRUCTURE/ SIZE-CLASS	DWARF SHRUB	MEDIUM SHRUB	TALL SHRUB	TALL SHRUB LICHEN		POLE	MEDIUM	SEED/ SAPLING	POLE	MEDIUM
			% Cove	er (StDev)			-TPA (StDev)		Basal Area (StDev)		
16211-B	WET MEADOW	4.2 (8.4)	14.9 (27.3)	3.7 (9.8)	1.4 (5.0)	1232.1 (550.5)	0 (0)	0 (0)	37.9 (18.8)	0 (0)	0 (0)
16211-C	POLE	14.1 (20.3)	27.5 (33.1)	5.4 (9.5)	1.2 (3.0)	889 (432.7)	210.1 (162.7)	38.7 (20.6)	65.8 (27.4)	37.9 (29.3)	28.4 (14.5)



Figure 46. Example of BpS 16211-B stand in Ahtna Traditional Use Territory (LANDFIRE Photo).

Dwarf Black Spruce Peatland- Sub-boreal (16212)

This BpS covers an estimated 219,354 acres of the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 29,498 acres. It occurs in valley bottoms and poorly drained floodplains on acidic soils with well-developed peat and permafrost is generally absent (Nature Serve 2008). The primary disturbance in these systems is changes in hydrology. Over time, thickening peat layers support low shrubs and eventually dwarf black spruce. For the first 50 years following disturbance, such as the melting of permafrost, or other hydrology changes an herbaceous disturbance state dominates. Conditions range from marsh to sedge meadow to sphagnum-dominated poor fen. Common species include *Carex aquatilis*, *C. rariflora*, *C. limosa*, *C. utriculata*, *Eriophorum russeolum*, *E. angustifolium*, *Sphagnum* spp., *Calamagrostis canadensis*, and *Equisetum fluviatile*. Under historical conditions this state represented 10% of the BpS.

As the peat layer thickens the sites move towards a dwarf shrub disturbance state. This can last up to 300 years. Common species include *Vaccinium oxycoccus*, *V. uliginosum*, *Andromeda polifolia*, *Ledum groenlandicum*, *Betula nana*, *Empetrum nigrum*, *Carex microglochin*, *C. rotundata*, *C. rariflora*, *C. lasiocarpa*, *C. livida*, *C. williamsii*, *Eriophorum brachyantherum*, *E. angustifolium* and *Drosera* spp. *Sphagnum* spp. form an abundant ground layer. This state was present across 20% of the BpS historically.

The climax state for this BpS is a dwarf black spruce bog. Tree cover ranges from 10-60%. Dominant understory species include *Ledum groenlandicum*, *Betula nana*, *Empetrum nigrum*, *Vaccinium vitisidaea*, *V. uliginosum*, *Chamaedaphne calyculata*, *Carex* spp., *Eriophorum angustifolium*, and *Sphagnum* spp. Under historical conditions this state represented 70% of the BpS.

Black Spruce Wet-Mesic Slope (16220)

This BpS covers an estimated 93,318 acres of the Ahtna Traditional Use Territory. On Ahtna Incorporated and Village controlled lands this BpS covers an estimated 18,119 acres. It occurs on north-facing slopes underlain by permafrost. Soils are poorly drained and acidic with a well-developed peat layer (NatureServe 2008). The dominant overstory vegetation is black spruce and the primary disturbance in these systems is fire. The mean fire return interval is 170 years, with considerable variation between sites. Since this type occurs on slopes, fire spread is possible, despite saturated soils.

From years 0-19 following fire the stand is a grass/shrub disturbance type. Common shrubs include Ledum groenlandicum, Ledum palustre, Betula nana, Empetrum nigrum, Vaccinium vitis-idaea and V. uliginosum. Herbs include Equisetum sylvaticum, Rubus chamaemorus and Carex spp. Mosses include Sphagnum spp., Pleurozium schreberi, and Polytrichum spp. On some sites, Alnus spp. and Salix spp. may be present. Black spruce seedlings are present in the understory (LANDFIRE). Under historical conditions this state represented 10% of the BpS.

By years 20-40 post fire the stand transition to a tree disturbance type. The class is dominated by seedling/sapling black spruce. The shrubs listed above are still present, but occur under a closed tree canopy. This state was present across 20% of the BpS historically.

By 40+ years post fire the stand is dominated by pole-sized black spruce. Many of the shrub species have begun to senesce due to the closed tree canopy. This class has the highest probability of stand replacing fire due to the high amount of fuel present on the site. Under historical conditions this state represented 35% of the BpS. This state has a total carbon availability of 71.4 tons/acre with an annual carbon production of 0.37 tons/acre. There are approximately 34 tons of biomass available per acre.

As the black spruce mature the canopy becomes more open and lichen species become more prevalent. Lichen species include *Cladina arbuscula*, *C. rangiferina* and *Nephroma articum*. Mosses include *Pleurozium schreberi*, *Polytrichum* spp., *Hylocomium splendens* and *Dicranum* spp., as well as *Sphagnum* spp. Low shrubs, including *Vaccinium vitis-idaea*, *V. uliginosum* and *Ledum groenlandicum*, are often present in the understory (<u>LANDFIRE</u>). The increasing lichen cover can make these stands important for wintering caribou. This state was present across 35% of the BpS historically. This state has a total carbon availability of 36.8 tons/acre with an annual carbon production of 0.22 tons/acre. There are approximately 18 tons of biomass available per acre.

Reference Conditions and Current Conditions

The landscape modeling done by the LANDFIRE team allows for an estimation of the acres in each BpS under historical reference conditions. In addition, the percentages within each disturbance class that were estimated by the models makes it possible to compare historical reference conditions and existing conditions. It should be noted that caution should be used when comparing historical and existing conditions acres. For both conditions the acreages are only estimates based on modeling and remote sensing. At this time an accuracy assessment has not been completed for either data source. Table 17 shows percentages of the landscape and acres of riparian types for the entire project area and Table 18 shows percentages and acres for riparian types on Ahtna lands within the project area.

Table 17. Acres and percentages for each disturbance state in riparian types for historical reference conditions and existing conditions in the Ahtna Traditional Use Territory.

	<u>D</u>	<u> Disturb</u>	ance St	tate (%	<u>.)</u>			<u>Distu</u>	<u>rbance State (a</u>	<u>cres)</u>	
BpS	Α	В	С	D	E	BpS	Α	В	С	D	E
16141	5	20	40	25	10	16141	24,912.31	99,649.24	199,298.48	124,561.55	49,824.62
16142	5	15	30	35	15	16142	10,218.59	30,655.78	61,311.57	71,530.16	30,655.78
16181	5	5	40	50	-	16181	15,567.63	15,567.63	124,541.02	155,676.28	N/A
16211	5	5	45	45	-	16211	58,787.65	58,787.65	529,088.87	529,088.87	N/A
16212	10	20	70	-	-	16212	21,935.38	43,870.75	153,547.63	N/A	N/A
16220	10	20	35	35	-	16220	9,331.81	18,663.61	32,661.32	32,661.32	N/A

Existing Conditions - Riparian/Wetland Types

<u>Disturbance State (%)</u>							<u>Disturbance State (acres)</u>						
BpS	Α	В	С	D	E	BpS	Α	В	С	D	E		
16141	46.1	13.9	6.4	31.2	2.4	16141	229,694.67	69,359.45	31,844.29	155,538.84	11,808.95		
16142	64.5	21.4	11.0	2.8	0.4	16142	131,730.78	43,702.18	22,440.99	5,752.02	745.91		
16181	25.8	27.2	0.5	46.5	-	16181	80,406.69	84,581.71	1,708.44	144,655.72	N/A		
16211	3.5	7.0	14.4	75.1	-	16211	41,330.33	82,176.73	169,148.97	883,097.00	N/A		
16212	9.4	47.0	43.6	-	-	16212	20,550.86	103,061.18	95,741.72	N/A	N/A		
16220	22.7	35.8	3.8	37.7	-	16220	21,174.23	33,388.61	3,532.52	35,222.70	N/A		

Table 18. Acres and percentages for each disturbance state in riparian types for historical reference conditions and existing conditions on lands owned by Ahtna, Inc.

Antina, mc.	•										
					<u>Estin</u>	nated Historical R	<u> eference Condi</u>	tions - Riparian/	Wetland Types		
	<u>C</u>	<u> Disturba</u>	nce St	tate (%)			<u>Di</u>	sturbance State (a	icres)	
BpS	Α	В	C	D	Ε	BpS	Α	В	С	D	E
16141	5	20	40	25	10	16141	3,682.35	14,729.39	29,458.78	18,411.74	7,364.70
16142	5	15	30	35	15	16142	944.89	2,834.67	5,669.33	6,614.22	2,834.67
16181	5	5	40	50	-	16181	1,114.34	1,114.34	8,914.74	11,143.43	N/A
16211	5	5	45	45	-	16211	15,776.91	15,776.91	141,992.20	141,992.20	N/A
16212	10	20	70	-	-	16212	2,949.76	5,899.51	20,648.30	N/A	N/A
16220	10	20	35	35	-	16220	1,811.94	3,623.88	6,341.79	6,341.79	N/A
						Existing Co	onditions - Ripa	rian/Wetland Ty	pes		
	<u>D</u>	Disturba	nce St	tate (%	<u>)</u>			<u>Di</u>	sturbance State (a	icres)	
BpS	Α	В	C	D	Ε	BpS	Α	В	С	D	E
16141	10.0	13.0	5.7	65.6	5.7	16141	7,392.19	9,564.54	4,231.73	48,276.83	4,188.36
16141	10.0	13.0	5./	65.6	5./	16141	7,392.19	9,564.54	4,231./3	48,276.83	

	<u>D</u>	<u>isturba</u>	ance St	ate (%)	<u>)</u>			<u>Dis</u>	<u>sturbance State (a</u>	acres)	
BpS	Α	В	C	D	E	BpS	Α	В	С	D	E
16141	10.0	13.0	5.7	65.6	5.7	16141	7,392.19	9,564.54	4,231.73	48,276.83	4,188.36
16142	58.0	20.6	15.0	5.4	1.0	16142	10,965.18	3,886.80	2,828.20	1,021.68	181.70
16181	33.3	12.8	1.5	52.4	-	16181	7,432.66	2,851.77	333.81	11,683.74	N/A
16211	3.8	4.2	9.3	82.7	-	16211	11,956.62	13,101.95	29,385.93	261,026.59	N/A
16212	8.7	30.4	60.9	-	-	16212	2,556.21	8,970.52	17,970.63	N/A	N/A
16220	21.5	25.5	3.0	50.1	-	16220	3,893.69	4,616.03	536.42	9,082.83	N/A

Regional (Village) Planning

Village Plans

The Ahtna Traditional Use Territory includes 7 villages that merged to form Ahtna, Inc along with the Chitina Native Corporation. Figure 47 shows the 8 planning regions. Each planning region will be discussed in detail in the following section. Several of the planning regions are located on the periphery of the Copper Basin and as a result have different climate and soils when compared to conditions described previously in this report. Additional detail on the unique biotic and abiotic conditions found in these planning regions will be discussed in more detail for each relevant village.

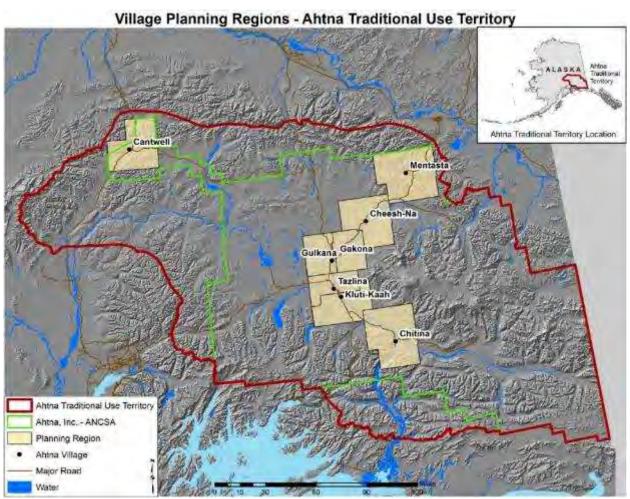


Figure 47. Village planning regions in the Ahtna Traditional Use Territory.

Site Improvement Areas

Site selection for improvement areas focused on two types of treatments. These were moose browse improvements and timber stand improvements. Moose browse improvements are intended to increase the foraging quality of a stand primarily by increasing the productivity of preferred willow species. This is done by using treatments that crush or cut mature willows to encourage regrowth and/or scarifying a site to stimulate willow seeding and regeneration.

Timber stand improvements are mainly intended to increase the growth rates and quality of timber, particularly in white spruce stands. Treatments consist of removing portions of the overstory along with diseased and misshapen trees to improve stand health and release remaining trees for increased growth. These treatments can also generate biomass and firewood for local communities. The selection criteria for improvement areas were similar for both types of treatments. The first consideration was that stands must occur on lands owned by Ahtna, Inc. Another consideration was that the stand should have good access in the form of an existing road or trail. Third, the selected stands needed to have productive soils that were also well suited to management activities. The next criterion was that the site had a high potential for willows for moose treatments or white spruce if considering a timber stand improvement. This was determined based on the existing vegetation type and biophysical setting which describes the potential vegetation of a site. Another factor in site selection was to avoid areas that provide high quality caribou habitat as both moose browse treatments and timber stand improvements could prove detrimental to caribou forage (primarily lichen). An additional consideration was whether the site had high potential for berry production, with efforts made to avoid disturbance to high quality sites. The final consideration was to keep the maximum stand size to around 40 acres to avoid creating large open areas. This was primarily a consideration for the moose browse improvement areas, but was also applied to the timber treatments to avoid negative impacts to existing moose habitat.

The goals for the site improvement areas are to increase the amount of moose forage, provide increased opportunities for the harvest of moose in the fall, improve the condition of timbered stands, increase the ease of access to stands, and provide biomass and firewood to neighboring communities. Treatment areas are also proposed that would add to the existing Primary Line of Defense (PLOD). The PLOD is intended to provide predetermined boundaries around areas of high values at risk such as residential, recreational, or commercial structures. PLOD boundaries are selected to maximize tactical efficacy, accessibility, ease of identification from the ground or air, and potential fire fighter safety. Within the PLOD it is also important to create defensible space around structures. By reducing fuels around structures the rate and intensity of advancing wildfire is reduced. Defensible space also provides more room for firefighters to work and protects surrounding forest land from a structure fire. The defensible space is usually defined as an area a minimum of 30 feet around a structure that has been cleared of all flammable brush or vegetation. For additional information visit www.firewise.org. Based on all of the stated criteria, sites were located for designated treatments. Detailed information on specific treatments for each village can be found in the appropriate village planning section. Treatments for moose habitat improvement could use a number of possible mechanical treatments designed to stimulate growth of preferred moose browse species such as willow, or potentially prescribed fire. Treatments for fuel mitigation would use similar methods, but are designed to reduce the amounts of flammable material in the primary lines of defense and provide a location where defensive actions can be taken to counter an approaching fire. Specific treatments need to be determined for each selected site. Treatment options are described in the Management Treatments section.

Cantwell Planning Region

The Cantwell Planning Region encompasses an area of 480,794 acres. Figure 48 displays the overall planning area along with associated infrastructure. Ownership within the planning region and the surrounding area is shown in Figure 49. As Figure 49 displays, Ahtna, Inc. owns 25.6% (123,231 acres) of the land in this area.

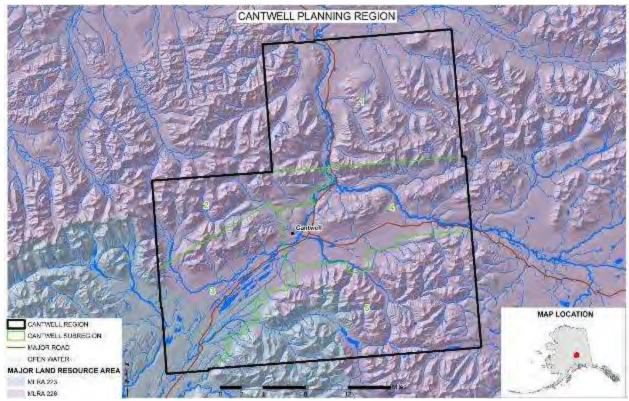


Figure 48. Overview of the Cantwell Planning Region of the Ahtna Traditional Use Territory.

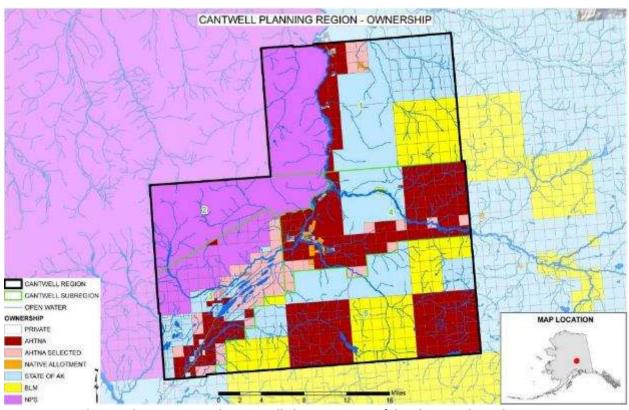


Figure 49. Land ownership patterns in the Cantwell Planning Region of the Ahtna Traditional Use Territory.

Geology

The geology of the Cantwell area was described in the <u>Denali National Park Area Soil Survey</u>. Rocks in the area consist of shale, andesite, schist, diorite, conglomerate, and limestone. The region around Cantwell is known as the Alaska Mountains Section and consists of steep alpine talus slopes and more gentle slopes locally flanked by glacial deposits. Surrounding the steep mountain slopes are rounded low mountains, plateaus, glacial plains, and hills. Running throughout these raised features are braided glacial-fed rivers and clear water streams with adjacent flood plains, alluvial fans, and terraces. There are soil materials of three main types. These are gravelly colluvium, drift, and loamy and gravelly alluvium. A major fault line, the Denali fault system, bisects the area running from east to west.

Climate

The location of the Cantwell area causes it to fall on the border between two major climatic areas of Alaska. Areas to the north have a sub-arctic continental climate, while areas to the south are transitional maritime- continental. The climate for these two types in the area was describe in the Denali National Park Area Soil Survey. "Based on climatic summaries from the Western Regional Climate Center, the sub-arctic continental climate characteristic to Interior Alaska consists of long cold winters and short warm summers. Mean minimum January air temperature at Minchumina along the northwest border of Denali Park are -12.6 °F (-24.8 °C) and the mean maximum July temperature is 71.6 °F (22 °C). Total precipitation is relatively low, totaling 12.8 inches (32.5 cm). The dry characteristic of the interior climatic zone is best understood by comparing total annual precipitation to water loss (evapotranspiration). Annual precipitation and potential evapotranspiration estimates for Minchumina are reported as 12.8 inches (32.5 cm) and 17.9 inches (45.4 cm) (Patric and Black 1968) and represent an annual moisture deficit of 5.1 inches (12.9 cm). Winter snow cover, an important insulator against subzero winter air temperatures, is relatively low, averaging only 18.1 inches (46 cm) during March.

The transitional maritime-continental climate of South Central which includes the South Central Mountains and Cook Inlet Lowlands Sections is characterized as a blend of the mild, moist maritime influences of the coastal zone of the Gulf of Alaska and the cold, dry continental influences of Interior Alaska. Mean minimum January air temperature at Talkeetna, outside the southern border of Denali Park, is 1.8 °F (-16.8 °C) and the mean maximum July temperature is 67.8 °F (19.9 °C) and permafrost is generally absent. Average precipitation at Talkeetna is over double that of the Minchumina station, at 27.9 inches (70.9 cm), owing to the more significant marine influence. Annual precipitation and potential evapotranspiration estimates for Talkeetna, the nearest available recording station is 27.9 inches (70.9 cm) and 18.7 inches (47.5 cm) (Patric and Black 1968) and represents a surplus of 9.2 inches (23.4 cm)."

Weather measurements from Cantwell indicate the area's location between these two climatic types. Minimum January air temperature is -8.7 °F (-22.6 °C) and maximum July temperature is 66.2 °F (19.0 °C). The mean annual precipitation is 17.0 inches (43.2 cm) and the March snow cover averages 30 inches (76.2 cm).

Soils

Soil texture in the Cantwell area is shown in Figure 50 and Figure 51 displays soil drainages.

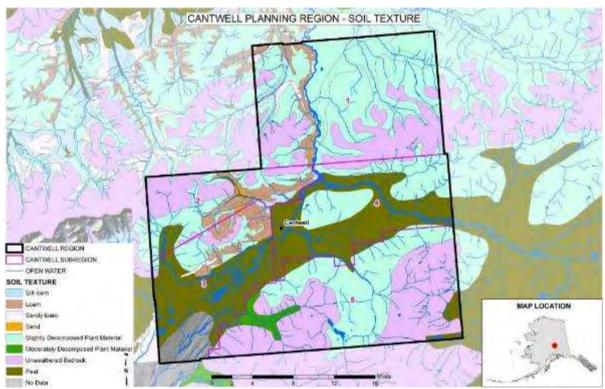


Figure 50. Soil texture in the Cantwell Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK651 and NRCS STATSGO database for Alaska.

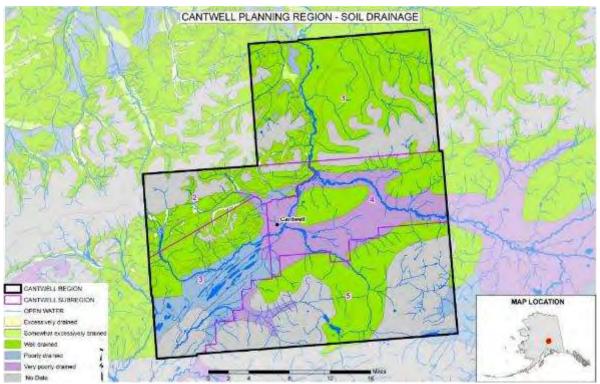


Figure 51. Soil drainage in the Cantwell Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK651 and NRCS STATSGO database for Alaska.

Permafrost

A significant factor influencing the vegetation in the landscape is the occurrence of permafrost under some of the project area. Permafrost is extensive on loamy textured soils within the boreal biome and is only occasionally observed in gravelly alpine soils. The Copper River Basin Soil Survey described the role of permafrost as: "Permafrost, or perennially frozen ground, underlies most of the Copper River basin. The depth at which it occurs and its ice content varies widely. Permafrost characteristically occurs as ice crystals disseminated throughout the soil. Although not extensive near the soil surface, massive ice wedges and lenses do occur in the subsoil in some areas. A perched water table and saturated conditions are common above the permafrost during the summer due to restricted drainage. The fire history of the site and the thickness of the insulating organic layer on the soil surface control depth to permafrost and water table, in part. Disturbance of the organic layer usually results in increased soil temperatures and a lowering of the permafrost level. As permafrost thaws, a large volume of water is released. Variation in the ice content of the permafrost and the rate of thawing results in differential subsidence of the soil surface and slumping on steeper slopes. The occurrence of permafrost requires special consideration when selecting lands for clearing and agriculture and during construction of roads and buildings." Permafrost considerations should be evaluated in other management decisions including selection of areas for moose habitat improvements. Figure 52 displays the occurrence of permafrost in the Cantwell project area as interpreted from the soil survey information.

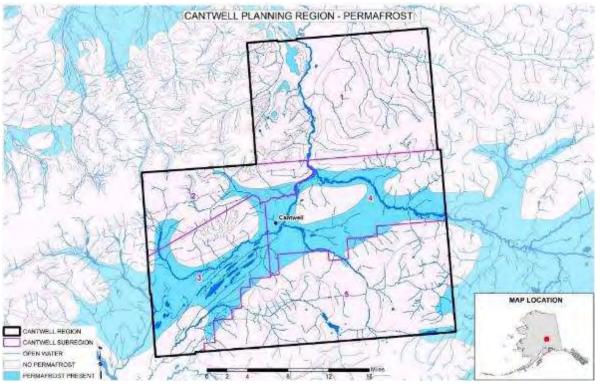


Figure 52. Permafrost in the Cantwell Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK651 and NRCS STATSGO database for Alaska.

Vegetation Description

The <u>Denali National Park Area Soil Survey</u> provided a general description of the vegetation occurring in the Cantwell area. It stated: "General patterns of vegetation in the study area are the result of two major influences: the elevation gradient of the Alaska Range, and the different climactic regions north

and south of the range. Much of the Denali Park is above tree line, and almost one-sixth is non-vegetated ice and rocky mountain slopes. In the vegetated zone, harsh conditions at high elevations limit plant communities to dwarf shrubs and herbaceous meadows in nivation hollows. Medium or tall shrubs are found lower down the slopes and these grade into forests or woodlands on well-drained substrates at lower elevations. Poor drainage at all elevations, because of glacial drift or permafrost, limits productivity. In lowlands, wet woodlands, shrubs, and herbaceous communities are found in a mosaic of fens, bogs, marshes and muskegs. Mountain vegetation of the Alaska Mountains Section is dominated by white mountain avens (*Dryas octopetala*) - dwarf ericaceous shrubs, which grade into medium-sized shrubs dominated by shrub birch and ericaceous shrubs such as blueberry (*Vaccinium uliginosum*), Labrador tea (*Ledum palustre* ssp. *decumbens* and *L. groenlandicum*) and crowberry (*Empetrum hermaphroditum*). On cooler, more northerly aspects these shrubs sometimes have high percentages of sedge and other herbaceous vegetation. Warmer low slopes, especially in the Kantishna Hills and Park headquarters areas, support white spruce/mixed scrub woodlands."

Disturbance Factors

As mentioned in reference to permafrost and vegetation, fire is a disturbance factor influencing the vegetation ecology in the project area. Figure 53 shows the fire history of the Cantwell area along with current fire protection zones. Although the level of fire occurring in this southcentral Alaska landscape is substantially less than that occurring in more interior areas of Alaska north of the Alaska Range, fire is still a significant disturbance when it occurs. However, there have been no recorded fires within the planning area since 1940. When fire does occur it serves to set back succession. It can also burn off the organic material at the ground surface, including peat that can occur on many sites. This can influence the thermal layer protecting the underlying permafrost on some sites, causing the permafrost to melt (thermokarst) and changing the site conditions through this process. Riparian areas are also influenced by flooding and ice events. These serve to set back succession of vegetation in riparian areas, and can even shift site conditions, particularly in the case of significant flooding events.

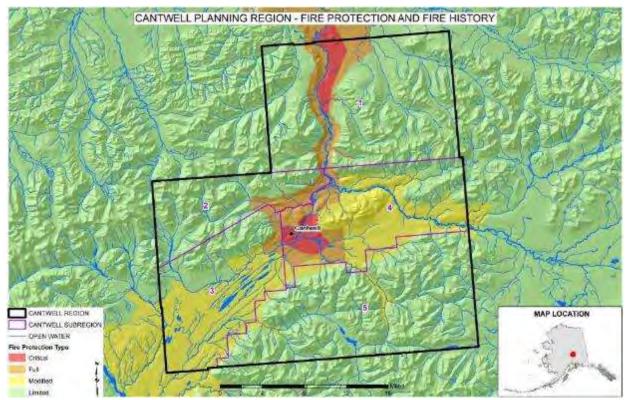


Figure 53. Current fire protection classes and fire history since 1940 in the Cantwell Planning Region of the Ahtna Traditional Use Territory. Data from Alaska Interagency Coordination Center.

Ecosystem Diversity

The biophysical settings present within the Cantwell Planning region are displayed in Figure 54. Table 19 displays the acres for each biophysical setting and disturbance class. Figure 55 is a map of ecosystem diversity (represented by biophysical setting and disturbance class) in the Cantwell Planning Region.

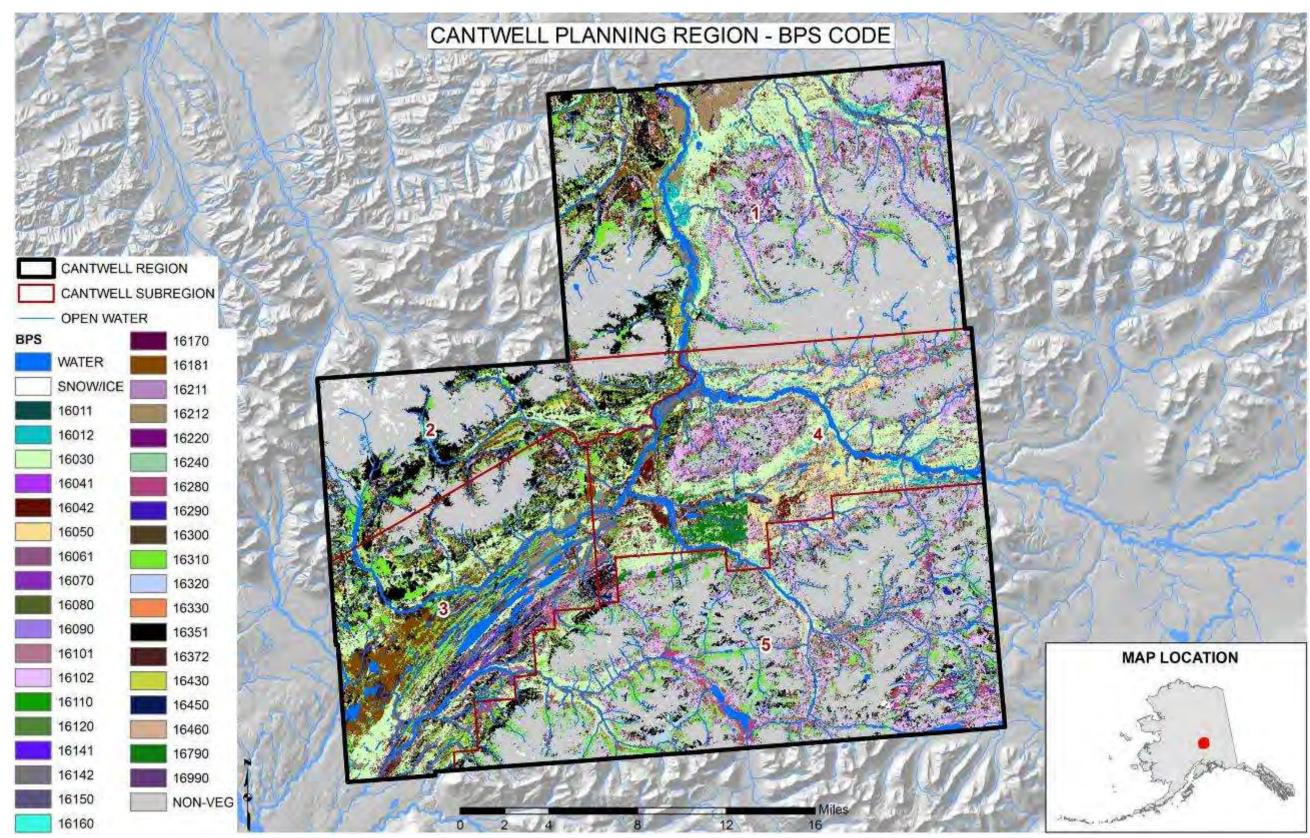


Figure 54. Biophysical settings (codes) in the Cantwell Planning Region of the Ahtna Traditional Use Territory. Data from LANDFIRE.

Table 19. Acres by BpS code and disturbance class (A – E) in the Cantwell Planning Region. The BpS vegetation label is provided as well.

	vided as well.		DwC C1-		
BpS Code _Dist Class	BpS Vegetation Label	Acres	BpS Code _Dist Class	BpS Vegetation Label	Acres
11	Open Water	7691.5	16142_D	Montane Floodplain-Subboreal	52.7
12	Perrennial Ice-Snow	3830.3	16142_E	Montane Floodplain-Subboreal	9.6
16011_A	Treeline White Spruce-Boreal	149.7	16150_A	Large River Floodplain	125.2
16011_B	Treeline White Spruce-Boreal	152.8	16150_B	Large River Floodplain	24.9
16012_A	Treeline White Spruce-SubBoreal	2380.7	16150_C	Large River Floodplain	23.1
16012_B	Treeline White Spruce-SubBoreal	3056.8	16150_D	Large River Floodplain	1.1
16012_C	Treeline White Spruce-SubBoreal	6.2	16150_E	Large River Floodplain	1.3
16030_A	White Spruce-Hardwood-Boreal	2816.0	16160_A	Riparian Stringer	7.1
16030_B	White Spruce-Hardwood-Boreal	40844.2	16160_B	Riparian Stringer	76.5
16030_C	White Spruce-Hardwood-Boreal	20555.8	16160_C	Riparian Stringer	0.9
16030_E	White Spruce-Hardwood-Boreal	42.0	16170_A	Shrub and Herbaceous Floodplain	19.3
16041_A	Mesic Black Spruce-Boreal	43.1	16170_B	Shrub and Herbaceous Floodplain	2.2
16041_B	Mesic Black Spruce-Boreal	218.4	16170_D	Shrub and Herbaceous Floodplain	6.0
16041_C	Mesic Black Spruce-Boreal	47.6	16170_E	Shrub and Herbaceous Floodplain	6.2
16041_D	Mesic Black Spruce-Boreal	102.7	16181_A	Herbaceous Fen	4145.9
16041_E	Mesic Black Spruce-Boreal	0.4	16181_B	Herbaceous Fen	5291.7
16042_A	Mesic Black Spruce-SubBoreal	4385.6	16181_C	Herbaceous Fen	0.9
16042_B	Mesic Black Spruce-SubBoreal	3574.6	16181_D	Herbaceous Fen	8453.7
16042_C	Mesic Black Spruce-SubBoreal	1511.2	16211_A	Dwarf Black Spruce Peatland-Boreal	436.6
16042_D	Mesic Black Spruce-SubBoreal	6.4	16211_B	Dwarf Black Spruce Peatland-Boreal	968.8
16042_E	Mesic Black Spruce-SubBoreal	1.3	16211_C	Dwarf Black Spruce Peatland-Boreal	1528.3
16050_A	Mesic Birch-Aspen	1284.8	16211_D	Dwarf Black Spruce Peatland-Boreal	627.4
16050_B	Mesic Birch-Aspen	3483.8	16212_A	Dwarf Black Spruce Peatland-Subboreal	1200.9
16050_E	Mesic Birch-Aspen	0.7	16212_B	Dwarf Black Spruce Peatland-Subboreal	8162.3
16061_B	Dry Aspen-Steppe Bluff	0.9	16212_C	Dwarf Black Spruce Peatland-Subboreal	4074.2
16070_A	Subalpine Balsam Poplar-Aspen	4.7	16220_A	Black Spruce Wet-Mesic Slope	233.7
16070_B	Subalpine Balsam Poplar-Aspen	0.7	16220_B	Black Spruce Wet-Mesic Slope	753.7
16080_A	Avalanche Slope Shrubland	170.6	16220_C	Black Spruce Wet-Mesic Slope	92.7
16080_B	Avalanche Slope Shrubland	610.3	16220_D	Black Spruce Wet-Mesic Slope	154.3
16090_A	Mesic Subalpine Alder	1543.0	16240_A	Deciduous Shrub Swamp	2845.5
16090_B	Mesic Subalpine Alder	5924.2	16240_B	Deciduous Shrub Swamp	2.0
16102_A	Mesic Scrub Birch-Willow	4254.0	16280_A	Low Shrub-Tussock Tundra	3927.7
16102_B	Mesic Scrub Birch-Willow	40223.9	16280_B	Low Shrub-Tussock Tundra	8763.5
16110_A	Mesic Bluejoint Meadow	1534.3	16280_C	Low Shrub-Tussock Tundra	3496.0
16110_B	Mesic Bluejoint Meadow	0.9	16290_A	Tussock Tundra	195.0
16120_A	Dry Grassland	2732.8	16290_B	Tussock Tundra	1220.
16141_A	Montane Floodplain-Boreal	1180.5	16300_A	Wet Black Spruce-Tussock	129.9
16141_B	Montane Floodplain-Boreal	570.9	16300_B	Wet Black Spruce-Tussock	654.1
16141_C	Montane Floodplain-Boreal	180.1	16300_C	Wet Black Spruce-Tussock	43.4
16141_D	Montane Floodplain-Boreal	2.7	16310_A	Alpine Dwarf Shrub Summit	37204.
16141_E	Montane Floodplain-Boreal	0.4	16310_B	Alpine Dwarf Shrub Summit	1.3
16142_A	Montane Floodplain-Subboreal	8255.7	16320_A	Alpine Talus and Bedrock	24014.
16142_B	Montane Floodplain-Subboreal	3542.1	16320_B	Alpine Talus and Bedrock	3858.6
16142_C	Montane Floodplain-Subboreal	2145.2	16330_A	Alpine Mesic Herbaceous Meadow	1472.

BpS Code _Dist Class	BpS Vegetation Label	Acres	BpS Code _Dist Class	BpS Vegetation Label	Acres
16330_B	Alpine Mesic Herbaceous Meadow	0.2	16450_B	Alpine Mesic Herbaceous Meadow	0.4
16351_A	Alpine Ericaceous Dwarf-Shrubland	49075.2	16790_A	White Spruce-Hardwood-SubBoreal	3897.3
16351_B	Alpine Ericaceous Dwarf-Shrubland	21.6	16790_B	White Spruce-Hardwood-SubBoreal	963.9
16372_A	Alpine Floodplain	797.5	16790_C	White Spruce-Hardwood-SubBoreal	149.7
16372_B	Alpine Floodplain	1802.3	16790_D	White Spruce-Hardwood-SubBoreal	1.6
16372_C	Alpine Floodplain	158.3	16790_E	White Spruce-Hardwood-SubBoreal	0.7
16430_A	Alpine Dwarf Shrubland	6277.1	31	Barren Rock-Sand-Clay	123829.5
16450_A	Alpine Mesic Herbaceous Meadow	560.0			

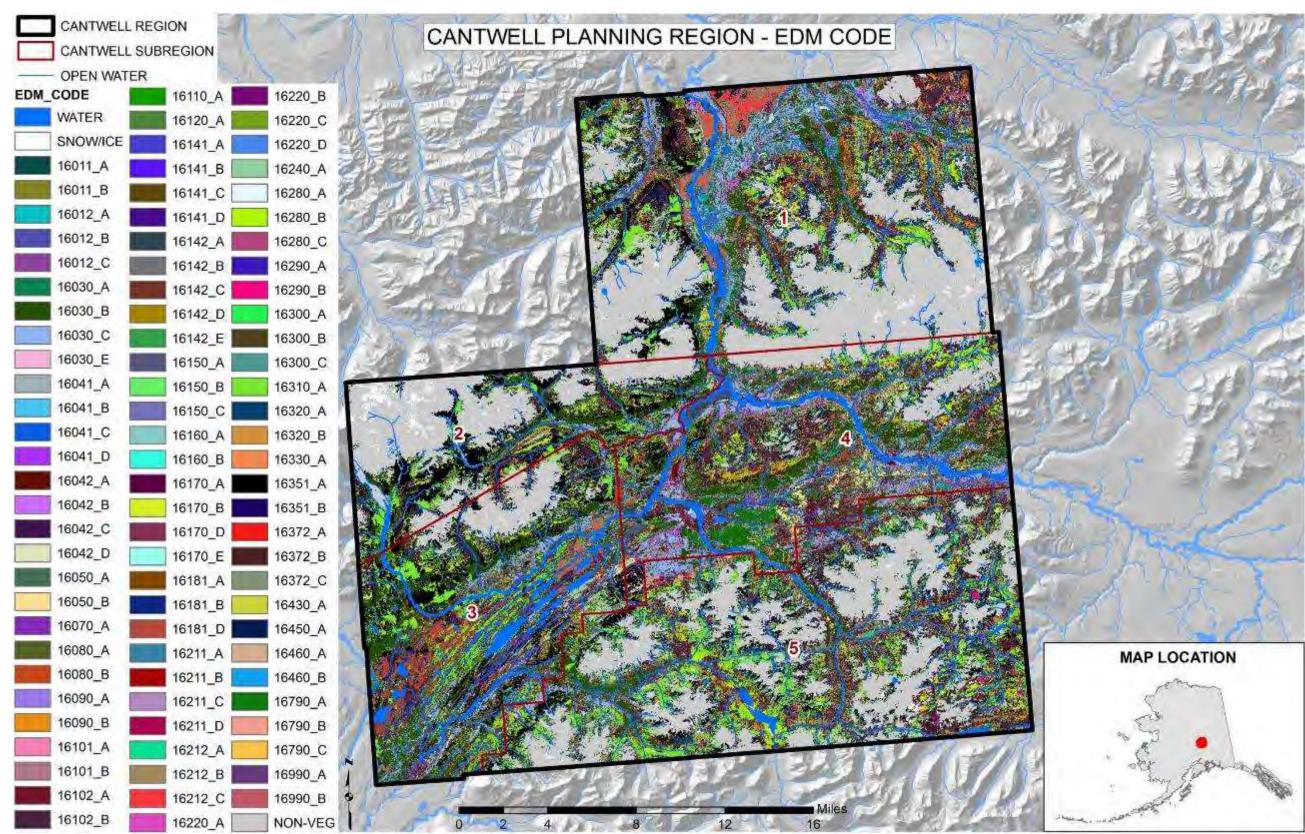


Figure 55. Map of ecosystem diversity in the Cantwell Planning Region of the Ahtna Traditional Use Territory. Numbers in the legend refer to the BpS code, while letters (A-E) refer to the disturbance class.

Berry Production Areas

Ecological sites that had the potential for producing desirable species of berries were identified. These sites were then further identified as to the disturbance states most likely to support berry production. While numerous factors can influence where berry-producing vegetation occurs, the combination of ecological sites and disturbance states allows for the identification of areas with higher probabilities of finding good berry production. Figure 56 shows the areas identified with the highest potential for berry production. It may be desirable to avoid disturbances to these areas, but ground-truthing of the actual presence of good berry production is recommended.

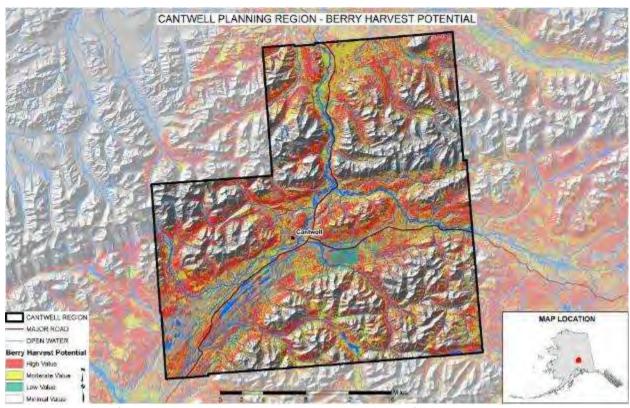


Figure 56. Potential for berry production in the Cantwell Planning Region of the Ahtna Traditional Use Territory based on ecological sites and disturbance states derived from remotely sensed information.

Moose and Caribou Model Results

The habitat suitability figure for moose spring habitat is found in Figure 57 and the landscape rating for moose spring habitat is found in Figure 58. The habitat suitability figure for moose summer habitat is found in Figure 59 and the landscape rating for moose summer habitat is found in Figure 60. The habitat suitability figure for moose winter habitat is found in Figure 61 and the landscape rating for moose winter habitat is found in Figure 62. The habitat suitability figure for caribou summer/calving habitat are found in Figure 63 and the landscape rating for caribou summer/calving habitat are found Figure 64. The habitat suitability figure for caribou winter habitat are found in Figure 65 and the landscape rating for caribou winter habitat are found Figure 66. A complete description of the moose habitat model can be found in Appendix A. A complete description of the caribou habitat model can be found in Appendix B.

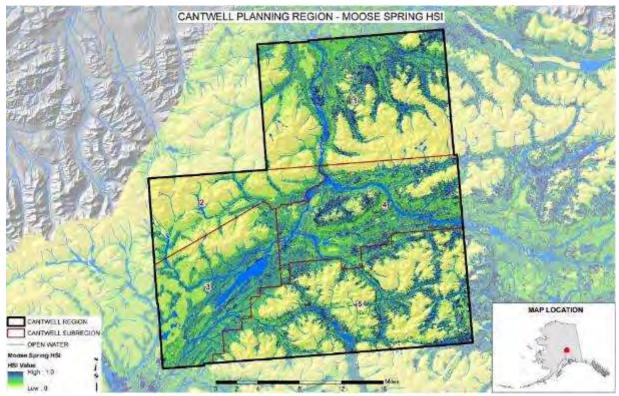


Figure 57. Habitat suitability for moose during spring in the Cantwell Planning Region of the Ahtna Traditional Use Territory.

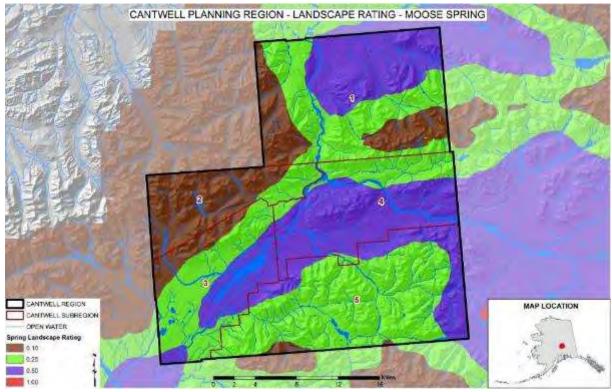


Figure 58. Landscape rating for moose during spring in the Cantwell Planning Region of the Ahtna Traditional Use Territory.

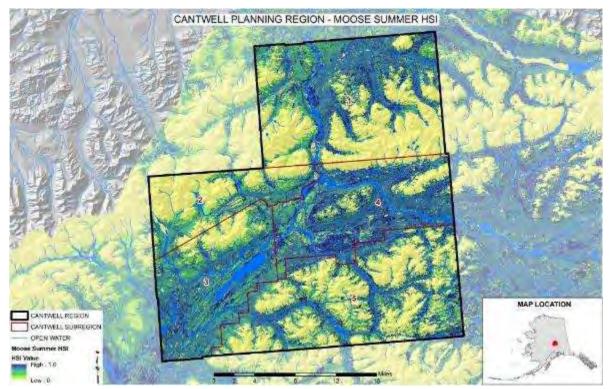


Figure 59. Habitat suitability for moose during summer in the Cantwell Planning Region of the Ahtna Traditional Use Territory.

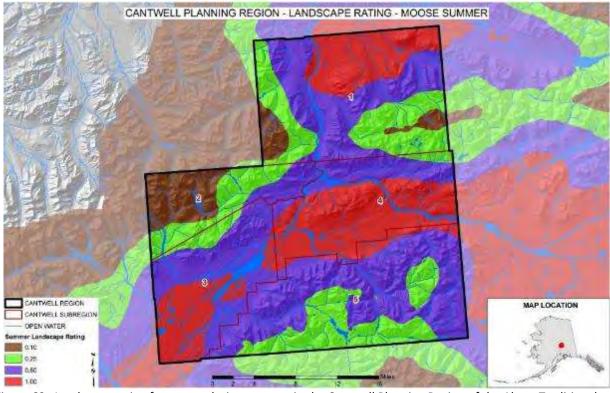


Figure 60. Landscape rating for moose during summer in the Cantwell Planning Region of the Ahtna Traditional Use Territory.

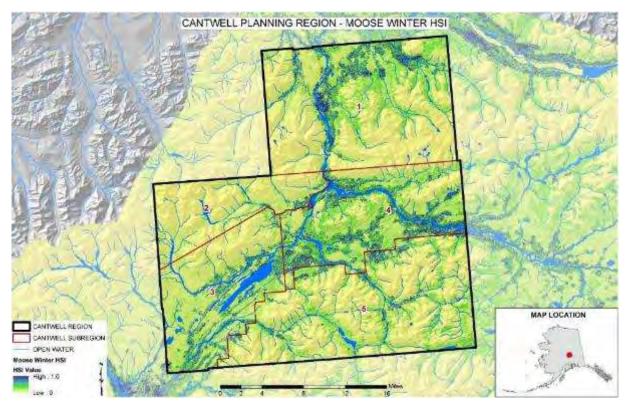


Figure 61. Habitat suitability for moose during winter in the Cantwell Planning Region of the Ahtna Traditional Use Territory.

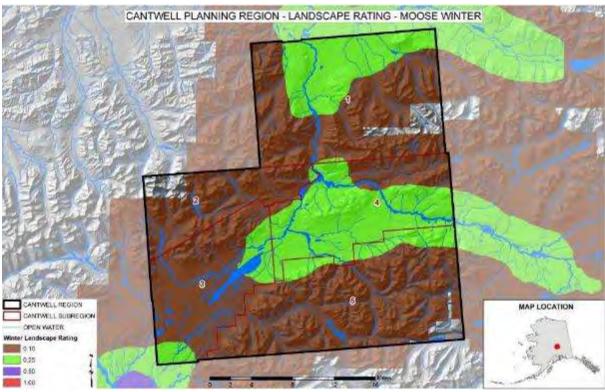


Figure 62. Landscape rating for moose during winter in the Cantwell Planning Region of the Ahtna Traditional Use Territory.

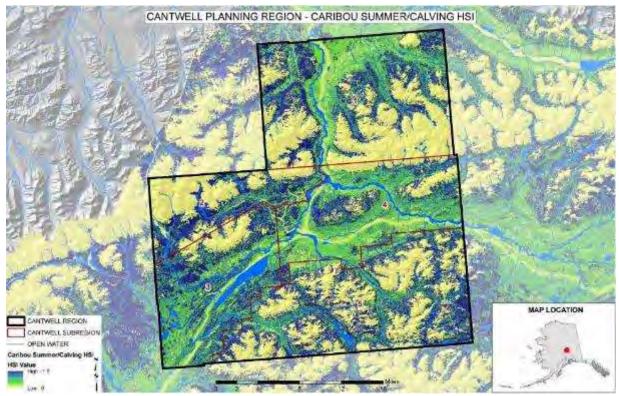


Figure 63. Habitat suitability for caribou during summer/calving in the Cantwell Planning Region of the Ahtna Traditional Use Territory.

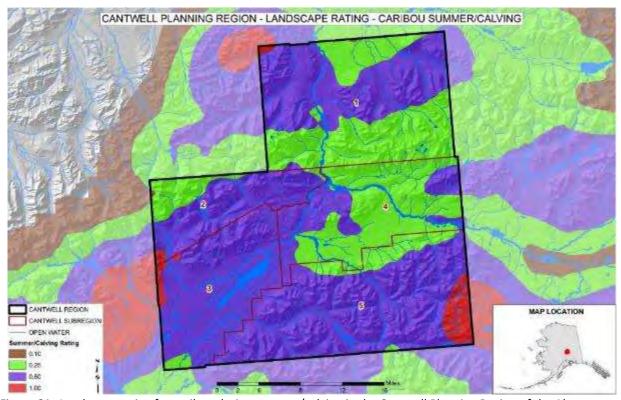


Figure 64. Landscape rating for caribou during summer/calving in the Cantwell Planning Region of the Ahtna Traditional Use Territory.

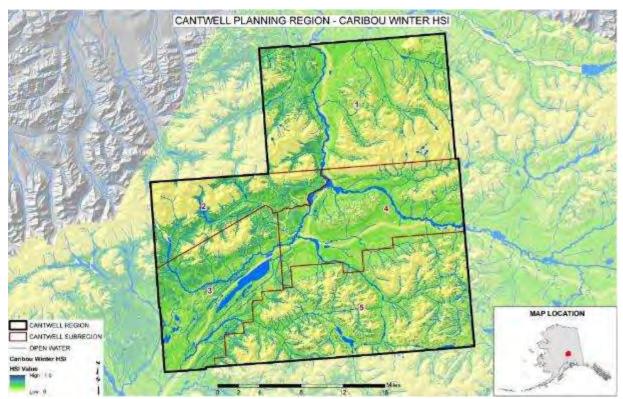


Figure 65. Habitat suitability for caribou during winter in the Cantwell Planning Region of the Ahtna Traditional Use Territory.

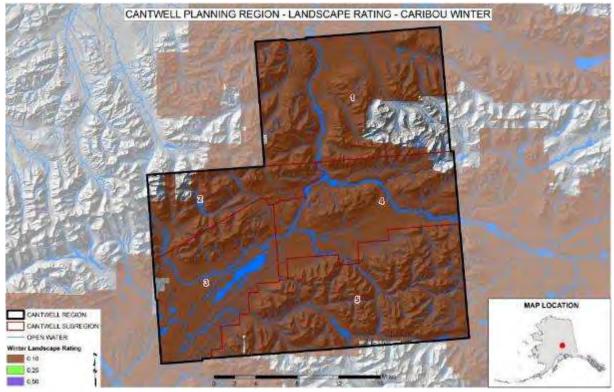


Figure 66. Landscape rating for caribou during winter in the Cantwell Planning Region of the Ahtna Traditional Use Territory.

Cantwell Site Improvement Areas

Potential treatment sites identified in the Cantwell area are displayed in figure 67. Based on all of the stated criteria, a number of sites were located for designated treatments. Figures 68-73 depict the locations of these areas. Appendix C provides a description of each of these sites. Sites ranging from approximately 17-62 acres in size were identified and are listed in Table 20. This table gives the site size and the treatment goal of primarily timber stand improvement and biomass production or moose browse enhancement. The table also displays an estimate of the total biomass of all standing above ground biomass. This estimate is not tonnage of usable biomass for energy but the total of all stems regardless of size. However, stands on the higher end of the total amount will typically have larger trees and are a candidate for biomass energy production.

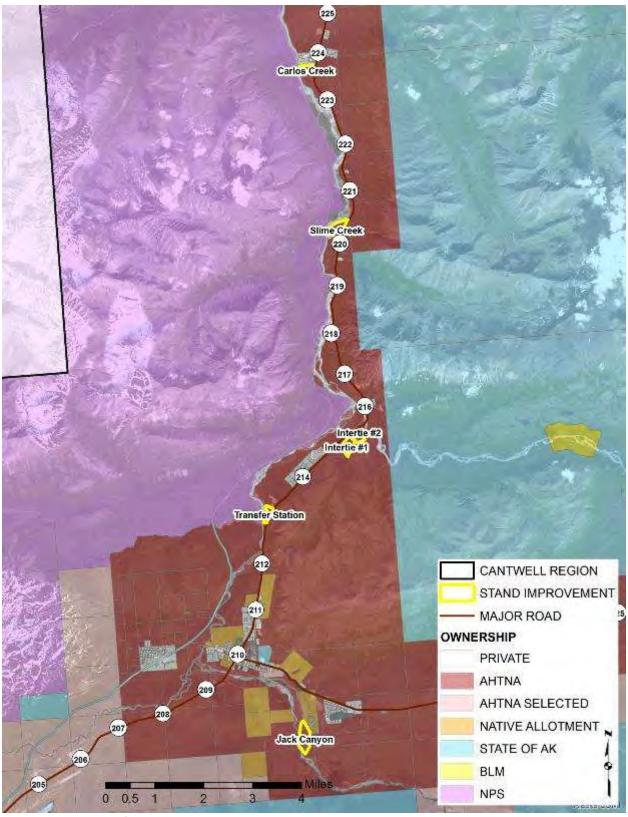


Figure 67. Overview of recommended treatment sites in the Cantwell management area.



Figure 68. Map of two proposed habitat improvement sites (Slime Creek and Carlos Creek) in the Cantwell Planning Region showing surface ownership and aerial imagery.

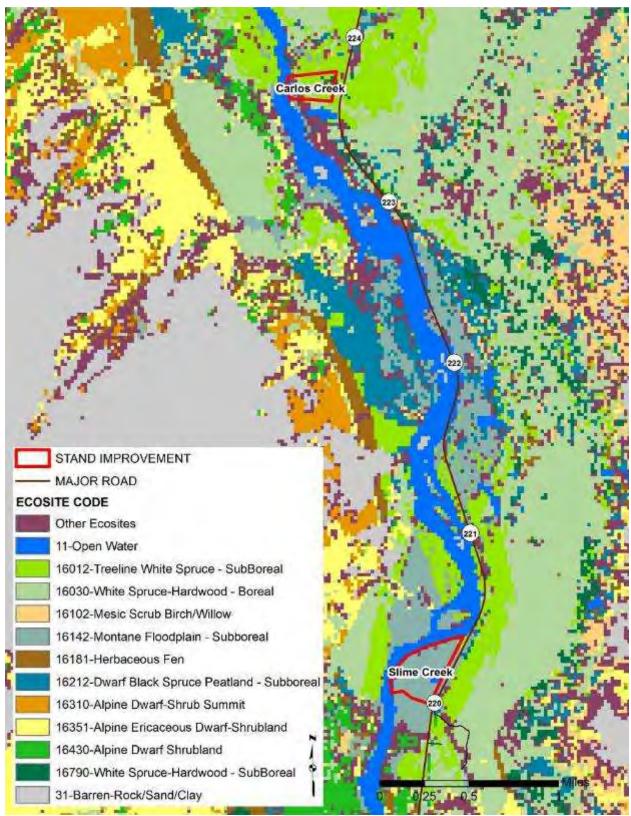


Figure 69. Map of two proposed treatment sites (Slime Creek and Carlos Creek) in the Cantwell Planning Region showing ecological sites.

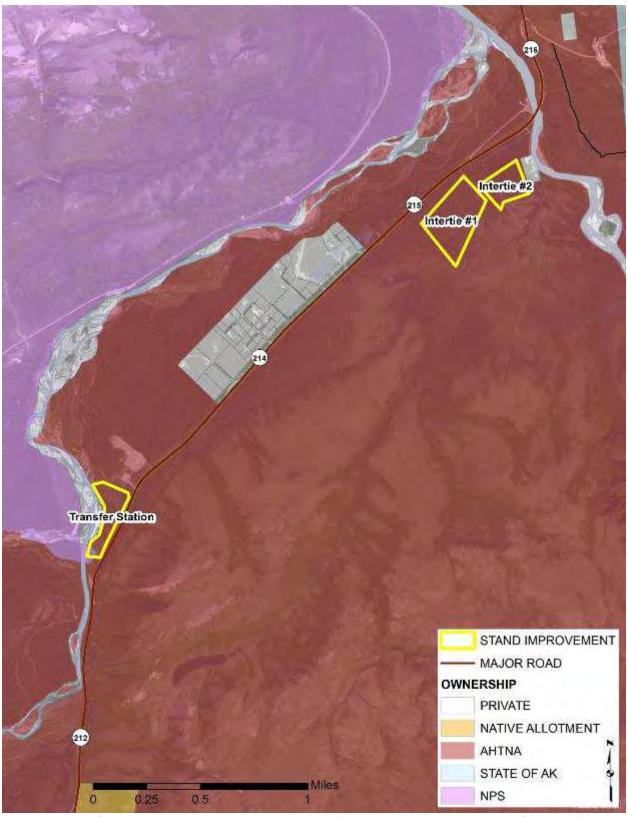


Figure 70. Map of three proposed habitat improvement sites (Intertie #1, Intertie #2, and Transfer Station) in the Cantwell Planning Region showing surface ownership and aerial imagery.

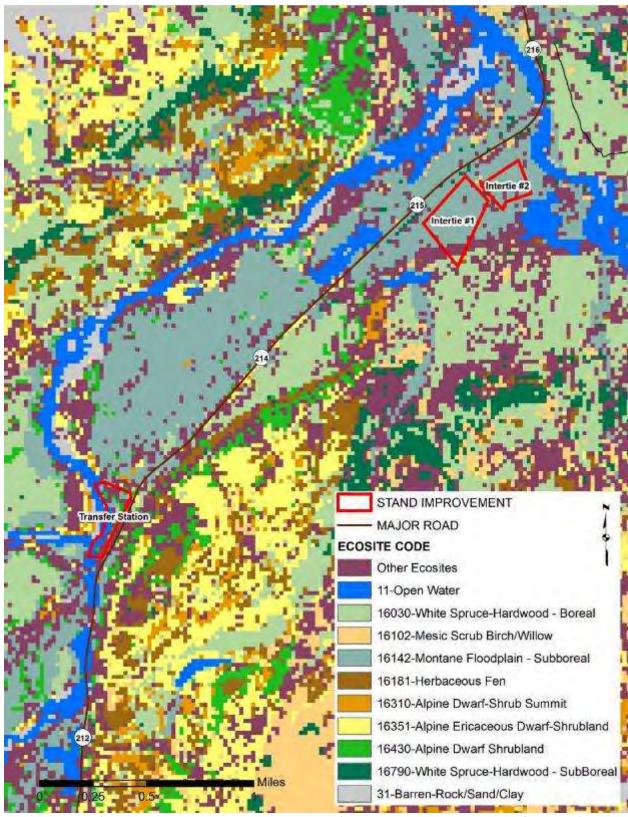


Figure 71. Map of three proposed habitat improvement sites (Intertie #1, Intertie #2, and Transfer Station) in the Cantwell Planning Region showing ecological sites.

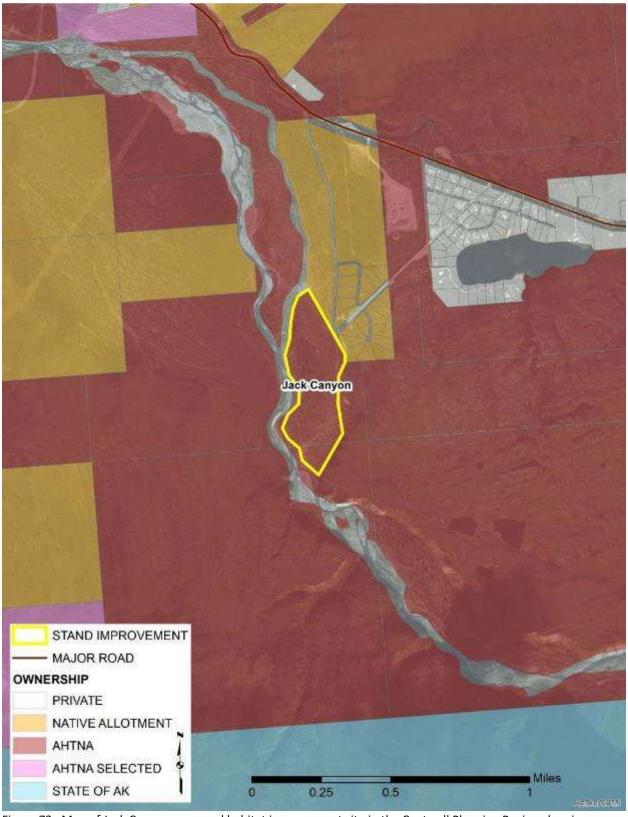


Figure 72. Map of Jack Canyon proposed habitat improvement site in the Cantwell Planning Region showing surface ownership and aerial imagery.

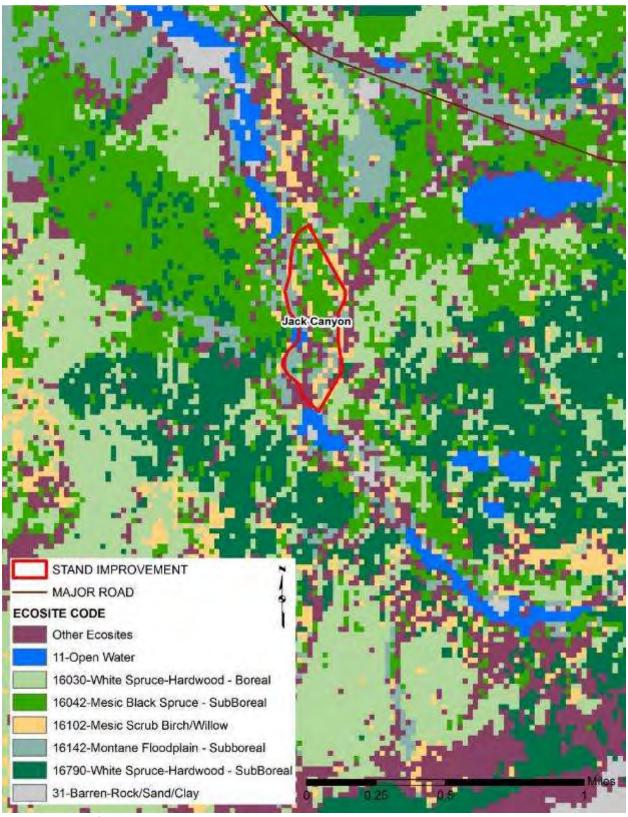


Figure 73. Map of Jack Canyon proposed habitat improvement site in the Cantwell Planning Region showing ecological sites.

Table 20. Treatment site with treatment goal, site size, total tons of biomass and the BpS code and disturbance class for each site in the Cantwell Planning Region.

Site Name	Treatment Goal	Acres	Biomass - Total BpS Code _Dist		
Site Name	Treatment Goal	Acres	(tons)	Code	
Carlos Creek	Moose Browse	23.8	394.5	16012_A	
Intertie #1	Moose Browse	39.6	663.5	16142_A	
Intertie #2	Moose Browse	17.0	471.3	16142_A	
Jack Canyon	Moose Browse	61.6	607.3	16042_A	
Slime Creek	Moose Browse	50.4	189.6	16142_A	
Transfer Station	Moose Browse	18.4	352.4	16042_B	

Chistochina Planning Region

The Chistochina Planning Region encompasses an area of 483,068 acres. Figure 74 displays the overall planning area along with associated infrastructure. Ownership of this and the surrounding area is shown in Figure 75. As Figure 75 displays, land ownership patterns are varied in this area with Ahtna, Inc. owning 37.4% (180,488 acres) of the land.

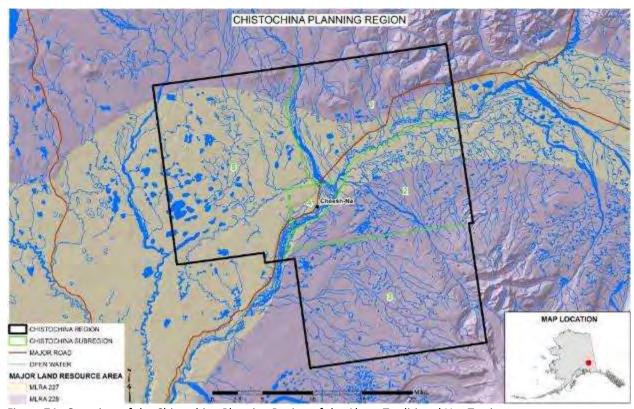


Figure 74. Overview of the Chistochina Planning Region of the Ahtna Traditional Use Territory.

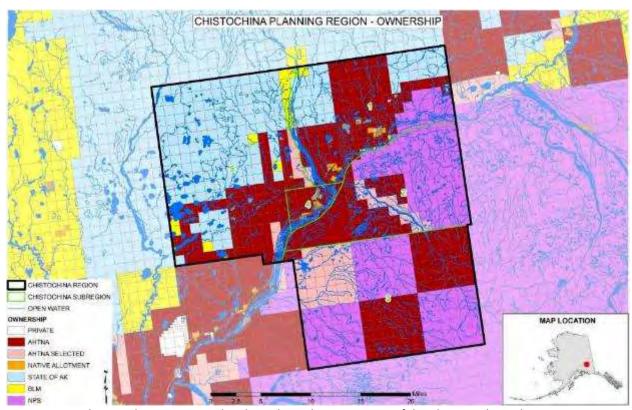


Figure 75. Land ownership patterns in the Chistochina Planning Region of the Ahtna Traditional Use Territory.

A description of the general geology, climate, soils, permafrost, and vegetation is found in Chapter 1 of this report. Figures showing these features specific to the Chistochina Planning Region are displayed below. Soil texture in the Chistochina area is shown in Figure 76 and Figure 77 displays soil drainages. Permafrost in the Chistochina area is shown in Figure 78.

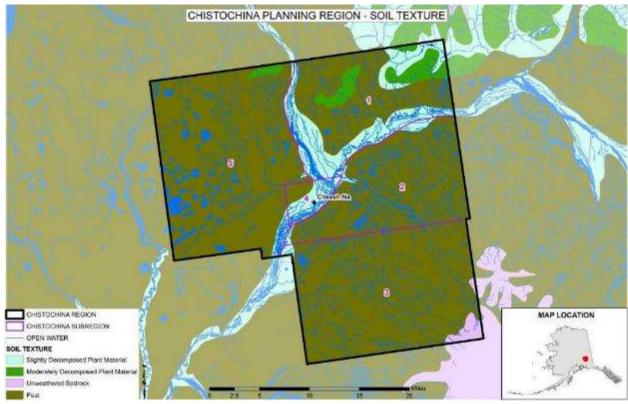


Figure 76. Soil texture in the Chistochina Planning Region of the Ahtna Traditional Use Territory. Data from NRCS STATSGO database for Alaska.

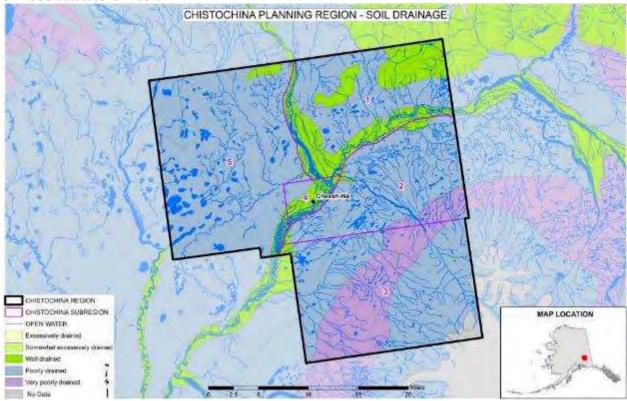


Figure 77. Soil drainage in the Chistochina Planning Region of the Ahtna Traditional Use Territory. Data from NRCS STATSGO database for Alaska.

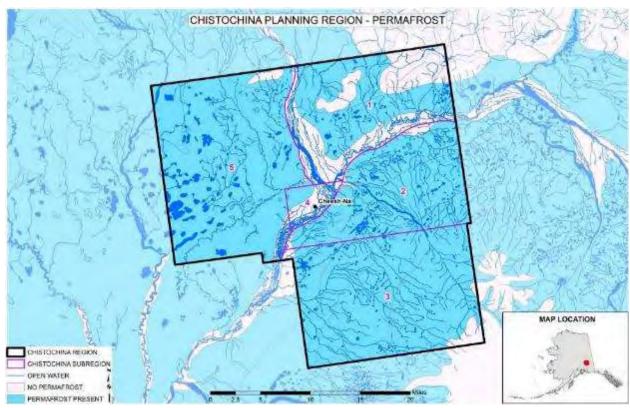


Figure 78. Permafrost in the Chistochina Planning Region of the Ahtna Traditional Use Territory. Data from NRCS STATSGO database for Alaska.

Disturbance Factors

Fire is the primary disturbance factor influencing the vegetation ecology in the Copper Basin. Figure 79 shows the fire history of the Chistochina area along with current fire protection zones. Although the level of fire occurring in this southcentral Alaska landscape is substantially less than that occurring in more interior areas of Alaska north of the Alaska Range, fire is still a significant disturbance when it occurs. When fire does occur it serves to set back succession. It can also burn off the organic material at the ground surface, including peat that can occur on many sites. This can influence the thermal layer protecting the underlying permafrost on some sites, causing the permafrost to melt (thermokarst) and changing the site conditions through this process. Riparian areas are also influenced by flooding and ice events. These serve to set back succession of vegetation in riparian areas, and can even shift site conditions, particularly in the case of significant flooding events.

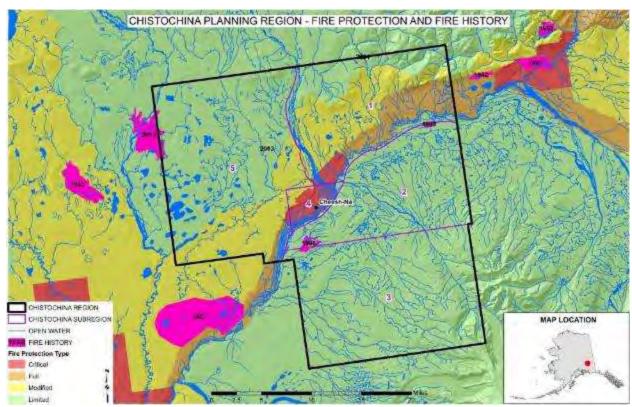


Figure 79. Current fire protection classes and fire history since 1940 in the Chistochina Planning Region of the Ahtna Traditional Use Territory. Data from Alaska Interagency Coordination Center.

Ecosystem Diversity

The biophysical settings present within the Chistochina Planning region are displayed in Figure 80. Table 21 displays the acres for each biophysical setting and disturbance class. Figure 81 is a map of ecosystem diversity (represented by biophysical setting and disturbance class) in the Chistochina Planning Region.

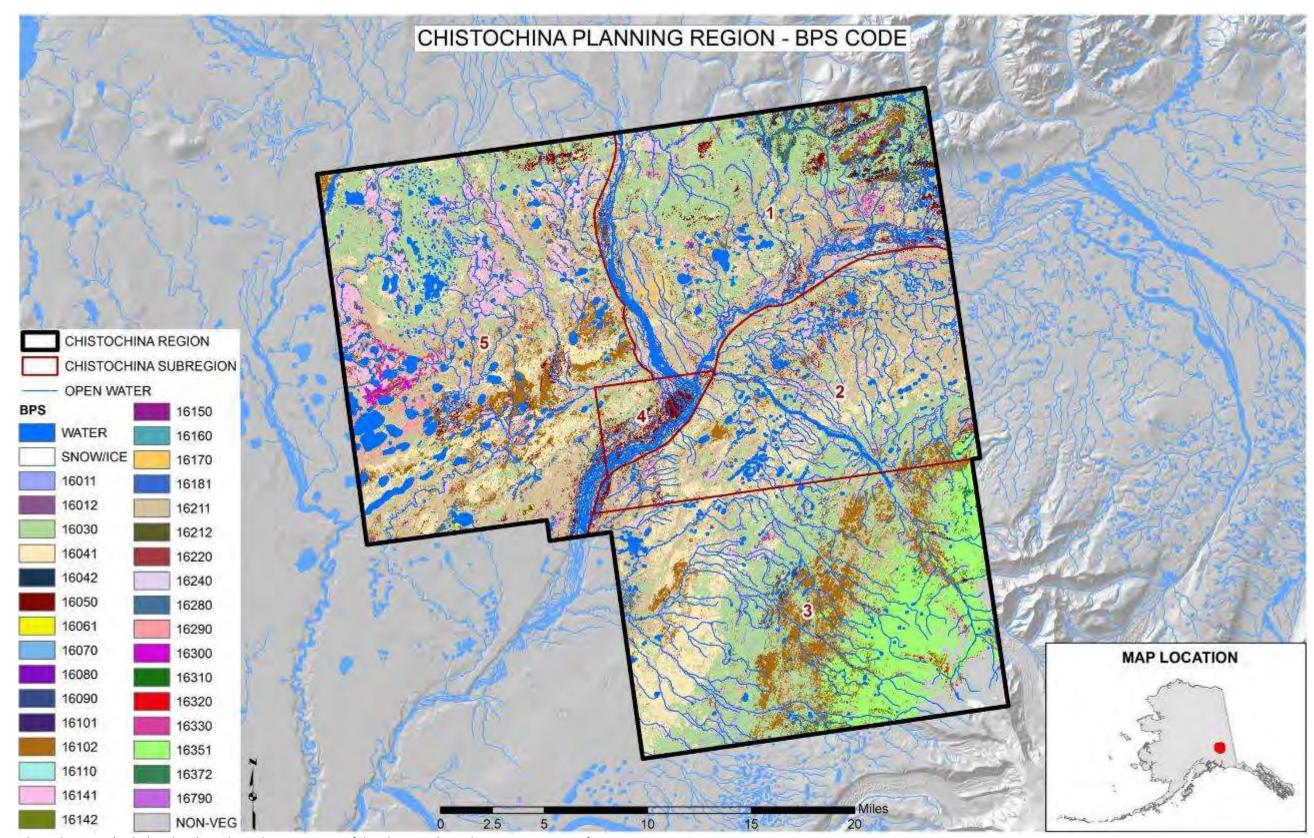


Figure 80. Biophysical settings (codes) in the Chistochina Planning Region of the Ahtna Traditional Use Territory. Data from LANDFIRE.

Table 21. Acres by BpS code and disturbance class (A - E) in the Chistochina Planning Region. The BpS vegetation label is provided as well.

BpS Code _Dist Class	BpS Vegetation Label	Acres	BpS Code _Dist Class	BpS Vegetation Label	Acres
11	Open Water	14805.5	16142_B	Montane Floodplain-Subboreal	285.8
12	Perrennial Ice-Snow	221.7	16142_C	Montane Floodplain-Subboreal	64.3
16011_A	Treeline White Spruce-Boreal	1067.9	16142 D	Montane Floodplain-Subboreal	554.7
16011_B	Treeline White Spruce-Boreal	1748.2	16142_E	Montane Floodplain-Subboreal	68.3
16012_A	Treeline White Spruce-SubBoreal	161.2	16150_A	Large River Floodplain	339.6
16012_B	Treeline White Spruce-SubBoreal	283.6	16150_B	Large River Floodplain	116.3
16012_C	Treeline White Spruce-SubBoreal	0.2	16150_C	Large River Floodplain	141.0
16030_A	White Spruce-Hardwood-Boreal	1720.4	16150_D	Large River Floodplain	1.1
16030_B	White Spruce-Hardwood-Boreal	27374.6	16150_E	Large River Floodplain	1.1
16030_C	White Spruce-Hardwood-Boreal	72610.4	16160_A	Riparian Stringer	455.7
16030_E	White Spruce-Hardwood-Boreal	65.8	16160_B	Riparian Stringer	198.6
16041_A	Mesic Black Spruce-Boreal	1858.6	16160_C	Riparian Stringer	10.7
16041_B	Mesic Black Spruce-Boreal	9368.2	16170_A	Shrub and Herbaceous Floodplain	2352.5
16041_C	Mesic Black Spruce-Boreal	4584.0	16170_B	Shrub and Herbaceous Floodplain	228.4
16041_D	Mesic Black Spruce-Boreal	49652.1	16170_C	Shrub and Herbaceous Floodplain	26.7
16041_E	Mesic Black Spruce-Boreal	46.3	16170_D	Shrub and Herbaceous Floodplain	54.7
16042_A	Mesic Black Spruce-SubBoreal	432.8	16170_E	Shrub and Herbaceous Floodplain	864.9
16042_B	Mesic Black Spruce-SubBoreal	884.0	16181_A	Herbaceous Fen	167.2
16042_C	Mesic Black Spruce-SubBoreal	384.1	16181_B	Herbaceous Fen	75.8
16042_D	Mesic Black Spruce-SubBoreal	0.7	16181_C	Herbaceous Fen	13.6
16050_A	Mesic Birch-Aspen	11740.2	16181_D	Herbaceous Fen	190.1
16050_B	Mesic Birch-Aspen	3770.9	16211_A	Dwarf Black Spruce Peatland-Boreal	3117.3
16050_D	Mesic Birch-Aspen	136.6	16211_B	Dwarf Black Spruce Peatland-Boreal	5480.0
16050_E	Mesic Birch-Aspen	95.2	16211_C	Dwarf Black Spruce Peatland-Boreal	16183.9
16061_A	Dry Aspen-Steppe Bluff	10.2	16211_D	Dwarf Black Spruce Peatland-Boreal	99363.4
16061_B	Dry Aspen-Steppe Bluff	98.5	16212_A	Dwarf Black Spruce Peatland-Subboreal	45.4
16061_C	Dry Aspen-Steppe Bluff	155.7	16212_B	Dwarf Black Spruce Peatland-Subboreal	189.3
16061_D	Dry Aspen-Steppe Bluff	171.5	16212_C	Dwarf Black Spruce Peatland-Subboreal	572.4
16070_A	Subalpine Balsam Poplar-Aspen	123.2	16220_A	Black Spruce Wet-Mesic Slope	318.5
16070_B	Subalpine Balsam Poplar-Aspen	298.0	16220_B	Black Spruce Wet-Mesic Slope	410.5
16080_A	Avalanche Slope Shrubland	84.1	16220_C	Black Spruce Wet-Mesic Slope	46.5
16080_B	Avalanche Slope Shrubland	203.3	16220_D	Black Spruce Wet-Mesic Slope	661.0
16090_A	Mesic Subalpine Alder	403.2	16240_A	Deciduous Shrub Swamp	292.2
16090_B	Mesic Subalpine Alder	1278.1	16240_B	Deciduous Shrub Swamp	55.8
16102_A	Mesic Scrub Birch-Willow	13847.4	16280_A	Low Shrub-Tussock Tundra	9859.7
16102_B	Mesic Scrub Birch-Willow	22257.1	16280_B	Low Shrub-Tussock Tundra	2296.9
16110_A	Mesic Bluejoint Meadow	184.1	16280_C	Low Shrub-Tussock Tundra	233.1
16110_B	Mesic Bluejoint Meadow	0.2	16290_A	Tussock Tundra	1386.6
16141_A	Montane Floodplain-Boreal	11640.6	16290_B	Tussock Tundra	569.8
16141_B	Montane Floodplain-Boreal	4896.7	16300_A	Wet Black Spruce-Tussock	220.2
16141_C	Montane Floodplain-Boreal	2009.6	16300_B	Wet Black Spruce-Tussock	428.3
16141_D	Montane Floodplain-Boreal	23976.9	16300_C	Wet Black Spruce-Tussock	2077.4
16141_E	Montane Floodplain-Boreal	2071.6	16310_A	Alpine Dwarf-Shrub Summit	656.3
16142_A	Montane Floodplain-Subboreal	13.3437	16310_B	Alpine Dwarf-Shrub Summit	11.6

BpS Code _Dist Class	BpS Vegetation Label	Acres	BpS Code _Dist Class	BpS Vegetation Label	Acres
16320_A	Alpine Talus and Bedrock	1209.8	16372_B	Alpine Floodplain	2.2
16320_B	Alpine Talus and Bedrock	53.2	16372_C	Alpine Floodplain	1.3
16330_A	Alpine Mesic Herbaceous Meadow	96.7	16430_A	Alpine Dwarf Shrubland	0.4
16330_B	Alpine Mesic Herbaceous Meadow	22.7	16790_A	White Spruce-Hardwood-SubBoreal	5.6
16351_A	Alpine Ericaceous Dwarf-Shrubland	29101.3	16790_B	White Spruce-Hardwood-SubBoreal	0.7
16351_B	Alpine Ericaceous Dwarf-Shrubland	896.3	31	Barren-Rock-Sand-Clay	14102.3
16372_A	Alpine Floodplain	9.34059			

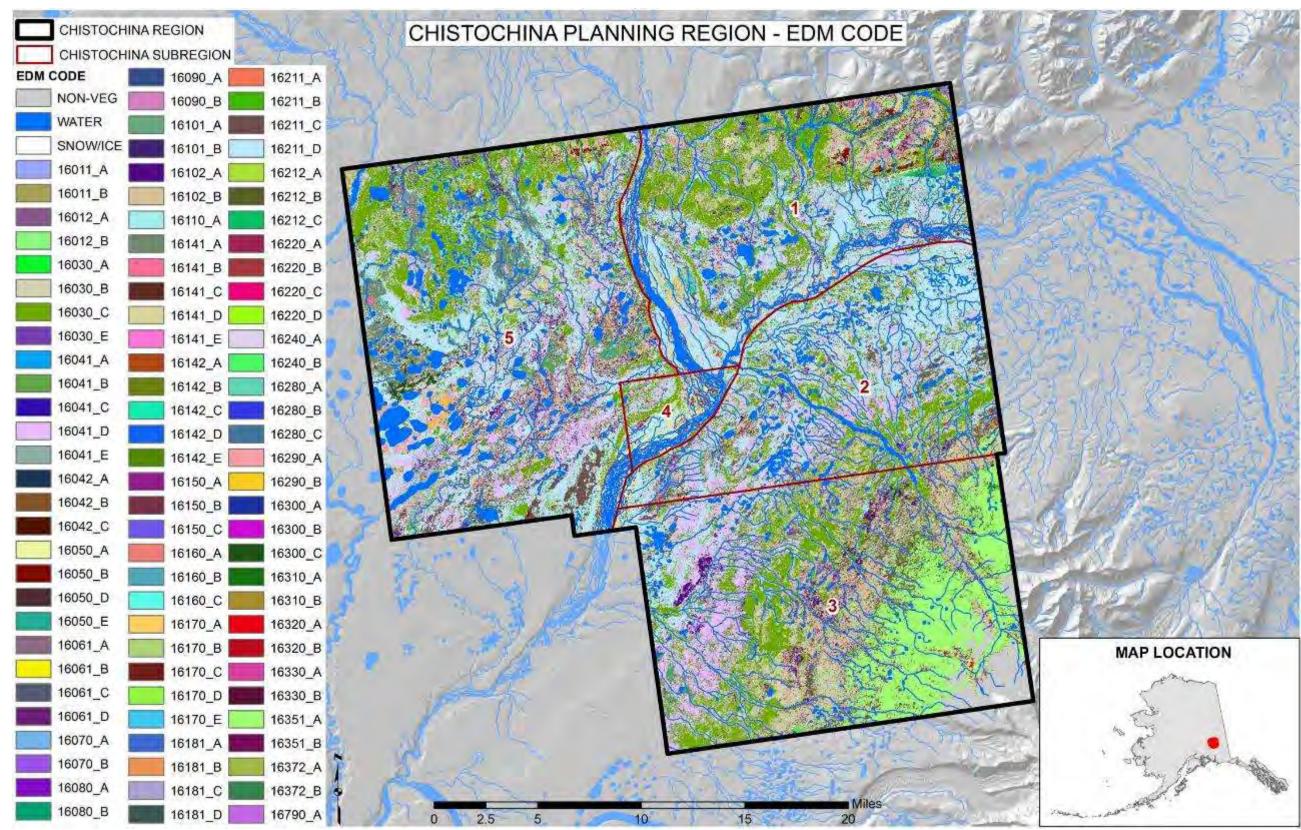


Figure 81. Map of ecosystem diversity in the Chistochina Planning Region of the Ahtna Traditional Use Territory. Numbers in the legend refer to the BpS code, while letters (A-E) refer to the disturbance class.

Berry Production Areas

Ecological sites that had the potential for producing desirable species of berries were identified. These sites were then further identified as to the disturbance states most likely to support berry production. While numerous factors can influence where berry-producing vegetation occurs, the combination of ecological sites and disturbance states allows for the identification of areas with higher probabilities of finding good berry production. Figure 82 shows the areas identified with the highest potential for berry production. It may be desirable to avoid disturbances to these areas, but ground-truthing of the actual presence of good berry production is recommended.

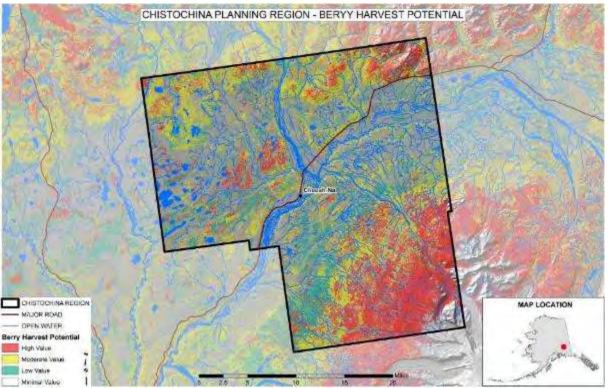


Figure 82. Potential for berry production in the Chistochina Planning Region of the Ahtna Traditional Use Territory based on ecological sites and disturbance states derived from remotely sensed information.

Moose and Caribou Model Results

The habitat suitability figure for moose spring habitat is found in Figure 83 and the landscape rating for moose spring habitat is found in Figure 84. The habitat suitability figure for moose summer habitat is found in Figure 85 and the landscape rating for moose summer habitat is found in Figure 86. The habitat suitability figure for moose winter habitat is found in Figure 87 and the landscape rating for moose winter habitat is found in Figure 88. The habitat suitability figure for caribou summer/calving habitat are found in Figure 89 and the landscape rating for caribou summer/calving habitat are found Figure 90. The habitat suitability figure for caribou winter habitat are found in Figure 91 and the landscape rating for caribou winter habitat are found Figure 92. A complete description of the moose habitat model can be found in Appendix A. A complete description of the caribou habitat model can be found in Appendix B.

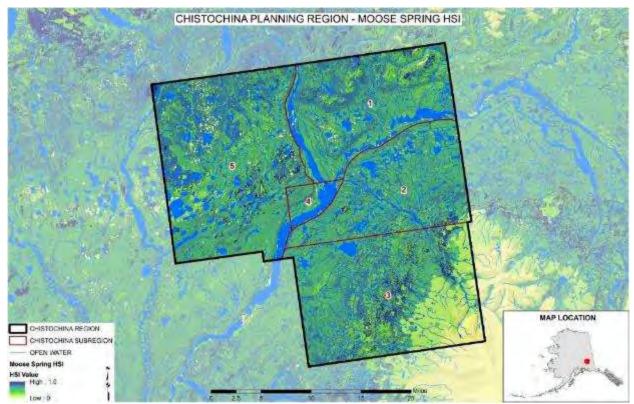


Figure 83. Habitat suitability for moose during spring in the Chistochina Planning Region of the Ahtna Traditional Use Territory.

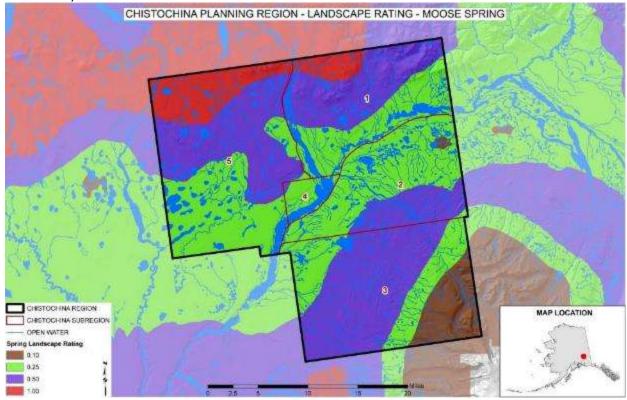


Figure 84. Landscape rating for moose during spring in the Chistochina Planning Region of the Ahtna Traditional Use Territory.

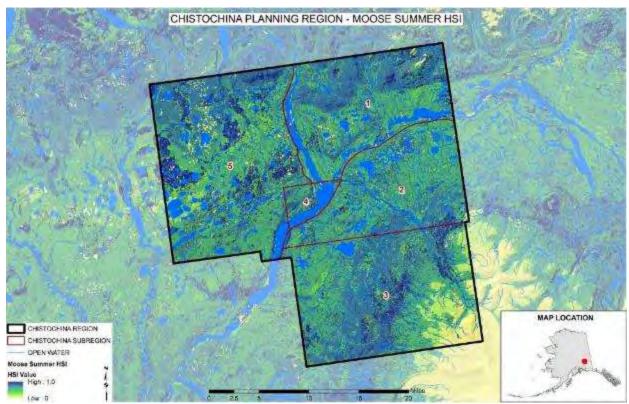


Figure 85. Habitat suitability for moose during summer in the Chistochina Planning Region of the Ahtna Traditional Use Territory.

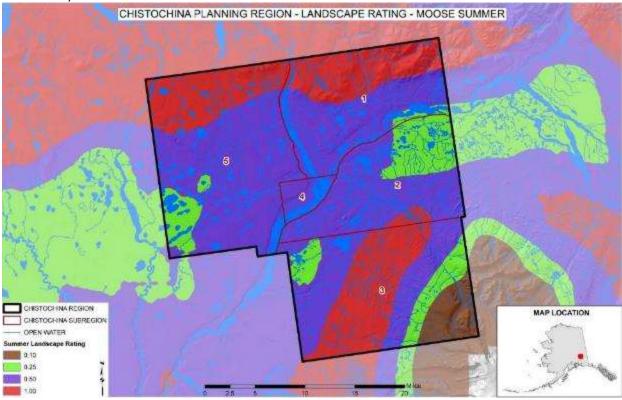


Figure 86. Landscape rating for moose during summer in the Chistochina Planning Region of the Ahtna Traditional Use Territory.

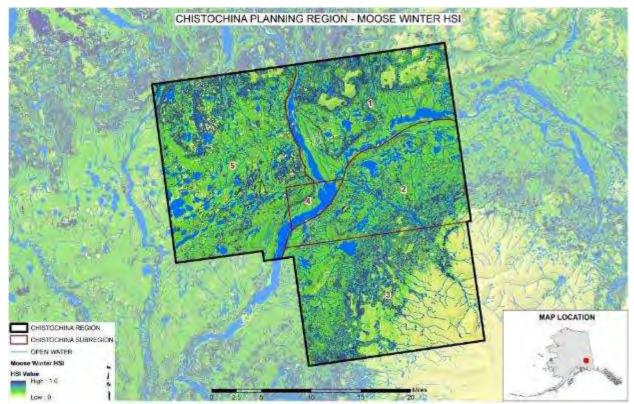


Figure 87. Habitat suitability for moose during winter in the Chistochina Planning Region of the Ahtna Traditional Use Territory.

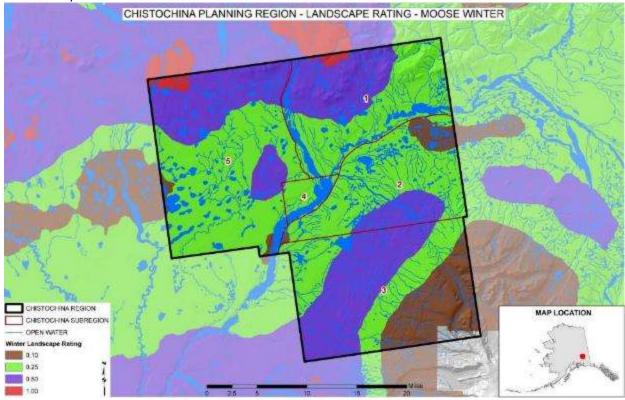


Figure 88. Landscape rating for moose during winter in the Chistochina Planning Region of the Ahtna Traditional Use Territory.

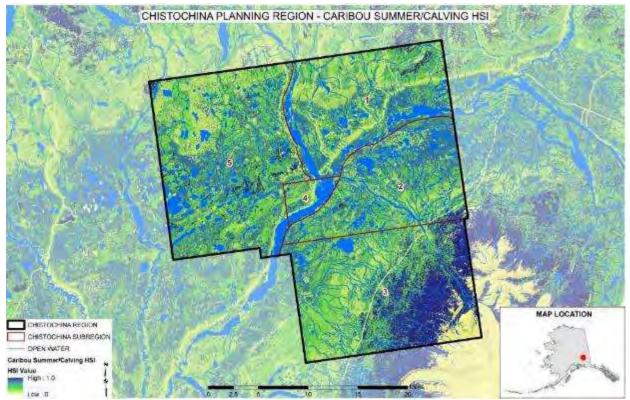


Figure 89. Habitat suitability for caribou during summer/calving in the Chistochina Planning Region of the Ahtna Traditional Use Territory.

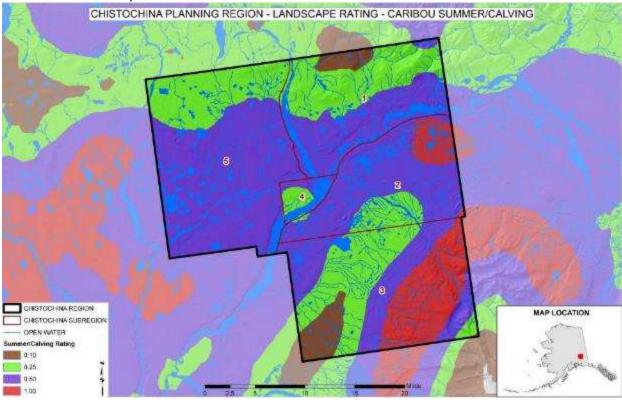


Figure 90. Landscape rating for caribou during summer/calving in the Chistochina Planning Region of the Ahtna Traditional Use Territory.

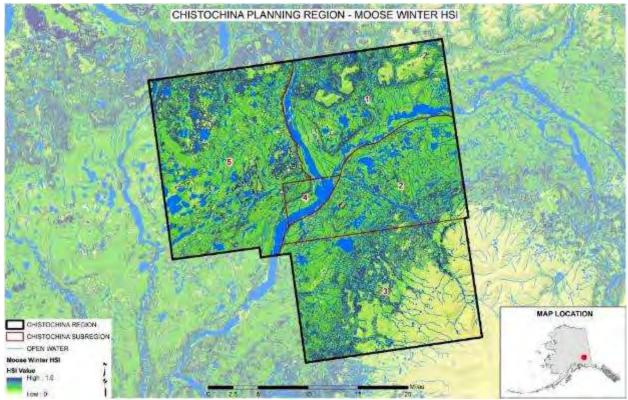


Figure 91. Habitat suitability for caribou during winter in the Chistochina Planning Region of the Ahtna Traditional Use Territory.

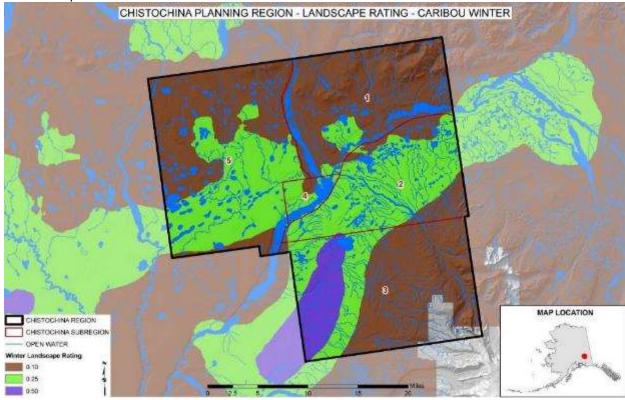


Figure 92. Landscape rating for caribou during winter in the Chistochina Planning Region of the Ahtna Traditional Use Territory.

Chistochina Site Improvement Areas

Potential treatment sites identified in the Chistochina area are displayed in figure 93. Based on all of the stated criteria, a number of sites were located for designated treatments. Figures 94-97 depict the locations of these areas. Appendix C provides a description of each of these sites. Sites ranging from approximately 16-55 acres in size were identified and are listed in Table 22. This table gives the site size and the treatment goal of primarily timber stand improvement and biomass production or moose browse enhancement. The table also displays an estimate of the total biomass of all standing above ground biomass. This estimate is not tonnage of usable biomass for energy but the total of all stems regardless of size. However, stands on the higher end of the total amount will typically have larger trees and are a candidate for biomass energy production.

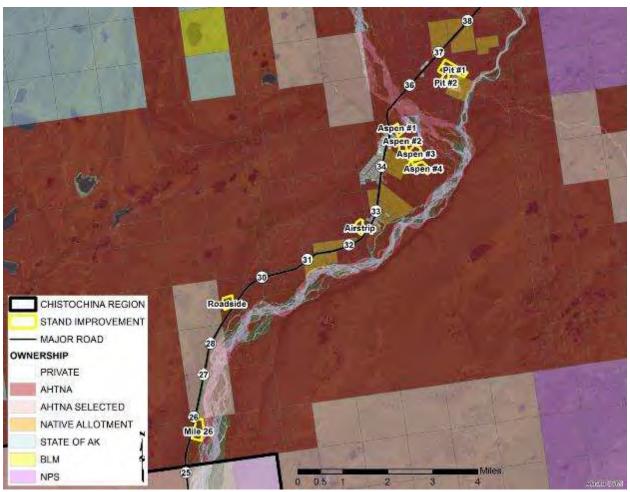


Figure 93. Overview of recommended treatment sites in the Chistochina management area.

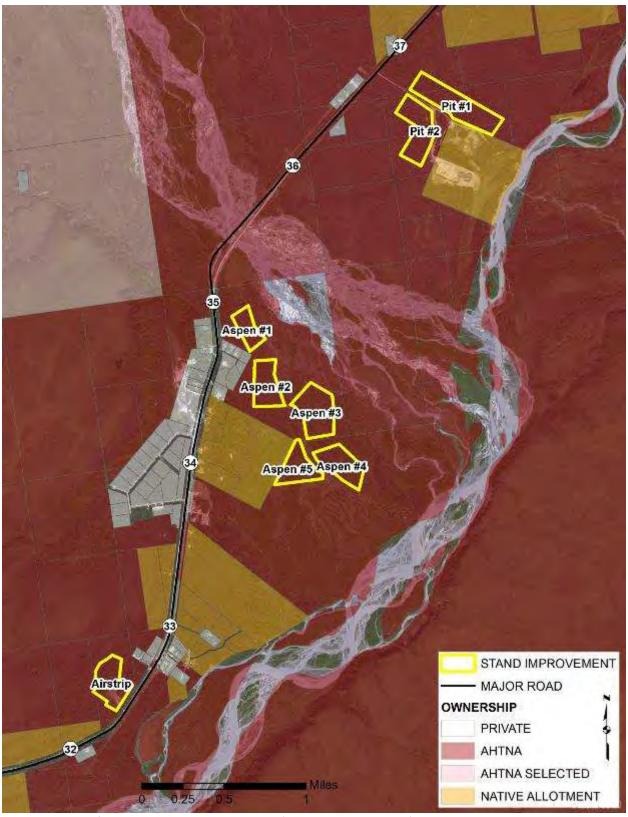


Figure 94. Map of three proposed treatment areas (Airstrip, Aspen, and Pit) in the Chistochina Planning Region showing surface ownership and aerial imagery.

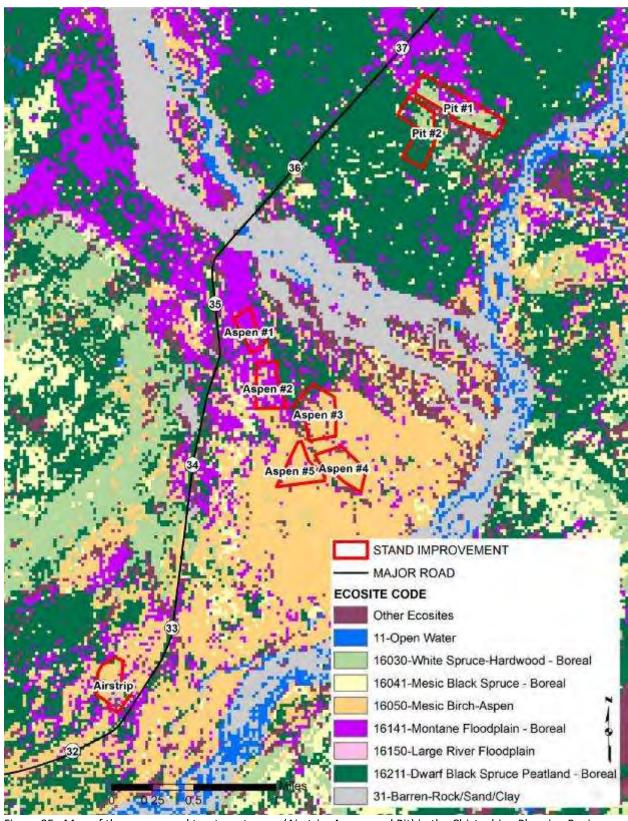


Figure 95. Map of three proposed treatment areas (Airstrip, Aspen, and Pit) in the Chistochina Planning Region showing ecological sites.

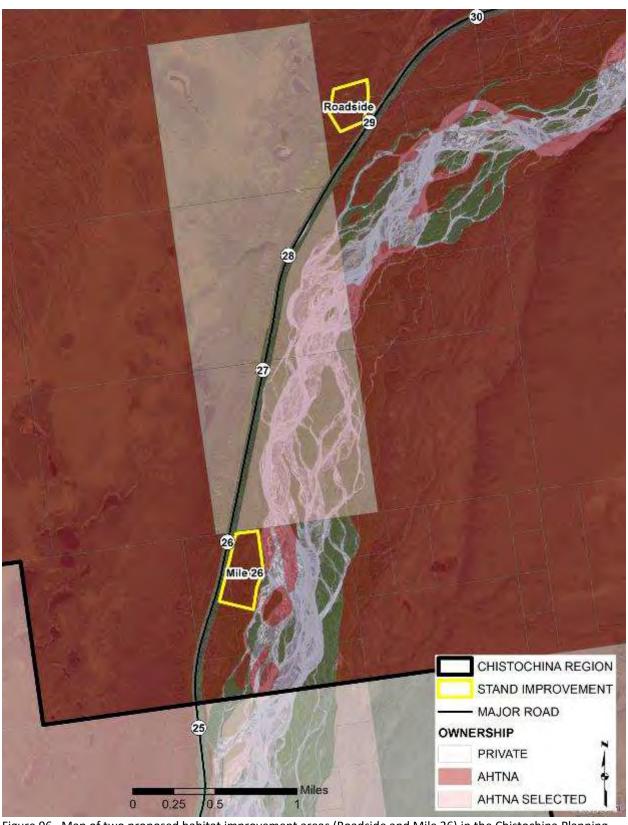


Figure 96. Map of two proposed habitat improvement areas (Roadside and Mile 26) in the Chistochina Planning Region showing surface ownership and aerial imagery.

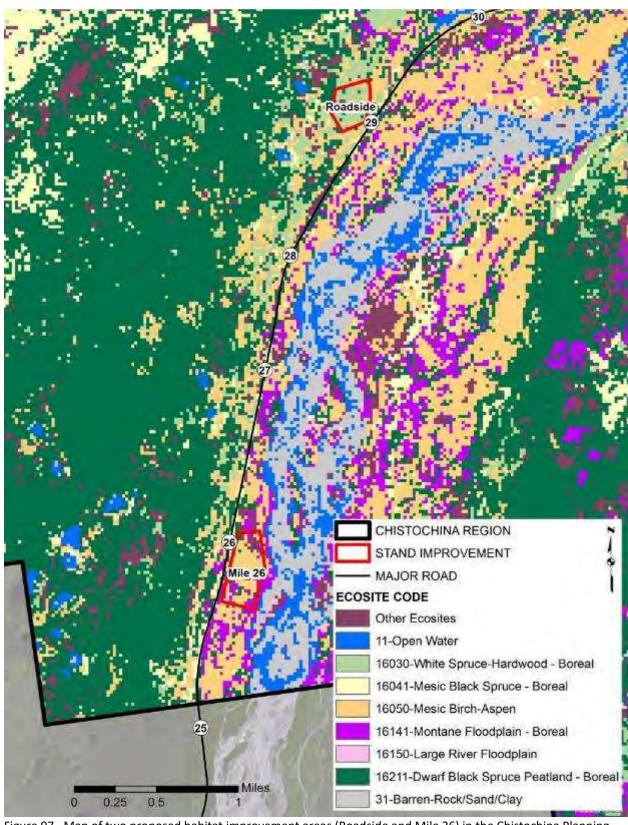


Figure 97. Map of two proposed habitat improvement areas (Roadside and Mile 26) in the Chistochina Planning Region showing ecological sites.

Table 22. Treatment site with treatment goal, site size, total tons of biomass and the BpS code and disturbance class for each site in the Chistochina Planning Region.

Sito Namo	Treatment Goal	Acres	Biomass - Total	BpS Code
Site Mairie			(tons)	_Dist Code
Airstrip	Moose Browse	28.9	110.3	16050_A
Aspen #1	Timber/Browse	16.2	10.0	16141_A
Aspen #2	Timber/Browse	26.1	3.1	16141_A
Aspen #3	Moose Browse	37.2	37.5	16050_A
Aspen #4	Moose Browse	27.4	39.5	16050_A
Aspen #5	Moose Browse	24.6	39.2	16050_A
Mile 26	Moose Browse	55.4	8.5	16050_A
Roadside	Timber/Browse	34.3	1111.1	16030_C

Chitina Planning Region

The Chitina Planning Region encompasses an area of 641,032 acres. Figure 98 displays the overall planning area along with associated infrastructure. Ownership of this and the surrounding area is shown in Figure 99. As Figure 99 displays, land ownership patterns are varied in this area with Ahtna owning 39.7% (254,179 acres) of the land. In addition, Chitina Native Corporation owns 16.5% (105,698 acres) of the planning region.

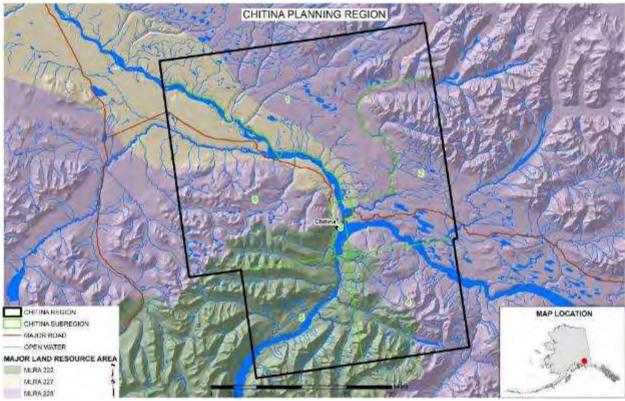


Figure 98. Overview of the Chitina Planning Region of the Ahtna Traditional Use Territory.

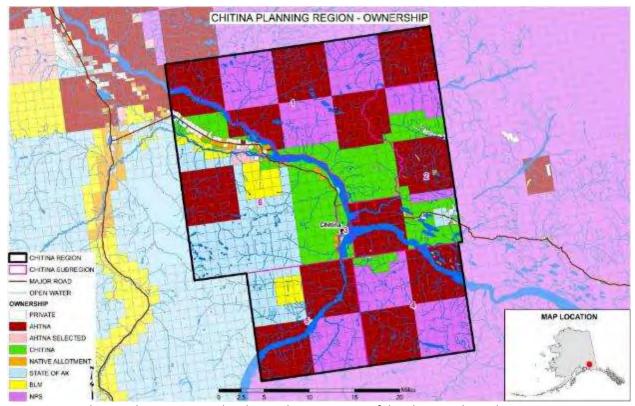


Figure 99. Land ownership patterns in the Chitina Planning Region of the Ahtna Traditional Use Territory.

Geology

The geology of the Chitina area was described previously based on the information from the <u>Copper River Basin Soil Survey</u>.

Climate

The Chitina area has a continental climate resulting from an influx of warm, moist air coming up the Copper River from the Gulf of Alaska. This makes it unique from the sub-arctic continental climate found in much of the Copper Basin. The climate in Chitina is characterized by long, cool winters and relatively warm summers. High winds are frequent due to the pressure gradient between coastal areas and the interior region of the Copper Basin.

The average minimum temperature in January is -16.8 °F (-27.1 °C); daily low temperatures of -50 °F (-46 °C) or less occur frequently during the winter and may last for two or more weeks. The average maximum temperature in July is 66.8 °F (19.3 °C) and on occasion exceed 85 °F (30 °C). Although the daily minimum temperature in summer averages in the forties, freezing temperatures have been recorded in every month resulting in a varied growing season length each year. Mean annual precipitation is 11.0 inches (27.9 cm) with 33% being received as rain during the growing season (June-August). Average annual snowfall is 26 inches (66.0 cm).

Soils

Soil texture in the Chitina project area is shown in Figure 100 and Figure 101 displays soil drainages.

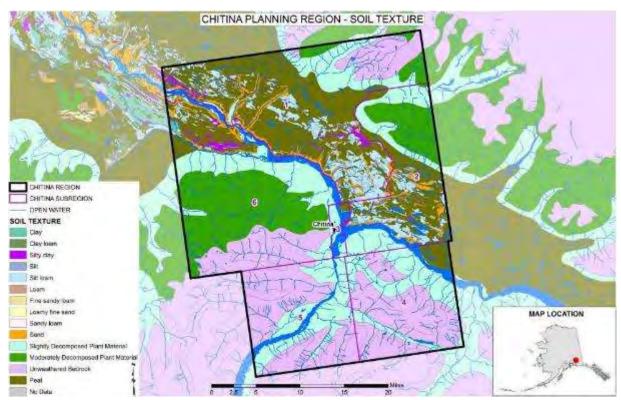


Figure 100. Soil texture in the Chitina Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

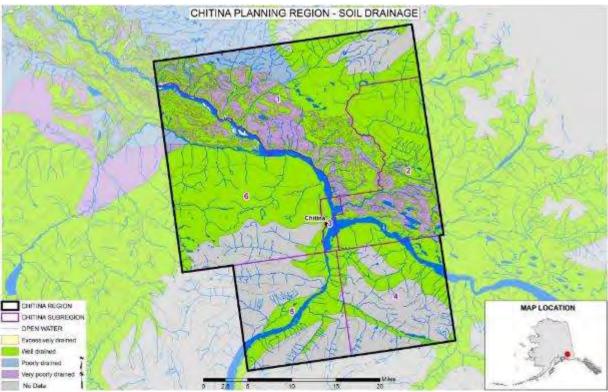


Figure 101. Soil drainage in the Chitina Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

Permafrost

Figure 102 displays the occurrence of permafrost in the Chitina project area as interpreted from the soil survey information.

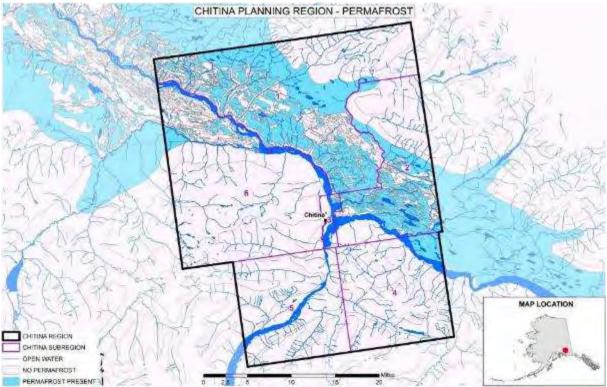


Figure 102. Permafrost in in the Chitina Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

Disturbance Factors

Fire is a major disturbance factor influencing the vegetation ecology in the project area. Figure 103 shows the fire history of the Chitina area along with current fire protection zones. While only one large fire has occurred since 1940 in the project area, fire is still a significant disturbance when it occurs. When fire does occur it serves to set back succession. It can also burn off the organic material at the ground surface, including peat that can occur on many sites. This can influence the thermal layer protecting the underlying permafrost on some sites, causing the permafrost to melt (thermokarst) and changing the site conditions through this process. Riparian areas are also influenced by flooding and ice events. These serve to set back succession of vegetation in riparian areas, and can even shift site conditions, particularly in the case of significant flooding events.

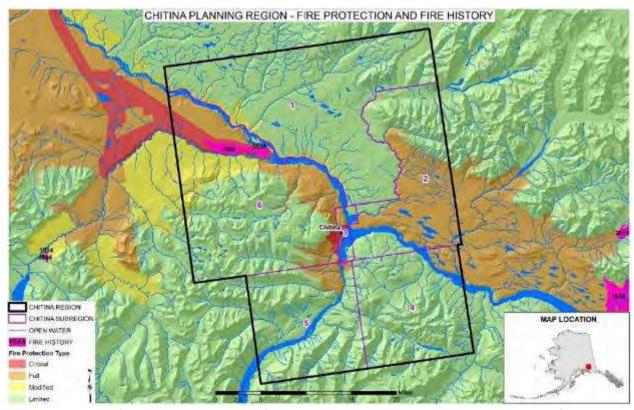


Figure 103. Current fire protection classes and fire history since 1940 in the Chitina Planning Region of the Ahtna Traditional Use Territory. Data from Alaska Interagency Coordination Center.

Ecosystem Diversity

The biophysical settings present within the Chitina Planning region are displayed in Figure 104. Table 23 displays the acres for each biophysical setting and disturbance class. Figure 105 is a map of ecosystem diversity (represented by biophysical setting and disturbance class) in the Chitina Planning Region.

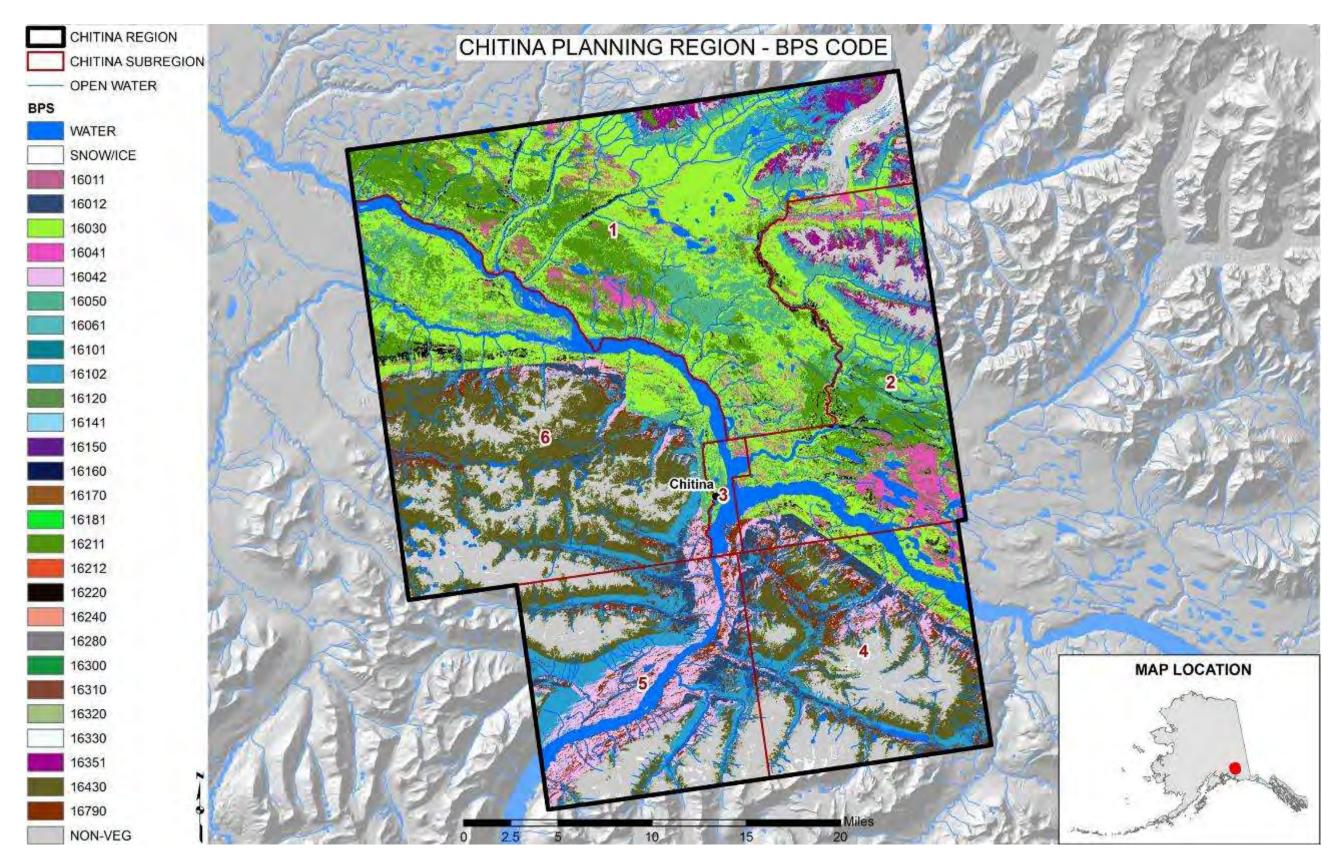


Figure 104. Biophysical settings (codes) in the Chitina Planning Region of the Ahtna Traditional Use Territory. Data from LANDFIRE.

Table 23. Acres by BpS code and disturbance class (A - E) in the Chitina planning region. The BpS vegetation label is provided as well.

IS provided as well.			Pac Codo		
BpS CodeDist Class	BpS Vegetation Label	Acres	BpS CodeDist Class	BpS Vegetation Label	Acres
11	Open Water	17723.3	16142_B	Montane Floodplain-Subboreal	0.4
12	Perrennial Ice-Snow	3879.2	16142_C	Montane Floodplain-Subboreal	0.2
16011_A	Treeline White Spruce-Boreal	151.7	16142_D	Montane Floodplain-Subboreal	0.9
16011_B	Treeline White Spruce-Boreal	733.5	16150_A	Large River Floodplain	486.6
16011_C	Treeline White Spruce-Boreal	0.2	16150_B	Large River Floodplain	100.5
16012_A	Treeline White Spruce-SubBoreal	1827.4	16150_C	Large River Floodplain	221.5
16012_B	Treeline White Spruce-SubBoreal	12358.0	16150_D	Large River Floodplain	11.1
16012_C	Treeline White Spruce-SubBoreal	8.2	16150_E	Large River Floodplain	1.1
16030_A	White Spruce-Hardwood-Boreal	2795.7	16160_A	Riparian Stringer	3.8
16030_B	White Spruce-Hardwood-Boreal	19679.1	16160_B	Riparian Stringer	6.0
16030_C	White Spruce-Hardwood-Boreal	107901.2	16170_A	Shrub and Herbaceous Floodplain	2.4
16030_E	White Spruce-Hardwood-Boreal	308.7	16170_B	Shrub and Herbaceous Floodplain	0.2
16041_A	Mesic Black Spruce-Boreal	812.9	16170_C	Shrub and Herbaceous Floodplain	0.2
16041_B	Mesic Black Spruce-Boreal	3682.2	16170_D	Shrub and Herbaceous Floodplain	1.8
16041_C	Mesic Black Spruce-Boreal	2921.8	16170_E	Shrub and Herbaceous Floodplain	2.7
16041_D	Mesic Black Spruce-Boreal	16317.1	16181_A	Herbaceous Fen	12.2
16041_E	Mesic Black Spruce-Boreal	4.9	16181_B	Herbaceous Fen	4.2
16042_A	Mesic Black Spruce-SubBoreal	4424.8	16181_C	Herbaceous Fen	2.0
16042_B	Mesic Black Spruce-SubBoreal	20517.5	16181_D	Herbaceous Fen	62.7
16042_C	Mesic Black Spruce-SubBoreal	7410.4	16211_A	Dwarf Black Spruce Peatland-Boreal	2843.3
16042_D	Mesic Black Spruce-SubBoreal	45.4	16211_B	Dwarf Black Spruce Peatland-Boreal	2707.2
16042_E	Mesic Black Spruce-SubBoreal	1.6	16211_C	Dwarf Black Spruce Peatland-Boreal	3593.2
16050_A	Mesic Birch-Aspen	22760.8	16211_D	Dwarf Black Spruce Peatland-Boreal	57194.2
16050_B	Mesic Birch-Aspen	15929.3	16212_A	Dwarf Black Spruce Peatland-Subboreal	2.4
16050_D	Mesic Birch-Aspen	37.8	16212_B	Dwarf Black Spruce Peatland-Subboreal	0.4
16050_E	Mesic Birch-Aspen	69.4	16212_C	Dwarf Black Spruce Peatland-Subboreal	3.6
16061_A	Dry Aspen-Steppe Bluff	95.9	16220_A	Black Spruce Wet-Mesic Slope	1386.2
16061_B	Dry Aspen-Steppe Bluff	343.8	16220_B	Black Spruce Wet-Mesic Slope	1058.4
16061_C	Dry Aspen-Steppe Bluff	423.0	16220_C	Black Spruce Wet-Mesic Slope	359.8
16061_D	Dry Aspen-Steppe Bluff	746.8	16220_D	Black Spruce Wet-Mesic Slope	5594.3
16070_A	Subalpine Balsam Poplar-Aspen	0.2	16240_A	Deciduous Shrub Swamp	55.8
16090_A	Mesic Subalpine Alder	0.7	16240_B	Deciduous Shrub Swamp	6.9
16102_A	Mesic Scrub Birch-Willow	13391.5	16280_A	Low Shrub-Tussock Tundra	2142.3
16102_B	Mesic Scrub Birch-Willow	43089.5	16280_B	Low Shrub-Tussock Tundra	268.9
16110_A	Mesic Bluejoint Meadow	1.6	16280_C	Low Shrub-Tussock Tundra	24.9
16110_B	Mesic Bluejoint Meadow	0.2	16290_B	Tussock Tundra	0.2
16120_A	Dry Grassland	3.6	16300_A	Wet Black Spruce-Tussock	8.5
16120_B	Dry Grassland	0.7	16300_B	Wet Black Spruce-Tussock	17.6
16141_A	Montane Floodplain-Boreal	489.7	16300_C	Wet Black Spruce-Tussock	7.3
16141_B	Montane Floodplain-Boreal	1225.0	16310_A	Alpine Dwarf-Shrub Summit	1412.9
16141_C	Montane Floodplain-Boreal	607.6	16310_B	Alpine Dwarf-Shrub Summit	222.0
16141_D	Montane Floodplain-Boreal	6634.3	16320_A	Alpine Talus and Bedrock	13007.7
16141_E	Montane Floodplain-Boreal	487.5	16320_B	Alpine Talus and Bedrock	209.7
16142_A	Montane Floodplain-Subboreal	0.4	16330_A	Alpine Mesic Herbaceous Meadow	7.6

BpS Code _Dist Class	BpS Vegetation Label	Acres	BpS Code _Dist Class	BpS Vegetation Label	Acres
16351_A	Alpine Ericaceous Dwarf-Shrubland	11226.1	16620_A	Emergent Marsh	0.2
16351_B	Alpine Ericaceous Dwarf-Shrubland	483.3	16790_A	White Spruce-Hardwood-SubBoreal	18801.3
16430_A	Alpine Dwarf Shrubland	70254.8	16790_B	White Spruce-Hardwood-SubBoreal	9383.5
16430_B	Alpine Dwarf Shrubland	5725.3	16790_C	White Spruce-Hardwood-SubBoreal	1097.3
16520_A	Subalpine Alder-Salmonberry	0.2	16790_D	White Spruce-Hardwood-SubBoreal	91.6
16520_B	Subalpine Alder-Salmonberry	0.2	16790_E	White Spruce-Hardwood-SubBoreal	12.7
16550_A	Montane Floodplain	0.2	31	Barren-Rock-Sand-Clay	101067.9

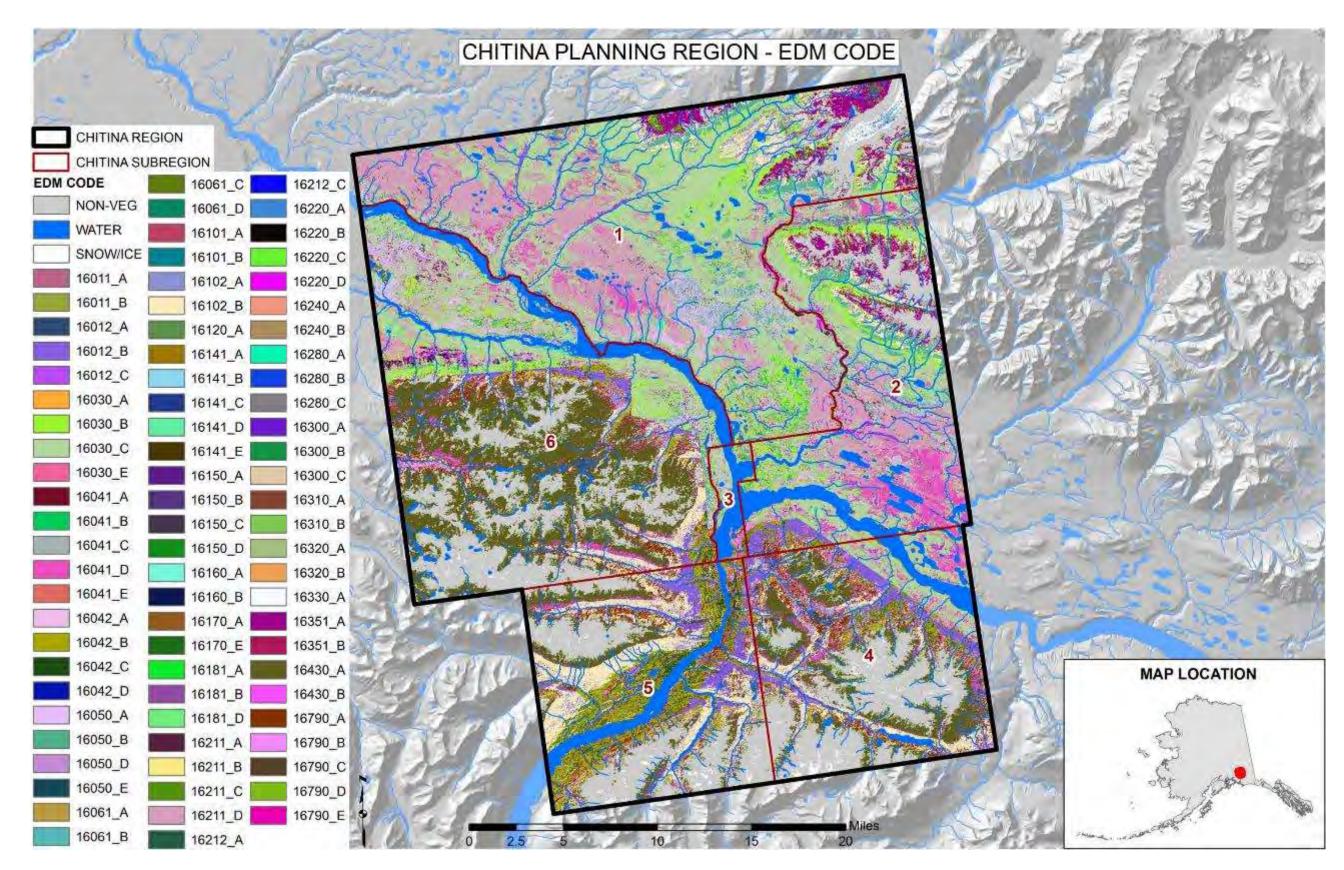


Figure 105. Map of ecosystem diversity in the Chitina Planning Region of the Ahtna Traditional Use Territory. Numbers in the legend refer to the BpS code, while letters (A-E) refer to the disturbance class.

Berry Production Areas

Ecological sites that had the potential for producing desirable species of berries were identified. These sites were then further identified as to the disturbance states most likely to support berry production. While numerous factors can influence where berry-producing vegetation occurs, the combination of ecological sites and disturbance states allows for the identification of areas with higher probabilities of finding good berry production. Figure 106 shows the areas identified with the highest potential for berry production. It may be desirable to avoid disturbances to these areas, but ground-truthing of the actual presence of good berry production is recommended.

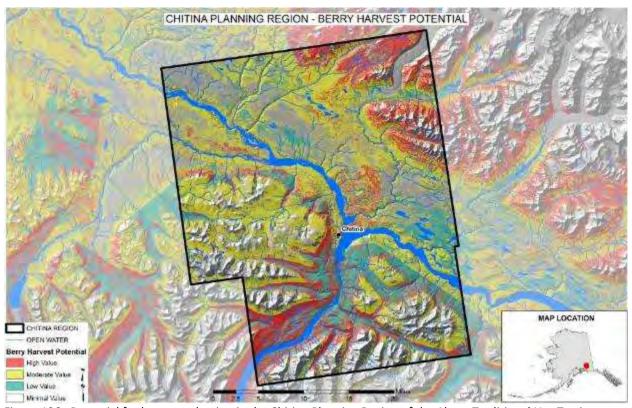


Figure 106. Potential for berry production in the Chitina Planning Region of the Ahtna Traditional Use Territory based on ecological sites and disturbance states derived from remotely sensed information.

Moose and Caribou Model Results

The habitat suitability figure for moose spring habitat is found in Figure 107 and the landscape rating for moose spring habitat is found in Figure 108. The habitat suitability figure for moose summer habitat is found in Figure 109 and the landscape rating for moose summer habitat is found in Figure 110. The habitat suitability figure for moose winter habitat is found in Figure 111 and the landscape rating for moose winter habitat is found in Figure 112. The habitat suitability figure for caribou summer/calving habitat are found in Figure 113 and the landscape rating for caribou summer/calving habitat are found Figure 114. The habitat suitability figure for caribou winter habitat are found in Figure 115 and the landscape rating for caribou winter habitat are found Figure 116. A complete description of the moose habitat model can be found in Appendix A. A complete description of the caribou habitat model can be found in Appendix B.

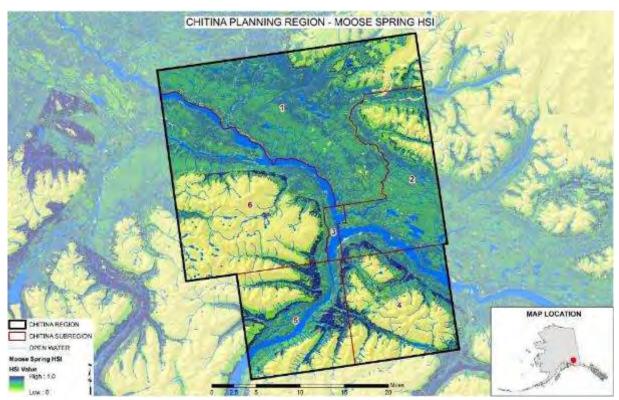


Figure 107. Habitat suitability for moose during spring in the Chitina Planning Region of the Ahtna Traditional Use Territory.

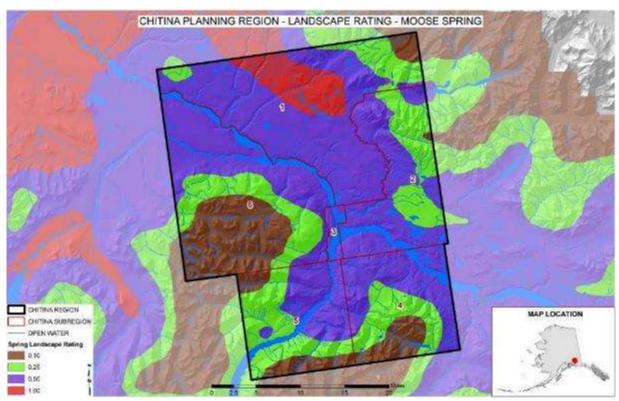


Figure 108. Landscape rating for moose during spring in the Chitina Planning Region of the Ahtna Traditional Use Territory.

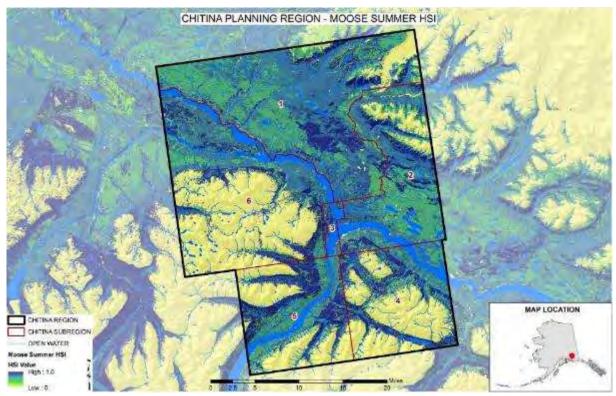


Figure 109. Habitat suitability for moose during summer in the Chitina Planning Region of the Ahtna Traditional Use Territory.

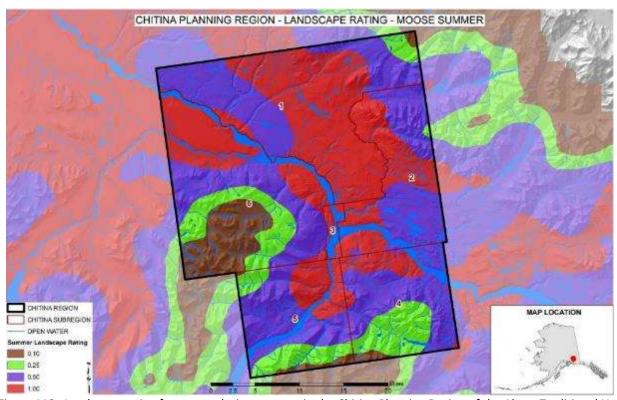


Figure 110. Landscape rating for moose during summer in the Chitina Planning Region of the Ahtna Traditional Use Territory.

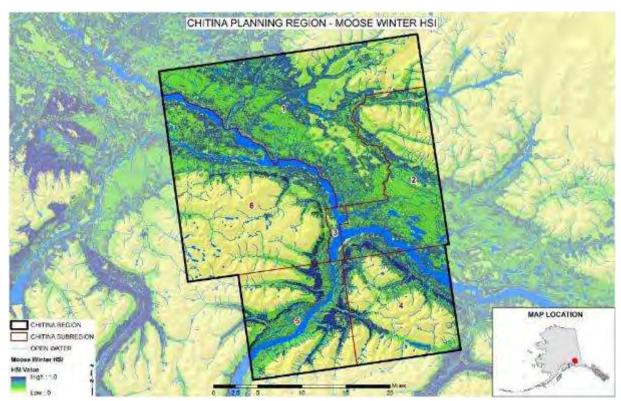


Figure 111. Habitat suitability for moose during winter in the Chitina Planning Region of the Ahtna Traditional Use Territory.

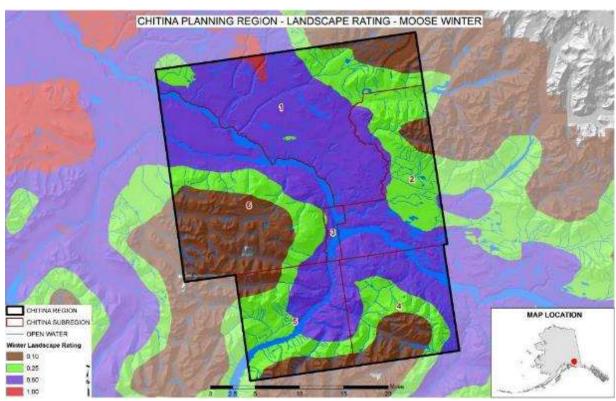


Figure 112. Landscape rating for moose during winter in the Chitina Planning Region of the Ahtna Traditional Use Territory.

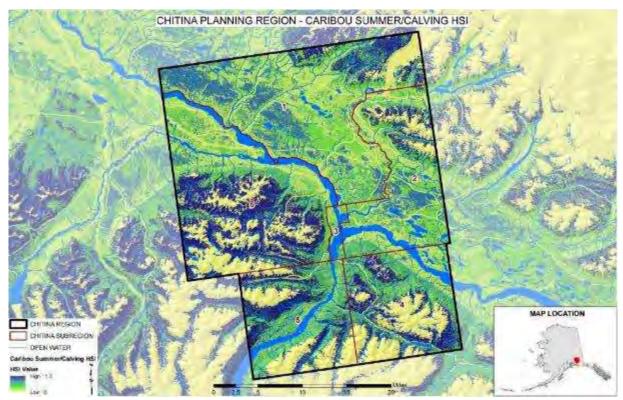


Figure 113. Habitat suitability for caribou during summer/calving in the Chitina Planning Region of the Ahtna Traditional Use Territory.

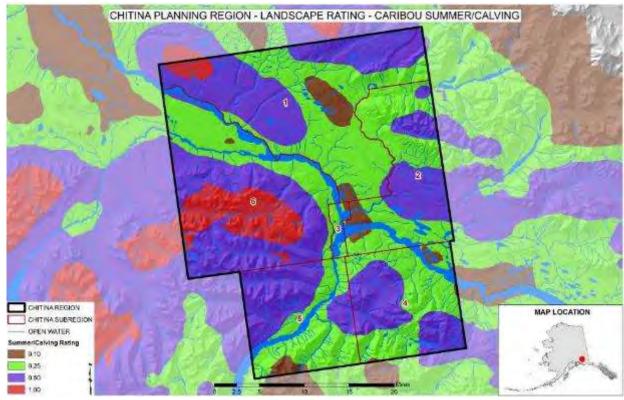


Figure 114. Landscape rating for caribou during summer/calving in the Chitina Planning Region of the Ahtna Traditional Use Territory.

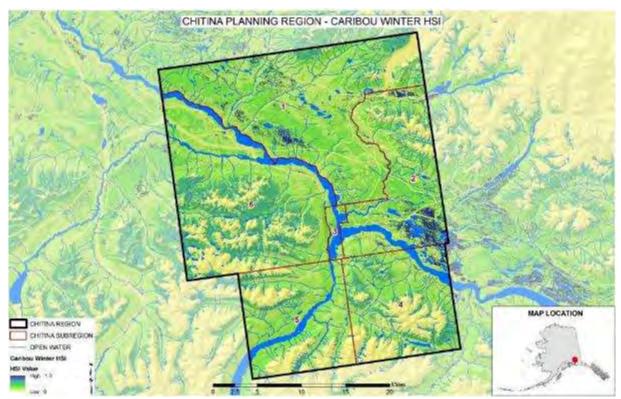


Figure 115. Habitat suitability for caribou during winter in the Chitina Planning Region of the Ahtna Traditional Use Territory.

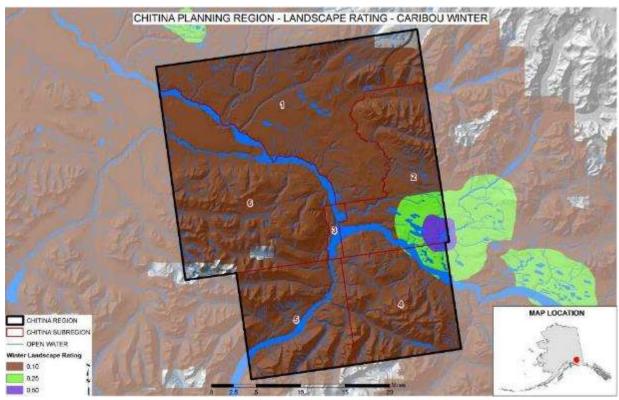


Figure 116. Landscape rating for caribou during winter in the Chitina Planning Region of the Ahtna Traditional Use Territory.

Chitina Site Improvement Areas

Potential treatment sites identified in the Chitina area are displayed in figure 117. Based on all of the stated criteria, a number of sites were located for designated treatments. Figures 118-121 depict the locations of these areas. Appendix C provides a description of each of these sites. Sites ranging from approximately 7-46 acres in size were identified and are listed in Table 24. This table gives the site size and the treatment goal of primarily timber stand improvement and biomass production or moose browse enhancement. The table also displays an estimate of the total biomass of all standing above ground biomass. This estimate is not tonnage of usable biomass for energy but the total of all stems regardless of size. However, stands on the higher end of the total amount will typically have larger trees and are a candidate for biomass energy production.

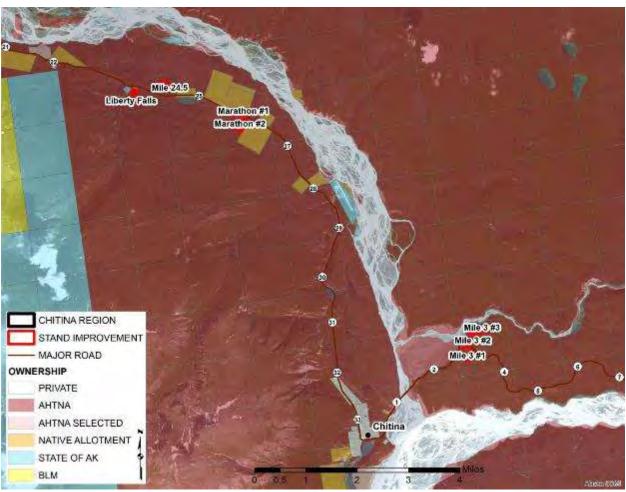


Figure 117. Overview of recommended treatment sites in the Chitina management area.

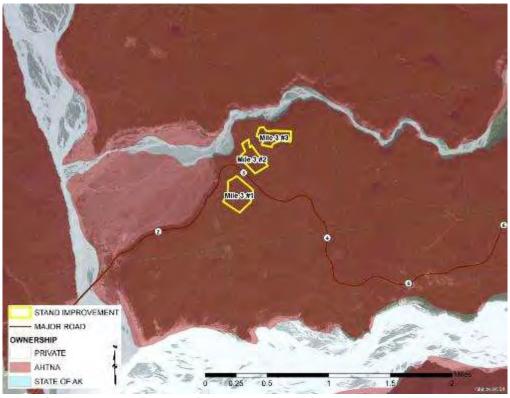


Figure 118. Map of proposed treatment areas (Mile 3 #1,#2, and #3) in the Chitina Planning Region showing surface ownership and aerial imagery.

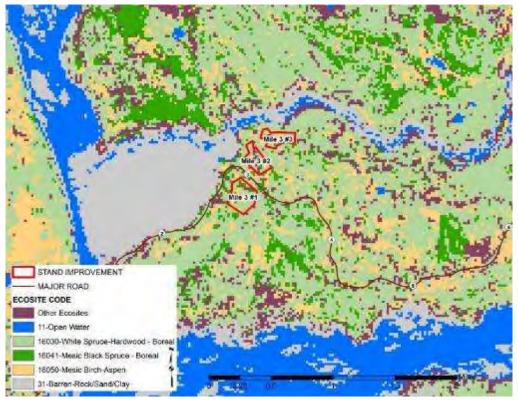


Figure 119. Map of proposed treatment areas (Mile 3 #1,#2, and #3) in the Chitina Planning Region showing ecological sites.

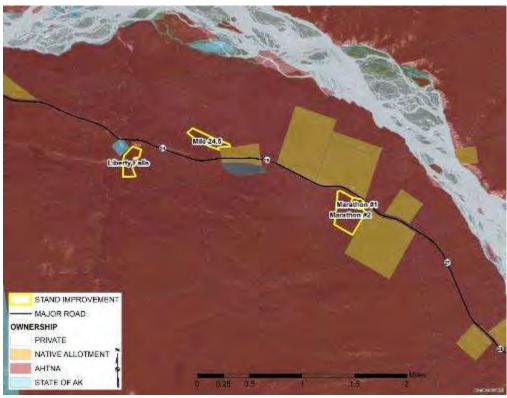


Figure 120. Map of proposed habitat improvement areas (Liberty Falls, Mile 24.5, and Marathon #1 and #2) in the Chitina Planning Region showing surface ownership and aerial imagery.

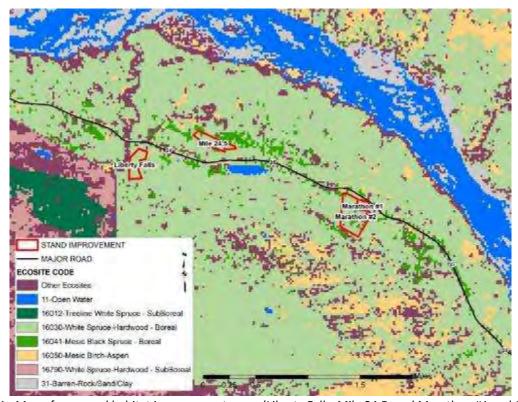


Figure 121. Map of proposed habitat improvement areas (Liberty Falls, Mile 24.5, and Marathon #1 and #2) in the Chitina Planning Region showing ecological sites.

Table 24. Treatment site with treatment goal, site size, total tons of biomass and the BpS code and disturbance class for each site in the Chitina Planning Region.

Site Name	Treatment Goal	Acres	Biomass - Total	BpS Code
Site Name			(tons)	_Dist Code
Liberty Falls		19.62	862.6	16030_C
Mile 24.5		19.10	90.7	16030_C
Marathon #1		6.95	2062.3	16030_C
Marathon #2		46.05	865.9	16030_C
Mile 3 #2	Moose Browse	17.78	239.8	16030_B
Mile 3 #3		16.88	253.4	16030_B
Mile 3 #1	Moose Browse	28.13	172.0	16030_C

Gakona Planning Region

The Gakona Planning Region includes an area of 641,274 acres. Figure 122 displays the overall planning area along with associated infrastructure. Ownership of this and the surrounding area is shown in Figure 123. As Figure 123 displays, land ownership patterns are varied in this area with Ahtna owning 42.0% (269,100 acres) of the land.

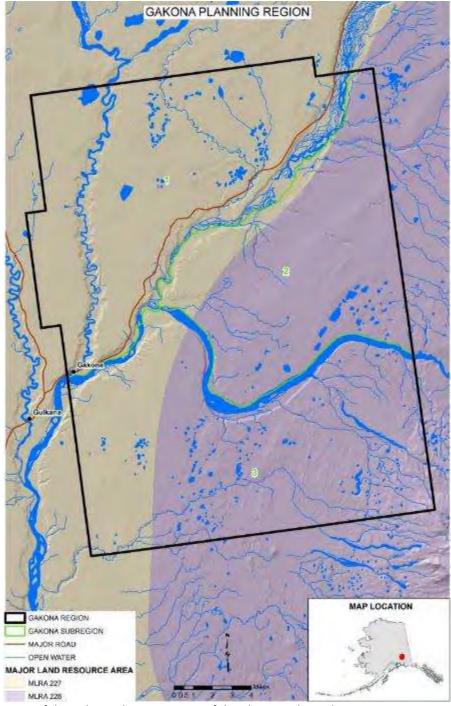


Figure 122. Overview of the Gakona Planning Region of the Ahtna Traditional Use Territory.

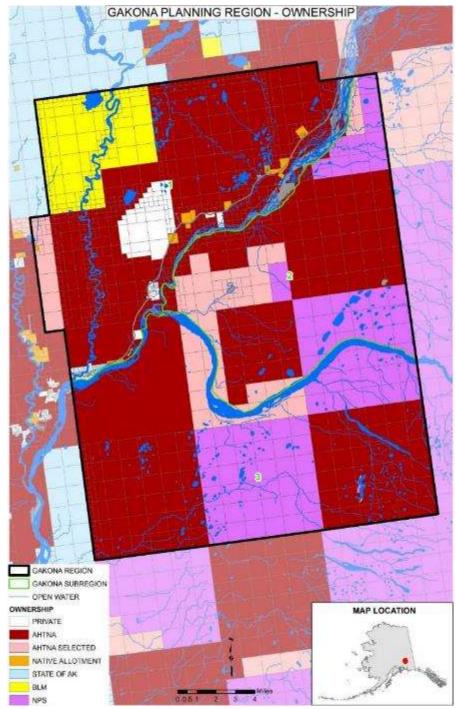


Figure 123. Land ownership patterns in the Gakona Planning Region of the Ahtna Traditional Use Territory.

A description of the general geology, climate, soils, permafrost, and vegetation is found in Chapter 1 of this report. Figures showing these features specific to the Gakona Planning Region are displayed below. Soil texture in the Gakona area is shown in Figure 124 and Figure 125 displays soil drainages. Permafrost in the Gakona area is shown in Figure 126.

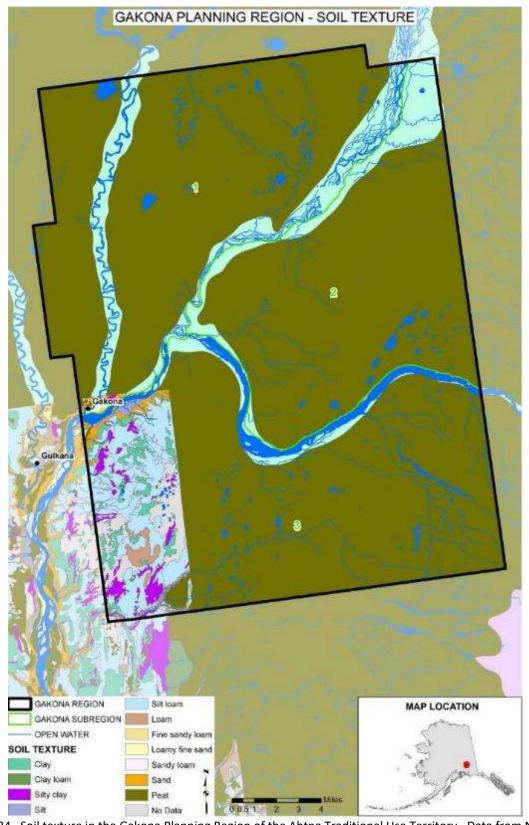


Figure 124. Soil texture in the Gakona Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

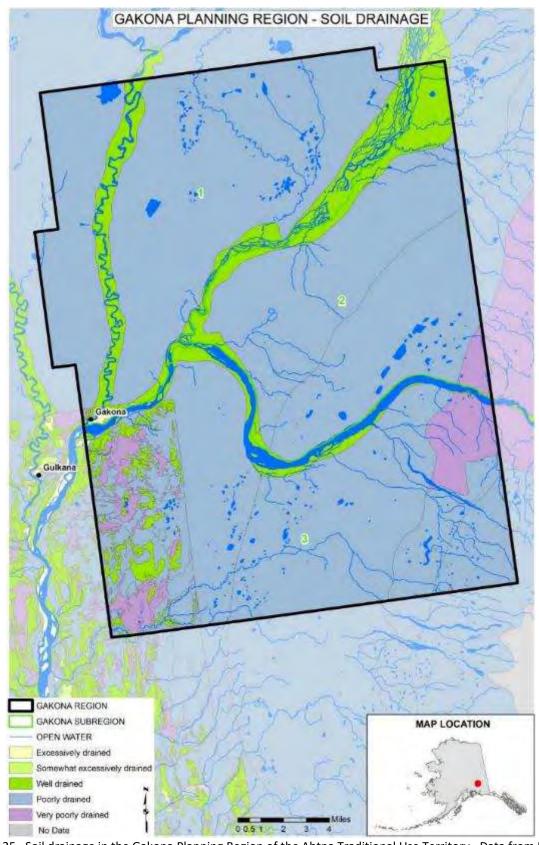


Figure 125. Soil drainage in the Gakona Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

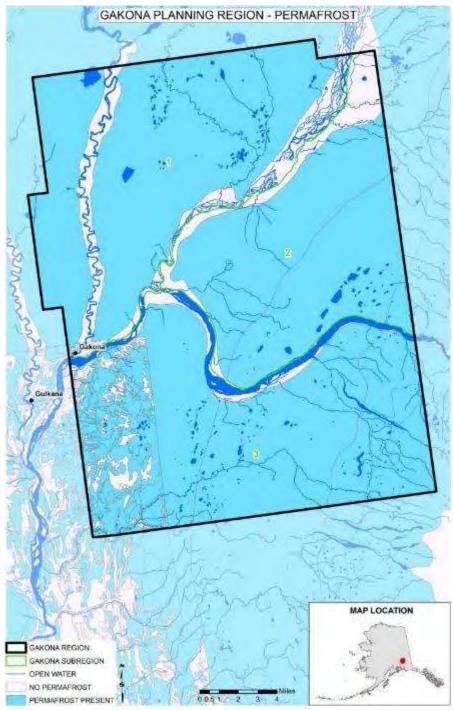


Figure 126. Permafrost in the Gakona Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

Disturbance Factors

Fire is the primary disturbance factor influencing the vegetation ecology in the Copper Basin. Figure 127 shows the fire history of the Gakona area along with current fire protection zones. Although the level of fire occurring in this southcentral Alaska landscape is substantially less than that occurring in more interior areas of Alaska north of the Alaska Range, fire is still a significant disturbance when it occurs.

However, there has not been a large fire in the project area since 1947. When fire does occur it serves to set back succession. It can also burn off the organic material at the ground surface, including peat that can occur on many sites. This can influence the thermal layer protecting the underlying permafrost on some sites, causing the permafrost to melt (thermokarst) and changing the site conditions through this process. Riparian areas are also influenced by flooding and ice events. These serve to set back succession of vegetation in riparian areas, and can even shift site conditions, particularly in the case of significant flooding events.

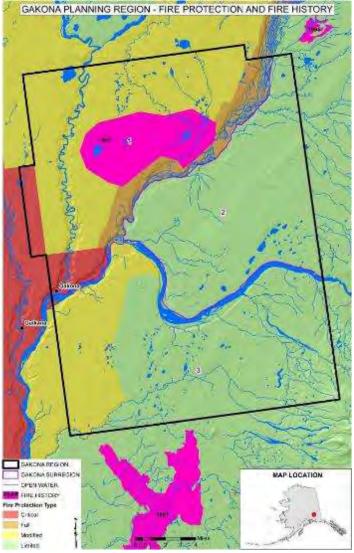


Figure 127. Current fire protection classes and fire history since 1940 in the Gakona Planning Region of the Ahtna Traditional Use Territory. Data from Alaska Interagency Coordination Center.

Ecosystem Diversity

The biophysical settings present within the Gakona Planning region are displayed in Figure 128. Table 25 displays the acres for each biophysical setting and disturbance class. Figure 129 is a map of ecosystem diversity (represented by biophysical setting and disturbance class) in the Gakona Planning Region.

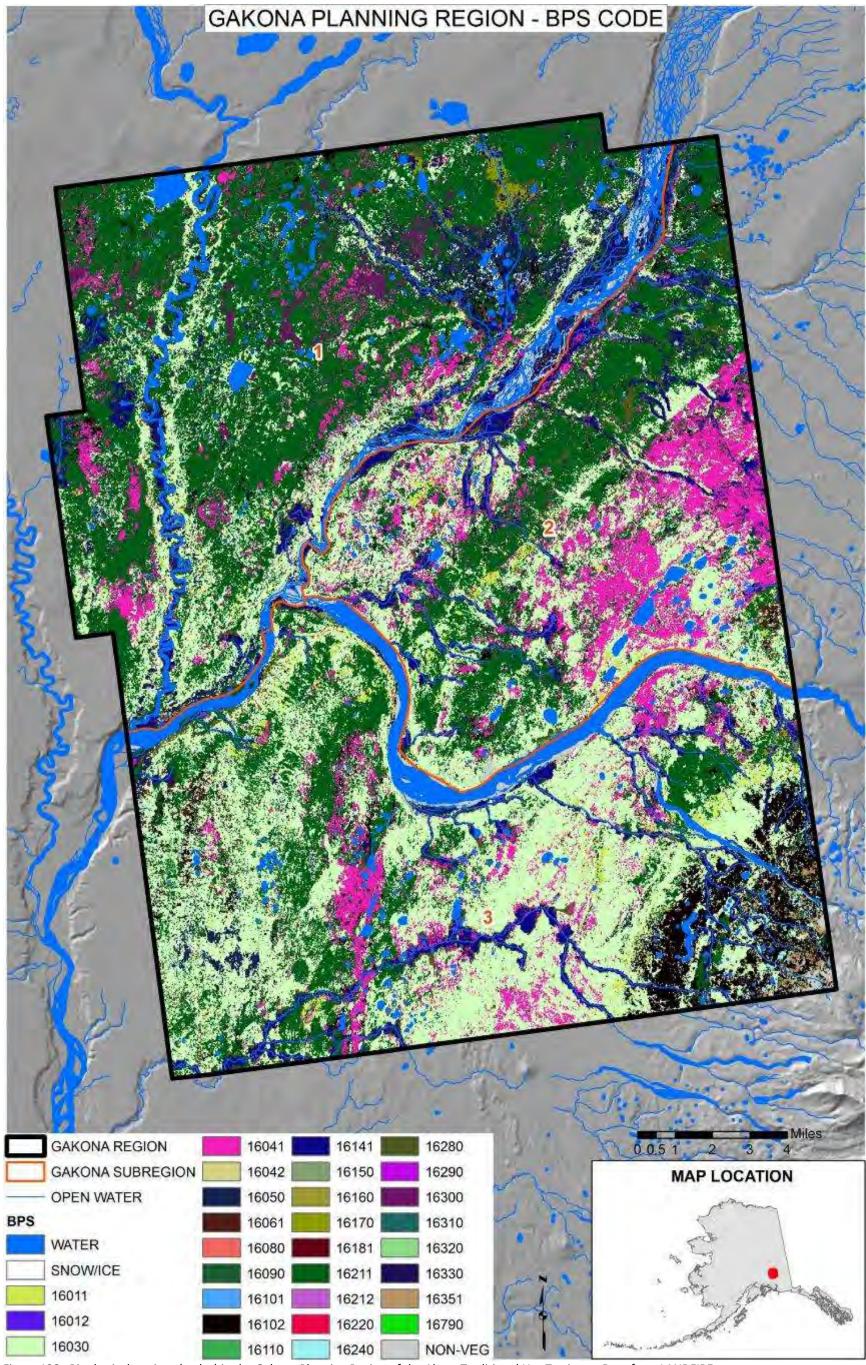


Figure 128. Biophysical settings (codes) in the Gakona Planning Region of the Ahtna Traditional Use Territory. Data from LANDFIRE.

Table 25. Acres by BpS code and disturbance class (A – E) in the Gakona Planning Region. The BpS vegetation label is provided as well.

BpS Code _Dist Class	·		BpS Code _Dist Class	BpS Vegetation Label	Acres	
11	Open Water	6267.8	16150_E	Large River Floodplain	0.4	
12	Perrennial Ice-Snow	39.6	16160_A	Riparian Stringer	13.1	
16011_A	Treeline White Spruce-Boreal	554.7	16160_C	Riparian Stringer	0.7	
16011_B	Treeline White Spruce-Boreal	3570.8	16170_A	Shrub and Herbaceous Floodplain	846.7	
16011_C	Treeline White Spruce-Boreal	0.7	16170_B	Shrub and Herbaceous Floodplain	61.8	
16012_A	Treeline White Spruce-SubBoreal	7.8	16170_C	Shrub and Herbaceous Floodplain	4.0	
16012_B	Treeline White Spruce-SubBoreal	10.2	16170_D	Shrub and Herbaceous Floodplain	17.3	
16030_A	White Spruce-Hardwood-Boreal	3060.4	16170_E	Shrub and Herbaceous Floodplain	307.8	
16030_B	White Spruce-Hardwood-Boreal	9997.1	16181_A	Herbaceous Fen	25.1	
16030_C	White Spruce-Hardwood-Boreal	72372.9	16181_B	Herbaceous Fen	3.3	
16030_E	White Spruce-Hardwood-Boreal	73.6	16181_C	Herbaceous Fen	0.4	
16041_A	Mesic Black Spruce-Boreal	1065.5	16181_D	Herbaceous Fen	13.8	
16041_B	Mesic Black Spruce-Boreal	2198.8	16211_A	Dwarf Black Spruce Peatland-Boreal	3723.8	
16041_C	Mesic Black Spruce-Boreal	937.6	16211_B	Dwarf Black Spruce Peatland-Boreal	3661.1	
16041_D	Mesic Black Spruce-Boreal	23704.6	16211_C	Dwarf Black Spruce Peatland-Boreal	8313.8	
16041_E	Mesic Black Spruce-Boreal	6.0	16211_D	Dwarf Black Spruce Peatland-Boreal	77256.0	
16042_A	Mesic Black Spruce-SubBoreal	0.4	16212_A	Dwarf Black Spruce Peatland-Subboreal	0.4	
16042_B	Mesic Black Spruce-SubBoreal	5.3	16212_B	Dwarf Black Spruce Peatland-Subboreal	2.0	
16042_C	Mesic Black Spruce-SubBoreal	0.9	16212_C	Dwarf Black Spruce Peatland-Subboreal	2.7	
16050_A	Mesic Birch-Aspen	9714.0	16220_A	Black Spruce Wet-Mesic Slope	53.8	
16050_B	Mesic Birch-Aspen	1659.5	16220_B	Black Spruce Wet-Mesic Slope	19.8	
16050_D	Mesic Birch-Aspen	10.2	16220_C	Black Spruce Wet-Mesic Slope	12.7	
16050_E	Mesic Birch-Aspen	15.1	16220_D	Black Spruce Wet-Mesic Slope	269.5	
16061_A	Dry Aspen-Steppe Bluff	46.7	16240_A	Deciduous Shrub Swamp	351.6	
16061_B	Dry Aspen-Steppe Bluff	222.0	16240_B	Deciduous Shrub Swamp	14.0	
16061_C	Dry Aspen-Steppe Bluff	160.3	16280_A	Low Shrub-Tussock Tundra	4927.8	
16061_D	Dry Aspen-Steppe Bluff	844.0	16280_B	Low Shrub-Tussock Tundra	1065.9	
16080_A	Avalanche Slope Shrubland	4.9	16280_C	Low Shrub-Tussock Tundra	116.8	
16080_B	Avalanche Slope Shrubland	21.3	16290_A	Tussock Tundra	23.8	
16090_A	Mesic Subalpine Alder	108.5	16290_B	Tussock Tundra	18.9	
16090_B	Mesic Subalpine Alder	203.7	16300_A	Wet Black Spruce-Tussock	250.2	
16102_A	Mesic Scrub Birch-Willow	4214.8	16300_B	Wet Black Spruce-Tussock	870.7	
16102_B	Mesic Scrub Birch-Willow	7426.2	16300_C	Wet Black Spruce-Tussock	1518.3	
16110_A	Mesic Bluejoint Meadow	270.4	16310_A	Alpine Dwarf-Shrub Summit	34.9	
16110_B	Mesic Bluejoint Meadow	1.3	16320_A	Alpine Talus and Bedrock	1140.4	
16141_A	Montane Floodplain-Boreal	803.7	16320_B	Alpine Talus and Bedrock	32.7	
16141_B	Montane Floodplain-Boreal	1539.2	16330_A	Alpine Mesic Herbaceous Meadow	14.0	
16141_C	Montane Floodplain-Boreal	608.7	16330_B	Alpine Mesic Herbaceous Meadow	1.1	
16141_D	Montane Floodplain-Boreal	10135.7	16351_A	Alpine Ericaceous Dwarf-Shrubland	878.9	
16141_E	Montane Floodplain-Boreal	459.2	16351_B	Alpine Ericaceous Dwarf-Shrubland	23.8	
16150_A	Large River Floodplain	381.0	16790_A	White Spruce-Hardwood-SubBoreal	5.1	
16150_B	Large River Floodplain	40.9	16790_B	White Spruce-Hardwood-SubBoreal	5.8	
16150_C	Large River Floodplain	113.4	31	Barren-Rock-Sand-Clay	8564.4	
16150_D	Large River Floodplain	4.225505				

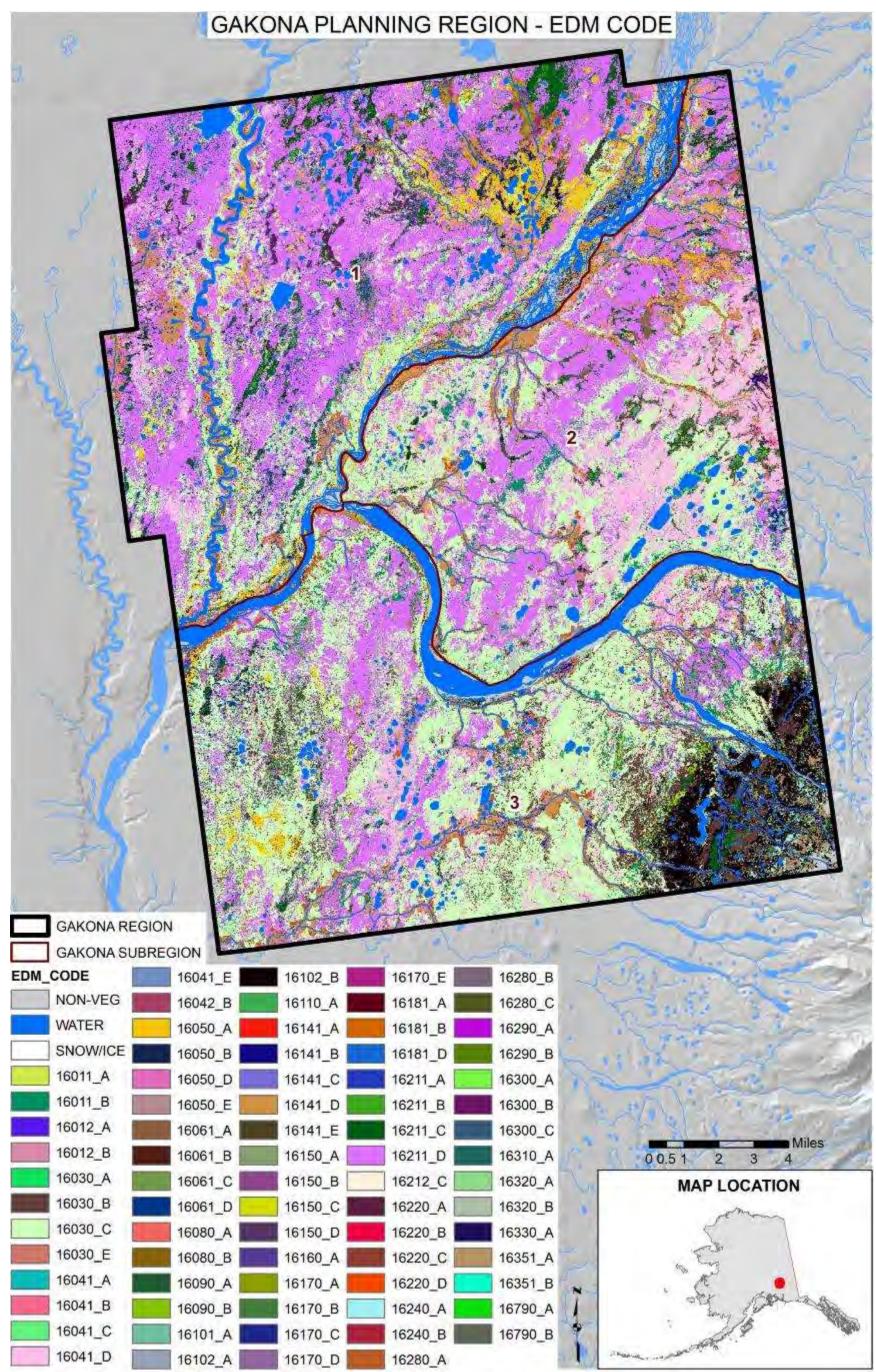


Figure 129. Map of ecosystem diversity in the Gakona Planning Region of the Ahtna Traditional Use Territory. Numbers in the legend refer to the BpS code, while letters (A-E) refer to the disturbance class.

Berry Production Areas

Ecological sites that had the potential for producing desirable species of berries were identified. These sites were then further identified as to the disturbance states most likely to support berry production. While numerous factors can influence where berry-producing vegetation occurs, the combination of ecological sites and disturbance states allows for the identification of areas with higher probabilities of finding good berry production. Figure 130 shows the areas identified with the highest potential for berry production. It may be desirable to avoid disturbances to these areas, but ground-truthing of the actual presence of good berry production is recommended.

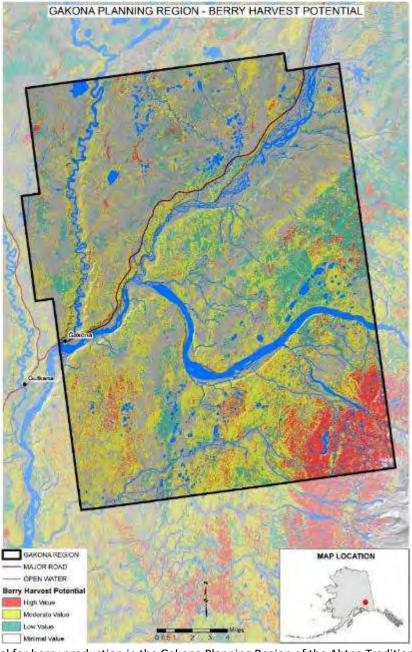


Figure 130. Potential for berry production in the Gakona Planning Region of the Ahtna Traditional Use Territory based on ecological sites and disturbance states derived from remotely sensed information.

Moose and Caribou Model Results

The habitat suitability figure for moose spring habitat is found in Figure 131 and the landscape rating for moose spring habitat is found in Figure 132. The habitat suitability figure for moose summer habitat is found in Figure 133 and the landscape rating for moose summer habitat is found in Figure 134. The habitat suitability figure for moose winter habitat is found in Figure 135 and the landscape rating for moose winter habitat is found in Figure 136. The habitat suitability figure for caribou summer/calving habitat are found in Figure 137 and the landscape rating for caribou summer/calving habitat are found Figure 138. The habitat suitability figure for caribou winter habitat are found in Figure 139 and the landscape rating for caribou winter habitat are found Figure 140. A complete description of the moose habitat model can be found in Appendix A. A complete description of the caribou habitat model can be found in Appendix B.

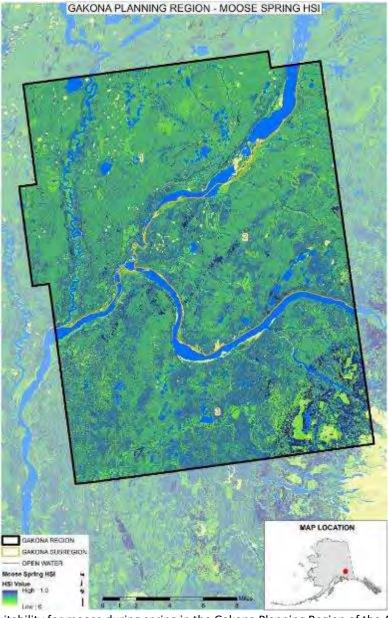


Figure 131. Habitat suitability for moose during spring in the Gakona Planning Region of the Ahtna Traditional Use Territory.

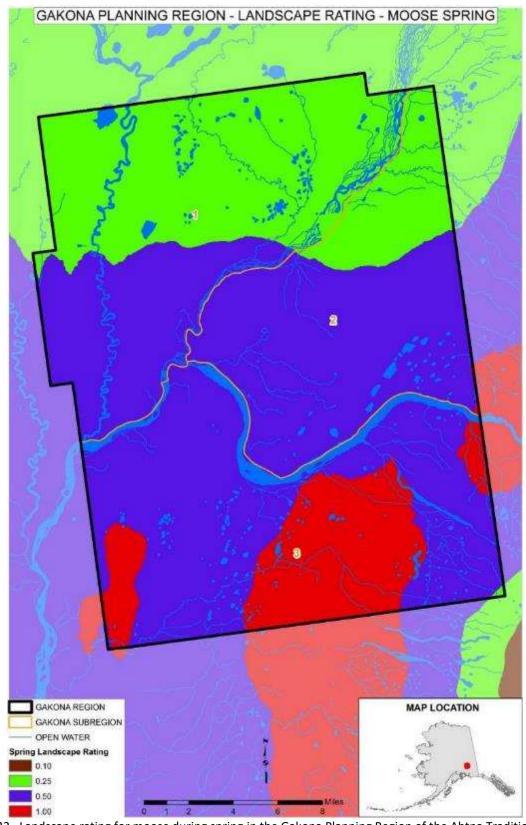


Figure 132. Landscape rating for moose during spring in the Gakona Planning Region of the Ahtna Traditional Use Territory.

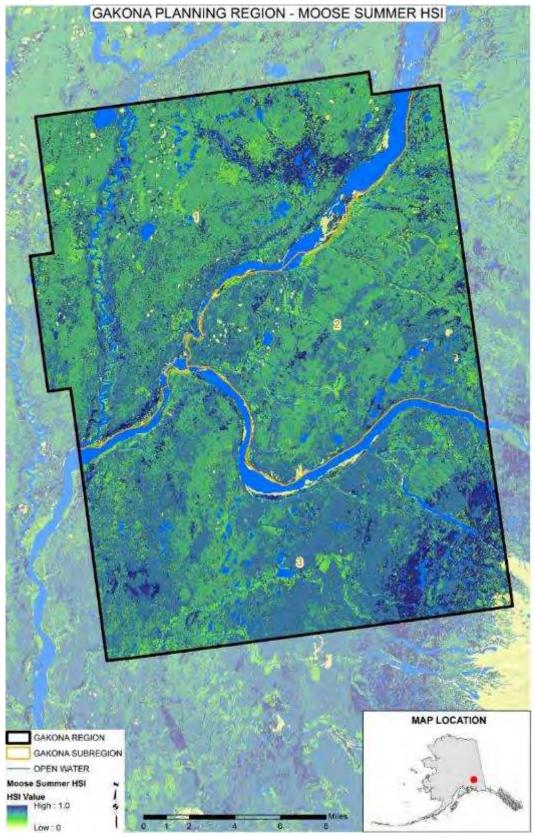


Figure 133. Habitat suitability for moose during summer in the Gakona Planning Region of the Ahtna Traditional Use Territory.

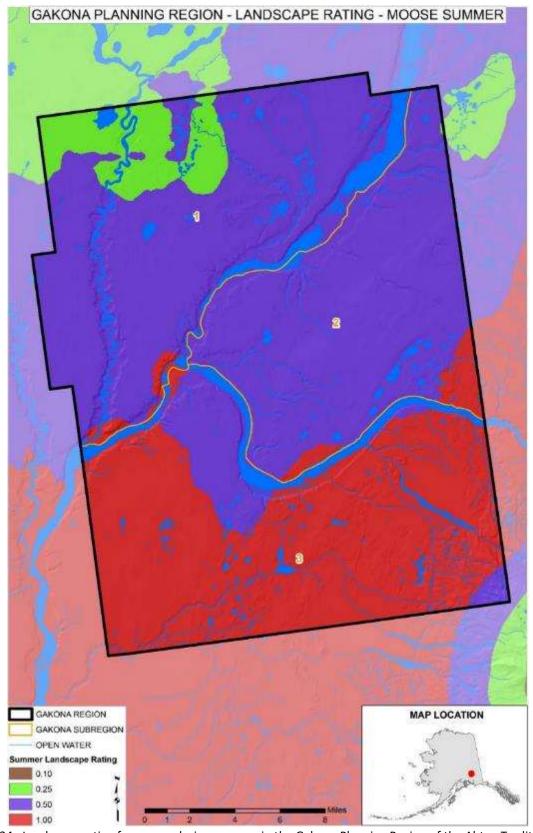


Figure 134. Landscape rating for moose during summer in the Gakona Planning Region of the Ahtna Traditional Use Territory.

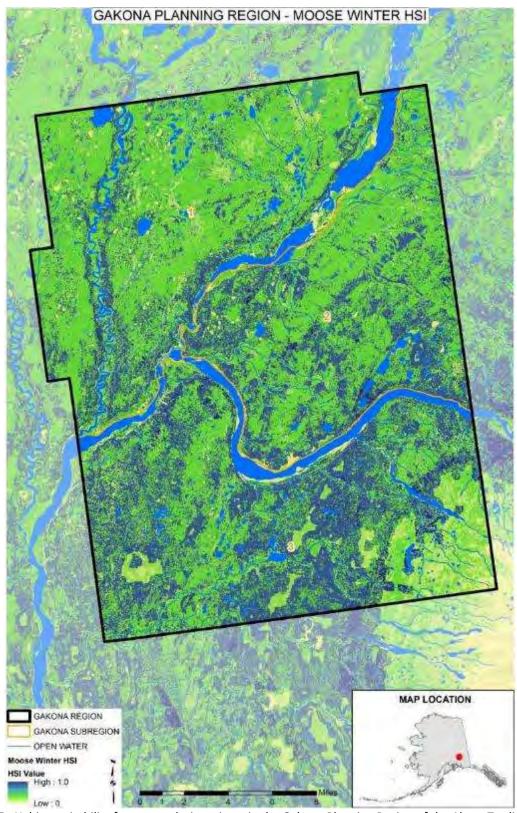


Figure 135. Habitat suitability for moose during winter in the Gakona Planning Region of the Ahtna Traditional Use Territory.

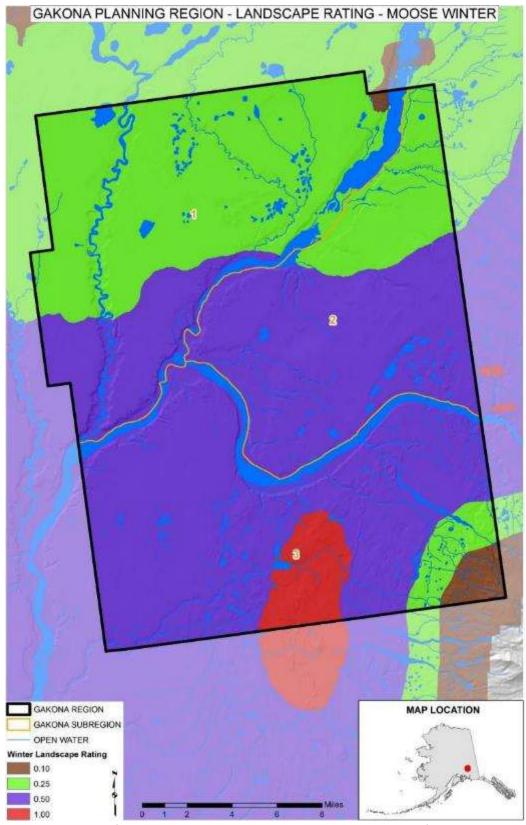


Figure 136. Landscape rating for moose during winter in the Gakona Planning Region of the Ahtna Traditional Use Territory.

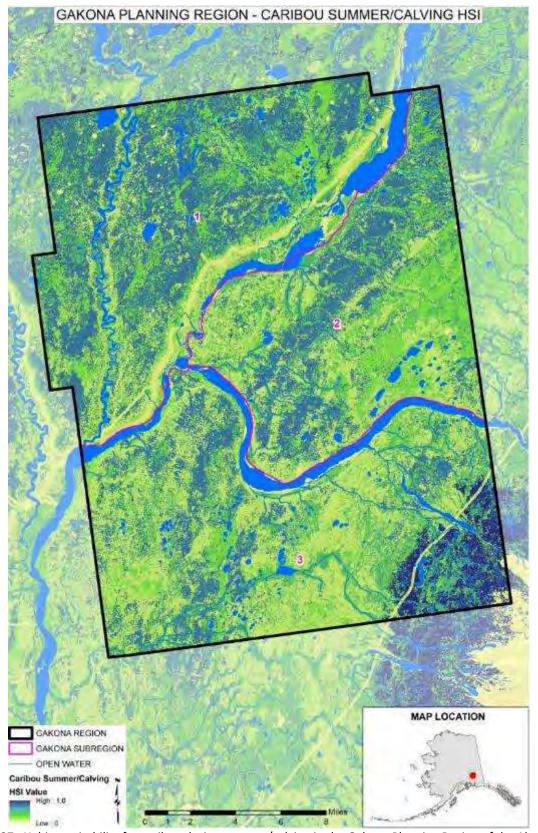


Figure 137. Habitat suitability for caribou during summer/calving in the Gakona Planning Region of the Ahtna Traditional Use Territory.

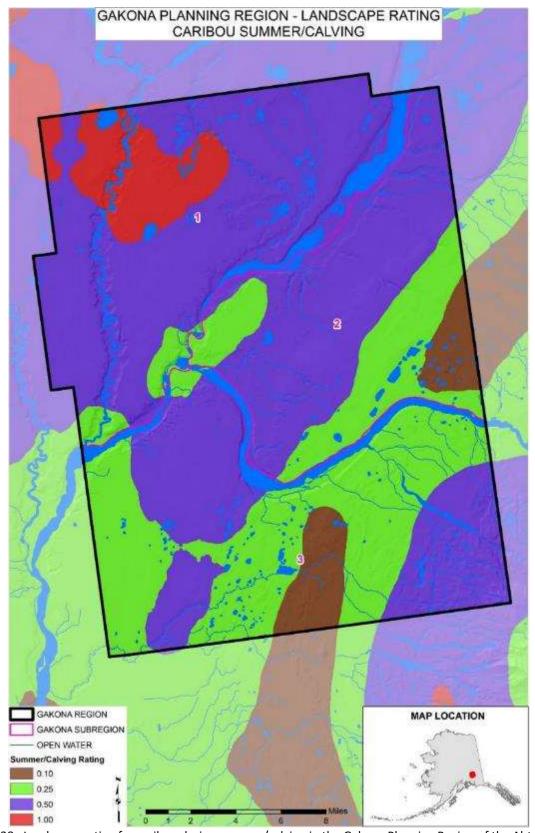


Figure 138. Landscape rating for caribou during summer/calving in the Gakona Planning Region of the Ahtna Traditional Use Territory.

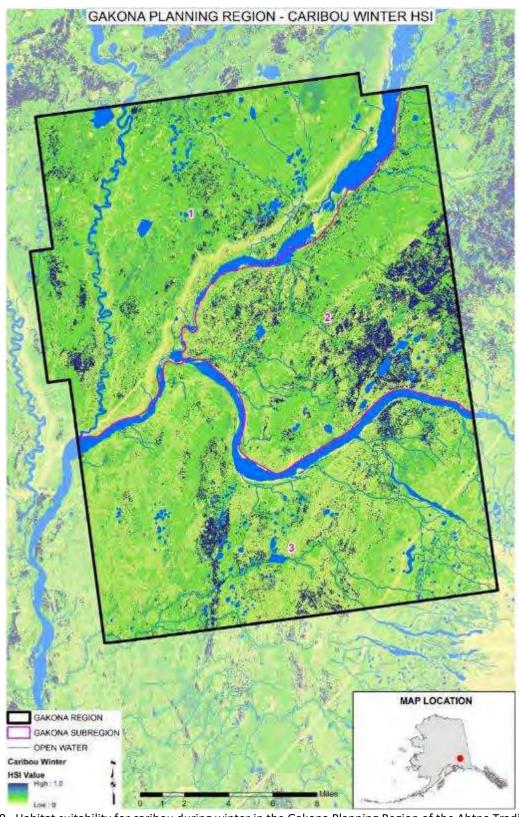


Figure 139. Habitat suitability for caribou during winter in the Gakona Planning Region of the Ahtna Traditional Use Territory.

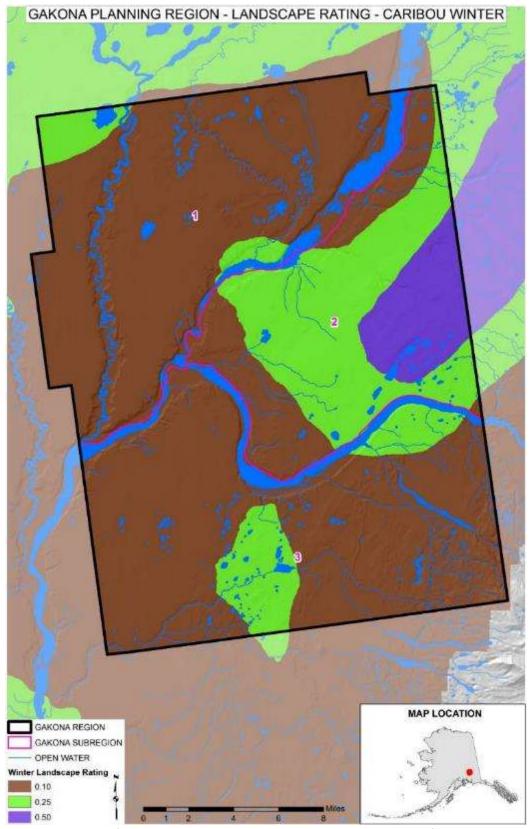


Figure 140. Landscape rating for caribou during winter in the Gakona Planning Region of the Ahtna Traditional Use Territory.

Gakona Site Improvement Areas

Potential treatment sites identified in the Gakona area are displayed in figure 141. Based on all of the stated criteria, a number of sites were located for designated treatments. Figures 142-147 depict the locations of these areas. Appendix C provides a description of each of these sites. Sites ranging from approximately 9-39 acres in size were identified and are listed in Table 26. This table gives the site size and the treatment goal of primarily timber stand improvement and biomass production or moose browse enhancement. The table also displays an estimate of the total biomass of all standing above ground biomass. This estimate is not tonnage of usable biomass for energy but the total of all stems regardless of size. However, stands on the higher end of the total amount will typically have larger trees and are a candidate for biomass energy production.

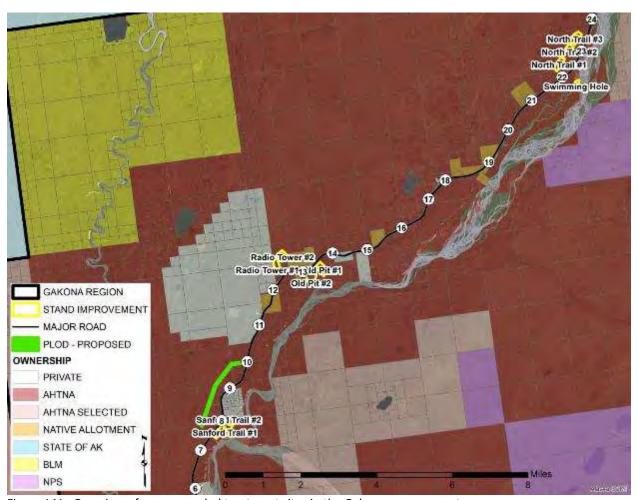


Figure 141. Overview of recommended treatment sites in the Gakona management area.

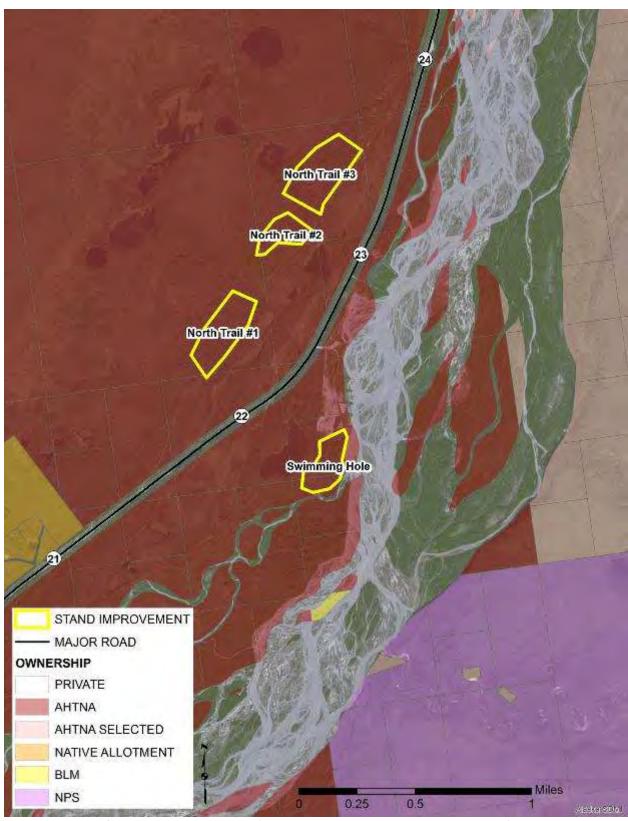


Figure 142. Map of proposed treatment areas (Swimming Hole and North Trail #1,#2, and #3) in the Gakona Planning Region showing surface ownership and aerial imagery.

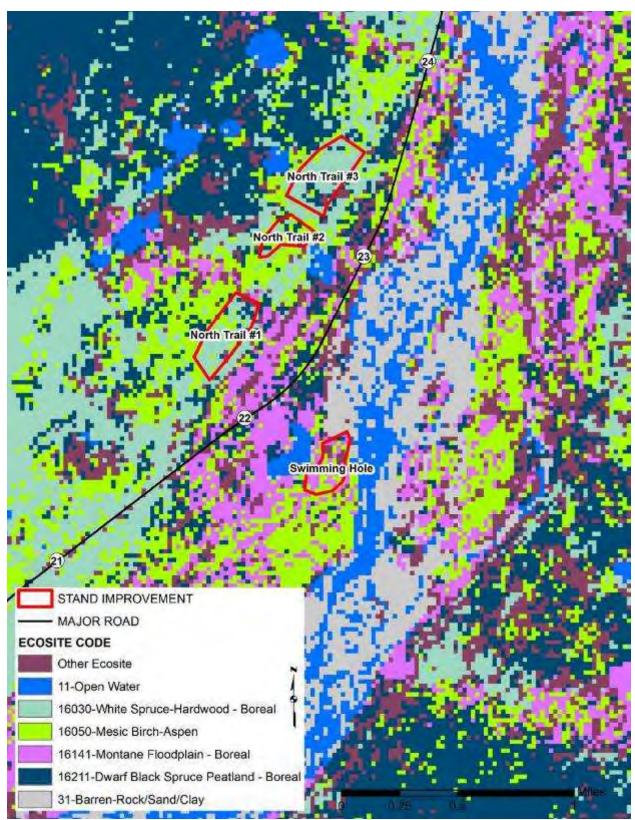


Figure 143. Map of proposed treatment areas (Swimming Hole and North Trail #1,#2, and #3) in the Gakona Planning Region showing ecological sites.



Figure 144. Map of proposed habitat improvement areas (Radio Tower #1 and #2 and Old Pit #1 and #2) in the Gakona Planning Region showing surface ownership and aerial imagery.

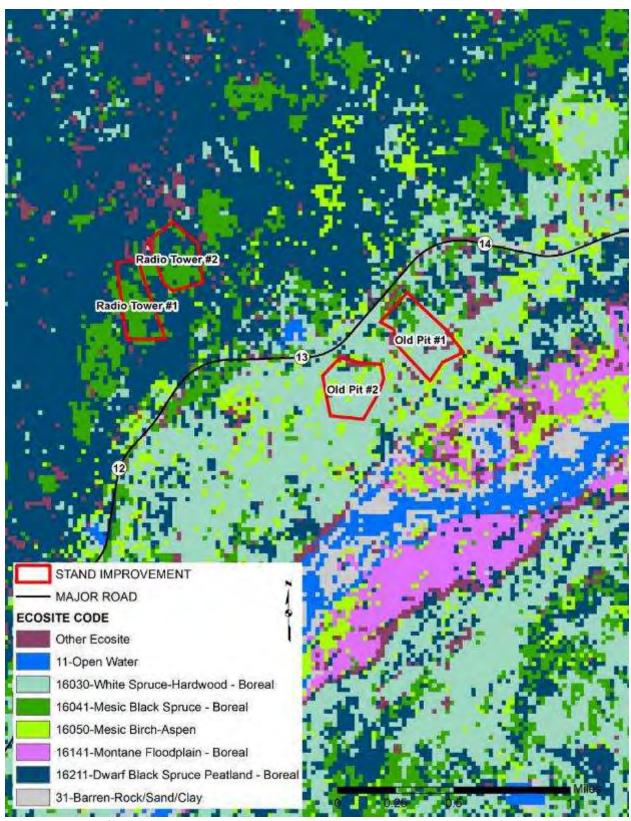


Figure 145. Map of proposed habitat improvement areas (Radio Tower #1 and #2 and Old Pit #1 and #2) in the Gakona Planning Region showing ecological sites.



Figure 146. Map of proposed habitat improvement areas (Sanford Trail #1, #2, and #3) and proposed PLOD in the Gakona Planning Region showing surface ownership and aerial imagery.

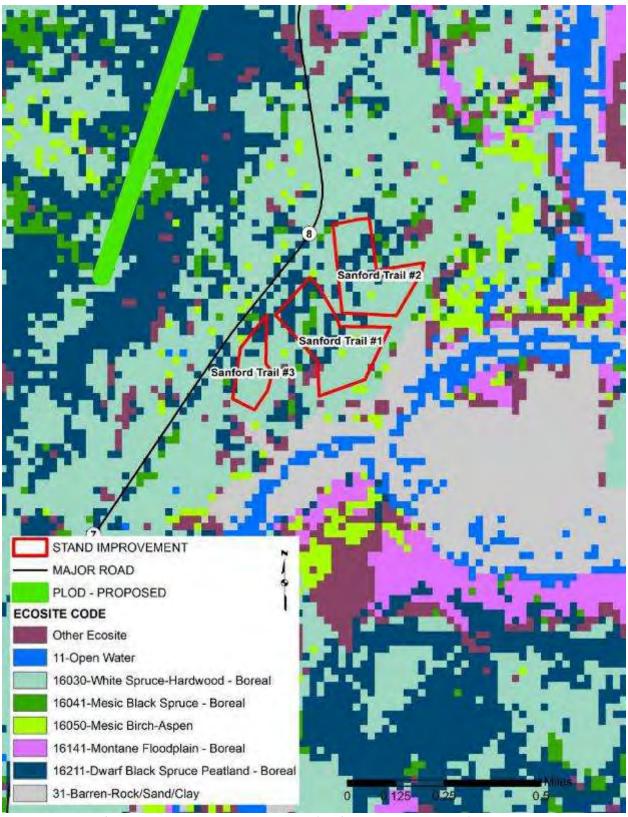


Figure 147. Map of proposed habitat improvement areas (Sanford Trail #1, #2, and #3) and proposed PLOD in the Gakona Planning Region showing ecological sites.

Table 26. Treatment site with treatment goal, site size, total tons of biomass and the BpS code and disturbance class for each site in the Gakona Planning Region.

Site Name	Treatment Goal	Acres	Biomass - Total	BpS Code	
Site Name	Treatment Goal	Acres	(tons)	_Dist Code	
North Trail #1	Timber Imp/Browse	30.6	845.5	16030_C	
North Trail #2	Timber Imp/Browse	14.2	672.4	16030_C	
North Trail #3	Timber Imp/Browse	36.0	626.4	16030_C	
Old Pit #1	Timber Improvement	38.9	1270.2	16030_C	
Old Pit #2	Timber Improvement	33.1	269.4	16030_C	
Radio Tower #1	Moose Browse	27.9	404.2	16041_D	
Radio Tower #2	Moose Browse	29.0	445.2	16041_D	
Sanford Trail #1	Timber Improvement	25.3	411.6	16030_C	
Sanford Trail #2	Timber Improvement	21.8	435.4	16030_C	
Sanford Trail #3	Timber Improvement	9.3	138.7	16030_C	
Swimming Hole	Moose Browse	19.6	515.8	16141_A	

Gulkana Planning Region

The Gulkana Planning Region includes an area of 270,781 acres. Figure 148 displays the overall planning area along with associated infrastructure. Ownership of this and the surrounding area is shown in Figure 149. As Figure 149 displays, land ownership patterns are varied in this area with Ahtna owning 43.5% (117,806 acres) of the land.

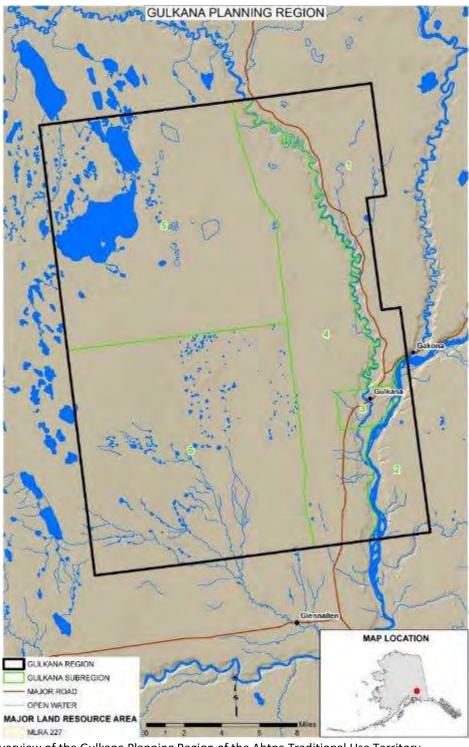


Figure 148. Overview of the Gulkana Planning Region of the Ahtna Traditional Use Territory.

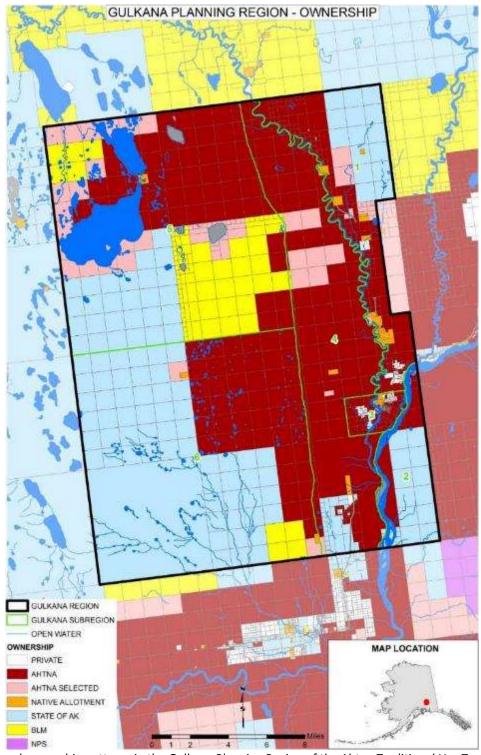


Figure 149. Land ownership patterns in the Gulkana Planning Region of the Ahtna Traditional Use Territory.

A description of the general geology, climate, soils, permafrost, and vegetation is found in Chapter 1 of this report. Figures showing these features specific to the Gulkana Planning Region are displayed below. Soil texture in the Gulkana area is shown in Figure 150 and Figure 151 displays soil drainages. Permafrost in the Gulkana area is shown in Figure 152.

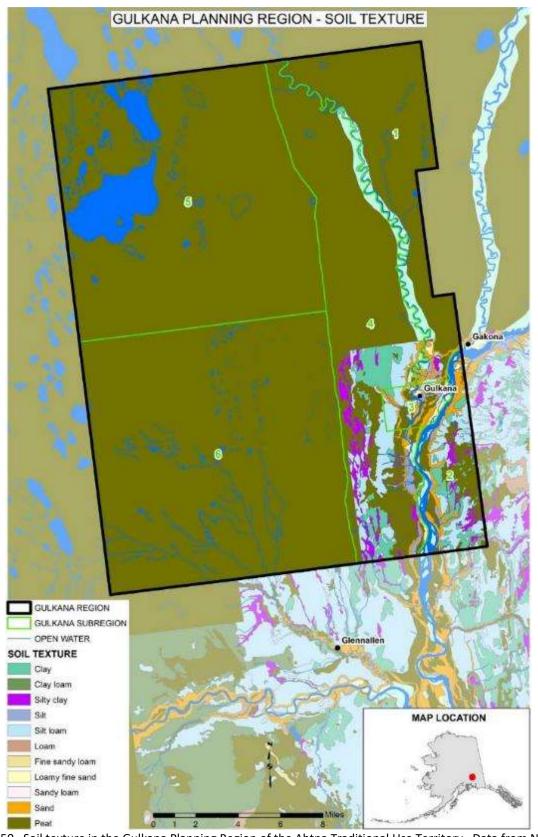


Figure 150. Soil texture in the Gulkana Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

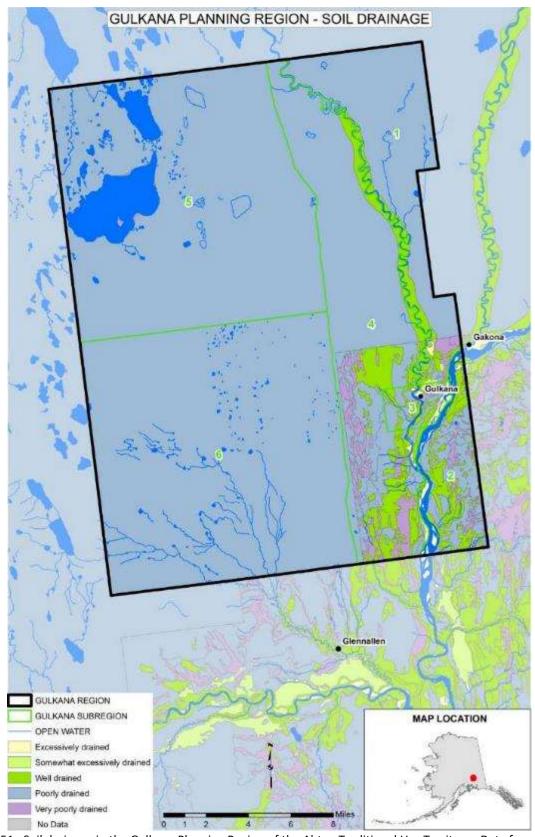


Figure 151. Soil drainage in the Gulkana Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

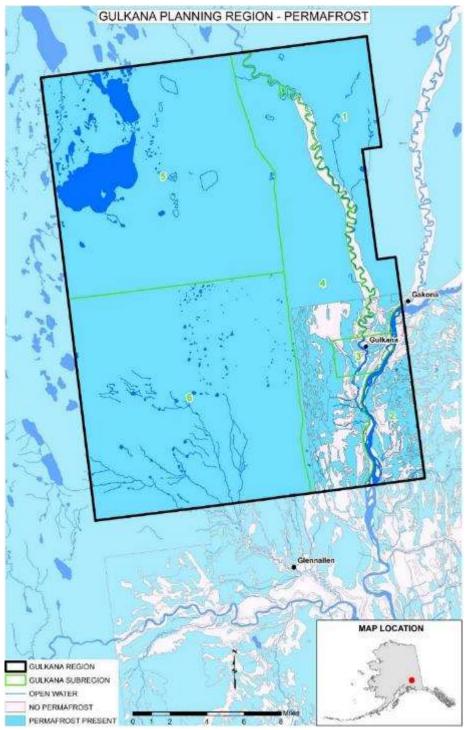


Figure 152. Permafrost in the Gulkana Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

Disturbance Factors

Fire is the primary disturbance factor influencing the vegetation ecology in the Copper Basin. Figure 153 shows the fire history of the Gulkana area along with current fire protection zones. Although the level of fire occurring in this southcentral Alaska landscape is substantially less than that occurring in more

interior areas of Alaska north of the Alaska Range, fire is still a significant disturbance when it occurs. However, there has not been a large fire in the project area since 1940. When fire does occur it serves to set back succession. It can also burn off the organic material at the ground surface, including peat that can occur on many sites. This can influence the thermal layer protecting the underlying permafrost on some sites, causing the permafrost to melt (thermokarst) and changing the site conditions through this process. Riparian areas are also influenced by flooding and ice events. These serve to set back succession of vegetation in riparian areas, and can even shift site conditions, particularly in the case of significant flooding events.

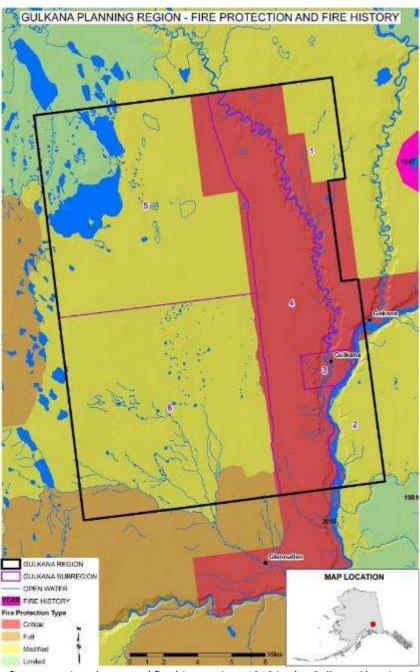


Figure 153. Current fire protection classes and fire history since 1940 in the Gulkana Planning Region of the Ahtna Traditional Use Territory. Data from Alaska Interagency Coordination Center.

Ecosystem Diversity

The biophysical settings present within the Gulkana Planning region are displayed in Figure 154. Table 27 displays the acres for each biophysical setting and disturbance class. Figure 155 is a map of ecosystem diversity (represented by biophysical setting and disturbance class) in the Gulkana Planning Region.

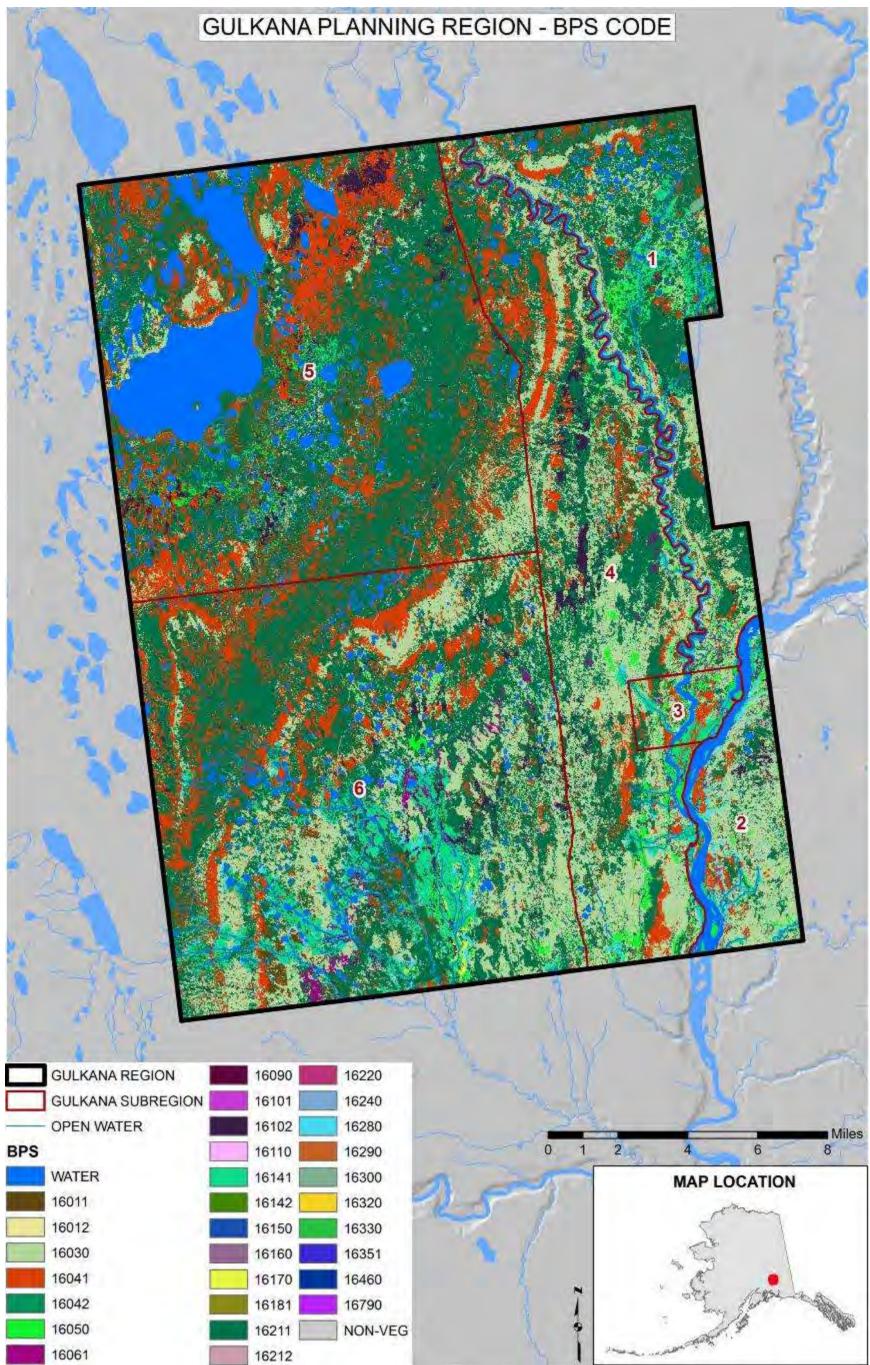


Figure 154. Biophysical settings (codes) in the Gulkana Planning Region of the Ahtna Traditional Use Territory. Data from LANDFIRE.

Table 27. Acres by BpS code and disturbance class (A – E) in the Gulkana Planning Region. The BpS vegetation label is provided as well.

pS Code Dist Class	BpS Vegetation Label	Acres	BpS Code _Dist Class	BpS Vegetation Label	
11	Open Water	17174.0	16150_D	Large River Floodplain	
.6011_A	Treeline White Spruce-Boreal	122.8	16160_A	Riparian Stringer	
16011_B	Treeline White Spruce-Boreal	1962.9	16160_B	Riparian Stringer	
16011_C	Treeline White Spruce-Boreal	0.2	16170_A	Shrub and Herbaceous Floodplain	
16012_A	Treeline White Spruce-SubBoreal	0.4	16170_B	Shrub and Herbaceous Floodplain	
16012_B	Treeline White Spruce-SubBoreal	11.8	16170_C	Shrub and Herbaceous Floodplain	
16030_A	White Spruce-Hardwood-Boreal	2690.1	16170_E	Shrub and Herbaceous Floodplain	
16030_B	White Spruce-Hardwood-Boreal	4325.6	16181_A	Herbaceous Fen	
16030_C	White Spruce-Hardwood-Boreal	49907.4	16181_B	Herbaceous Fen	
16030_E	White Spruce-Hardwood-Boreal	102.3	16181_C	Herbaceous Fen	
16041_A	Mesic Black Spruce-Boreal	887.4	16181_D	Herbaceous Fen	
16041_B	Mesic Black Spruce-Boreal	2016.9	16211_A	Dwarf Black Spruce Peatland-Boreal	
16041_C	Mesic Black Spruce-Boreal	829.5	16211_B	Dwarf Black Spruce Peatland-Boreal	
16041_D	Mesic Black Spruce-Boreal	40404.1	16211_C	Dwarf Black Spruce Peatland-Boreal	
16041_E	Mesic Black Spruce-Boreal	9.6	16211_D	Dwarf Black Spruce Peatland-Boreal	1
16042_A	Mesic Black Spruce-SubBoreal	2.2	16212_A	Dwarf Black Spruce Peatland-Subboreal	
16042_B	Mesic Black Spruce-SubBoreal	37.1	16212_B	Dwarf Black Spruce Peatland-Subboreal	
16042_C	Mesic Black Spruce-SubBoreal	2.2	16212_C	Dwarf Black Spruce Peatland-Subboreal	
16050_A	Mesic Birch-Aspen	4527.7	16220_A	Black Spruce Wet-Mesic Slope	
16050_B	Mesic Birch-Aspen	706.3	16220_B	Black Spruce Wet-Mesic Slope	
16050_D	Mesic Birch-Aspen	7.8	16220_C	Black Spruce Wet-Mesic Slope	
16050_E	Mesic Birch-Aspen	19.1	16220_D	Black Spruce Wet-Mesic Slope	
16061_A	Dry Aspen-Steppe Bluff	62.5	16240_A	Deciduous Shrub Swamp	
16061_B	Dry Aspen-Steppe Bluff	181.9	16280_A	Low Shrub-Tussock Tundra	
16061_C	Dry Aspen-Steppe Bluff	75.4	16280_B	Low Shrub-Tussock Tundra	
16061_D	Dry Aspen-Steppe Bluff	1454.2	16280_C	Low Shrub-Tussock Tundra	
16090_A	Mesic Subalpine Alder	42.3	16290_A	Tussock Tundra	
16090_B	Mesic Subalpine Alder	11.3	16300_A	Wet Black Spruce-Tussock	
16102_A	Mesic Scrub Birch-Willow	2845.5	16300_B	Wet Black Spruce-Tussock	
16102_B	Mesic Scrub Birch-Willow	4893.6	16300_C	Wet Black Spruce-Tussock	
16110_A	Mesic Bluejoint Meadow	39.6	16320_A	Alpine Talus and Bedrock	
16141_A	Montane Floodplain-Boreal	3936.2	16320_B	Alpine Talus and Bedrock	
16141_B	Montane Floodplain-Boreal	920.3	16330_A	Alpine Mesic Herbaceous Meadow	
16141_C	Montane Floodplain-Boreal	808.0	16351_A	Alpine Ericaceous Dwarf-Shrubland	
16141_D	Montane Floodplain-Boreal	6162.1	16790_A	White Spruce-Hardwood-SubBoreal	
16141_E	Montane Floodplain-Boreal	721.2	16790_B	White Spruce-Hardwood-SubBoreal	
16150_A	Large River Floodplain	120.3	16790_C	White Spruce-Hardwood-SubBoreal	
16150_B	Large River Floodplain	14.2	31	Barren-Rock-Sand-Clay	
16150_C	Large River Floodplain	31.8			

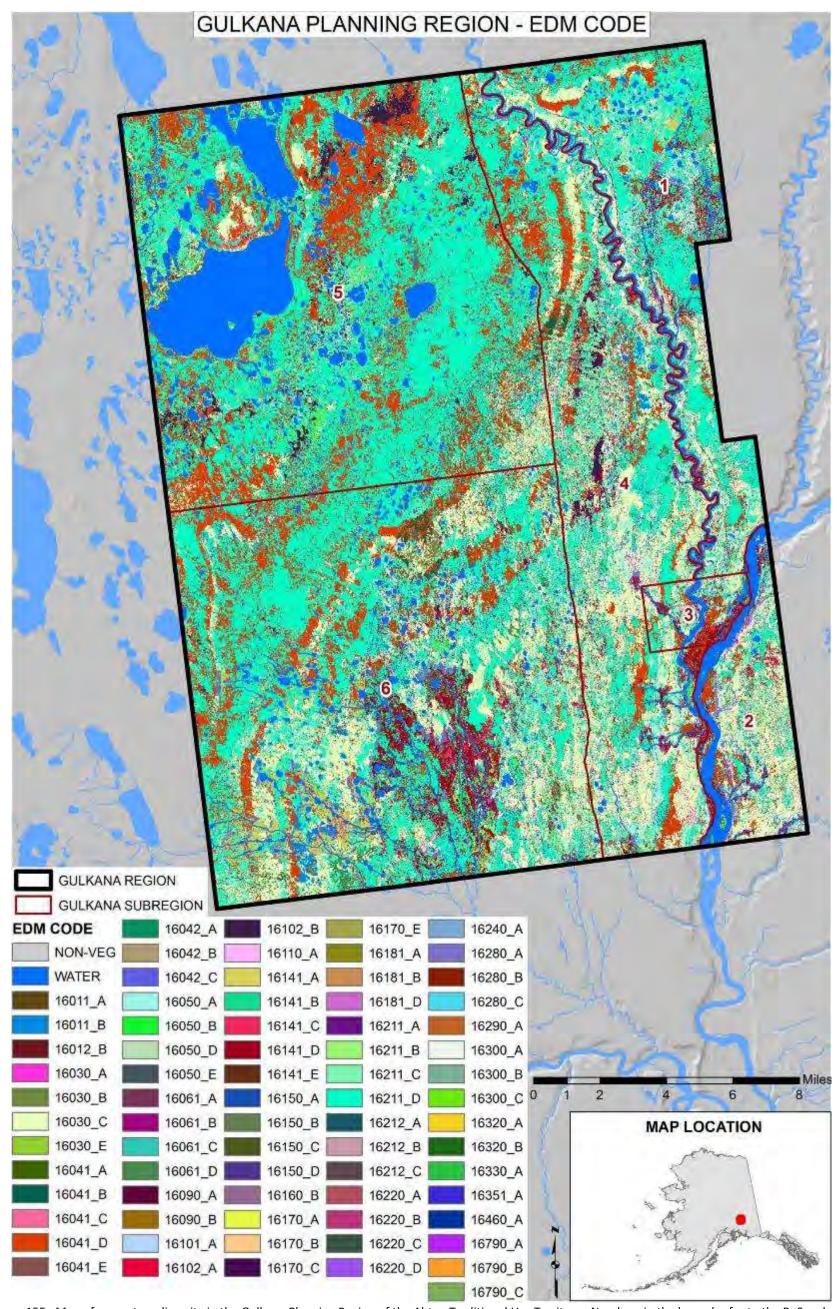


Figure 155. Map of ecosystem diversity in the Gulkana Planning Region of the Ahtna Traditional Use Territory. Numbers in the legend refer to the BpS code, while letters (A-E) refer to the disturbance class.

Berry Production Areas

Ecological sites that had the potential for producing desirable species of berries were identified. These sites were then further identified as to the disturbance states most likely to support berry production. While numerous factors can influence where berry-producing vegetation occurs, the combination of ecological sites and disturbance states allows for the identification of areas with higher probabilities of finding good berry production. Figure 156 shows the areas identified with the highest potential for berry production. It may be desirable to avoid disturbances to these areas, but ground-truthing of the actual presence of good berry production is recommended.

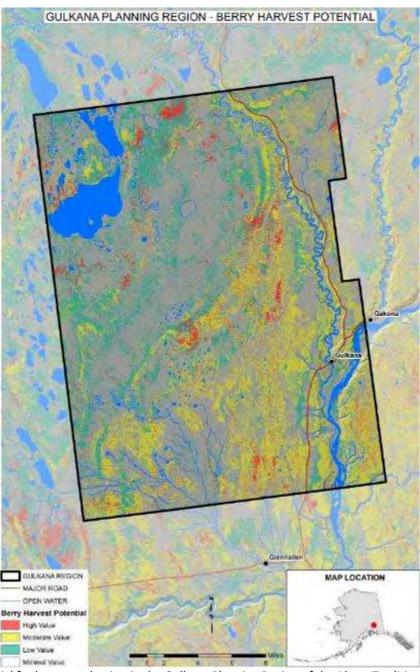


Figure 156. Potential for berry production in the Gulkana Planning Region of the Ahtna Traditional Use Territory based on ecological sites and disturbance states derived from remotely sensed information.

Moose and Caribou Model Results

The habitat suitability figure for moose spring habitat is found in Figure 157 and the landscape rating for moose spring habitat is found in Figure 158. The habitat suitability figure for moose summer habitat is found in Figure 159 and the landscape rating for moose summer habitat is found in Figure 160. The habitat suitability figure for moose winter habitat is found in Figure 161 and the landscape rating for moose winter habitat is found in Figure 162. The habitat suitability figure for caribou summer/calving habitat are found in Figure 163 and the landscape rating for caribou summer/calving habitat are found Figure 164. The habitat suitability figure for caribou winter habitat are found in Figure 165 and the landscape rating for caribou winter habitat are found Figure 166. A complete description of the moose habitat model can be found in Appendix A. A complete description of the caribou habitat model can be found in Appendix B.

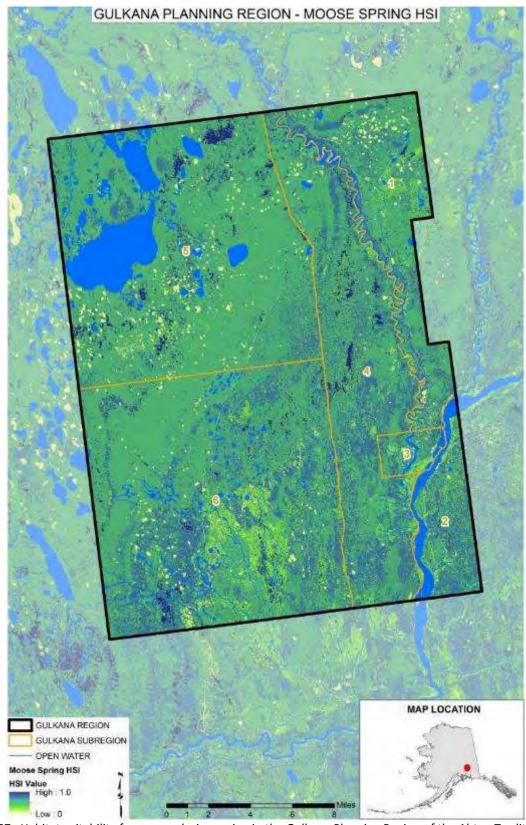


Figure 157. Habitat suitability for moose during spring in the Gulkana Planning Region of the Ahtna Traditional Use Territory.

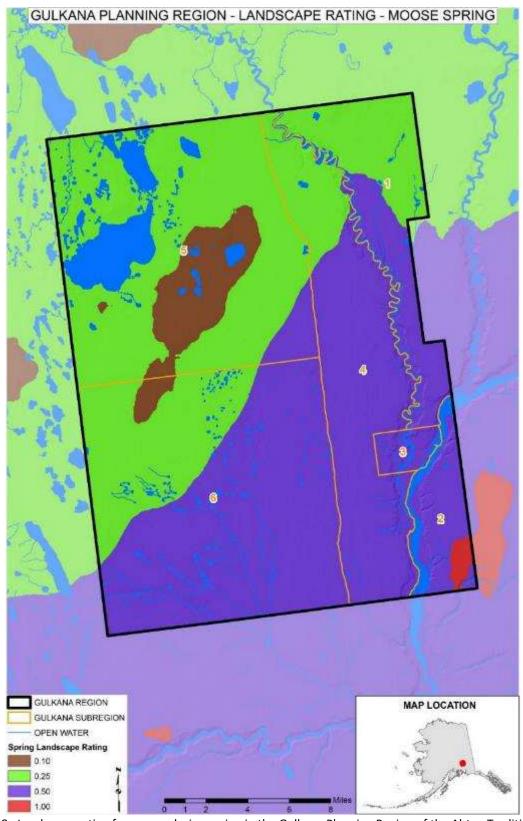


Figure 158. Landscape rating for moose during spring in the Gulkana Planning Region of the Ahtna Traditional Use Territory.

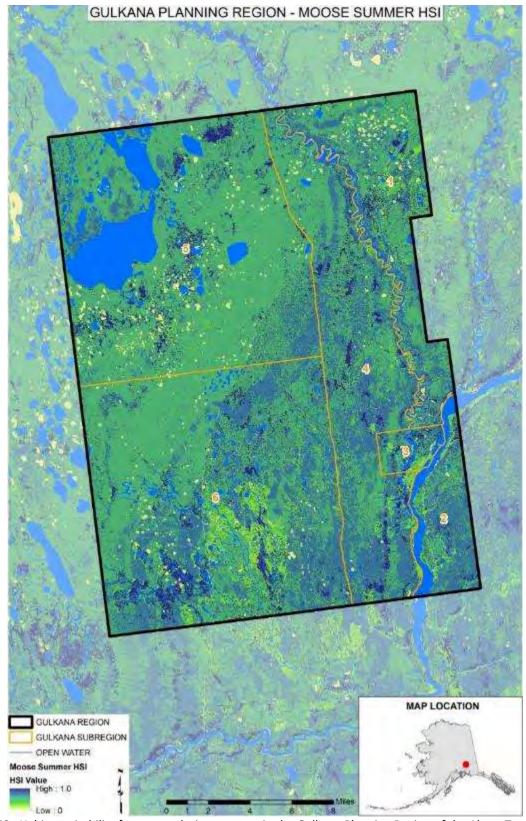


Figure 159. Habitat suitability for moose during summer in the Gulkana Planning Region of the Ahtna Traditional Use Territory.

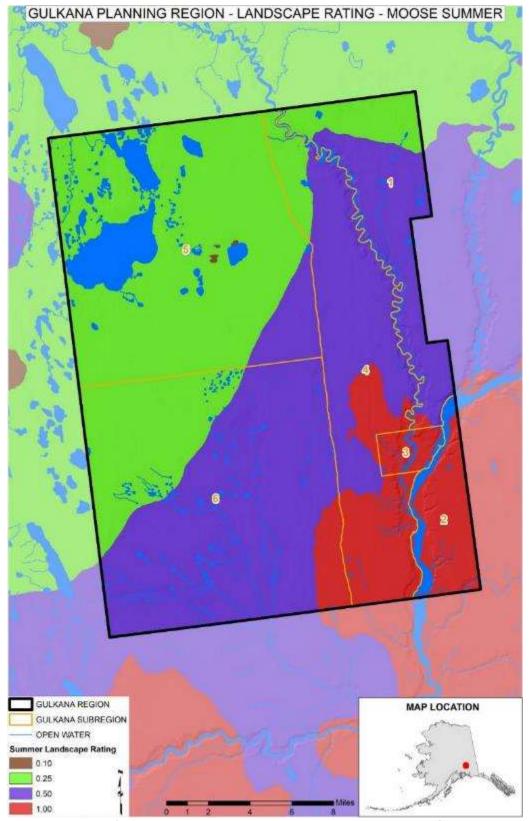


Figure 160. Landscape rating for moose during summer in the Gulkana Planning Region of the Ahtna Traditional Use Territory.

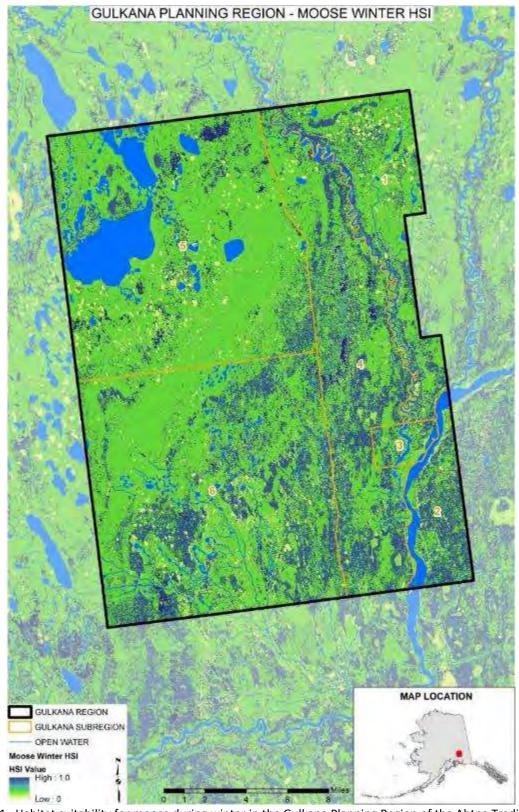


Figure 161. Habitat suitability for moose during winter in the Gulkana Planning Region of the Ahtna Traditional Use Territory.

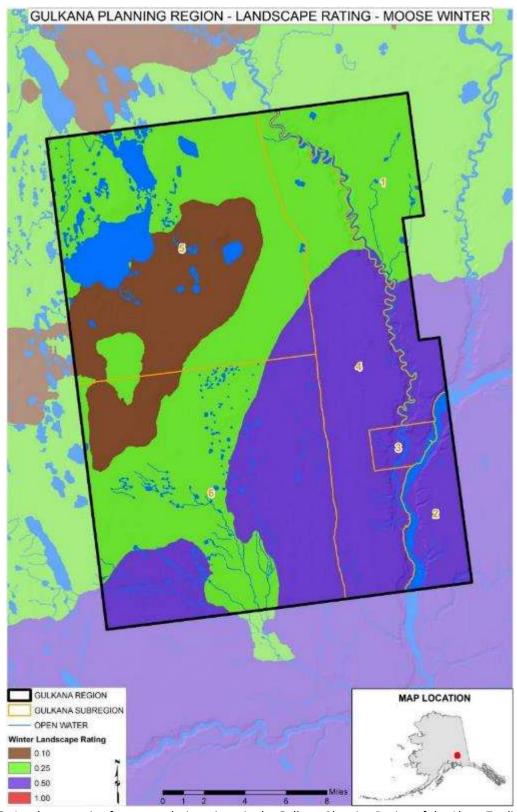


Figure 162. Landscape rating for moose during winter in the Gulkana Planning Region of the Ahtna Traditional Use Territory.

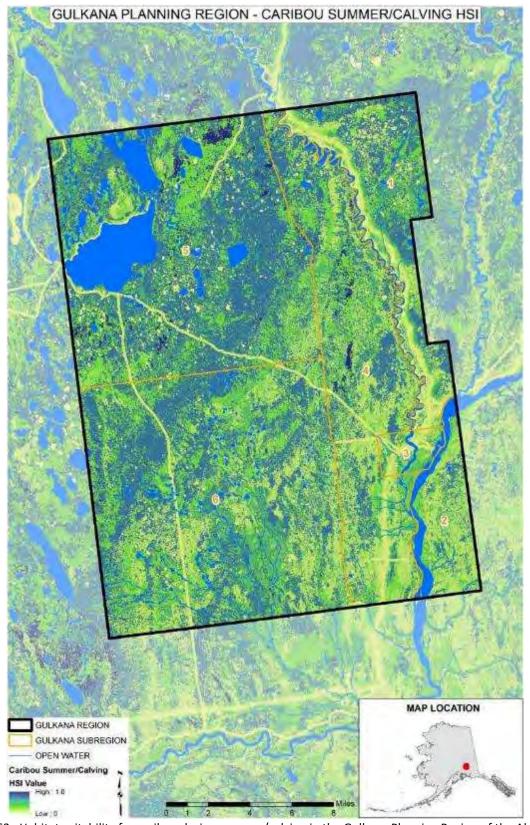


Figure 163. Habitat suitability for caribou during summer/calving in the Gulkana Planning Region of the Ahtna Traditional Use Territory.

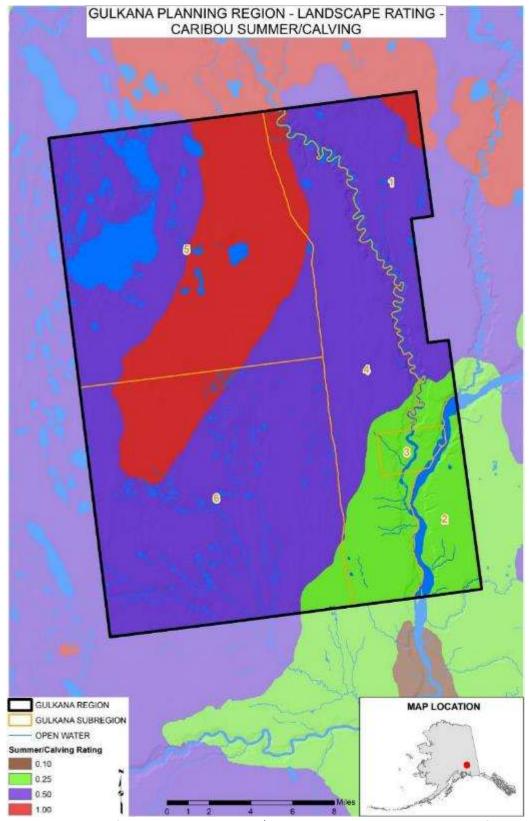


Figure 164. Landscape rating for caribou during summer/calving in the Gulkana Planning Region of the Ahtna Traditional Use Territory.

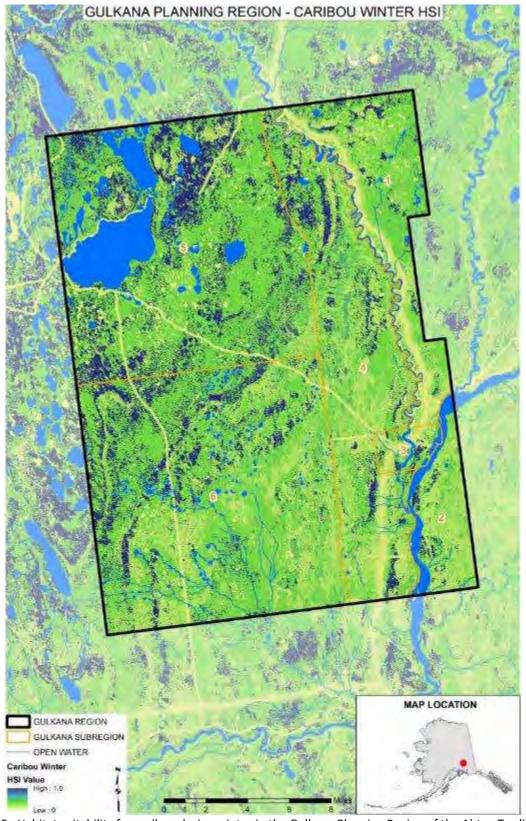


Figure 165. Habitat suitability for caribou during winter in the Gulkana Planning Region of the Ahtna Traditional Use Territory.

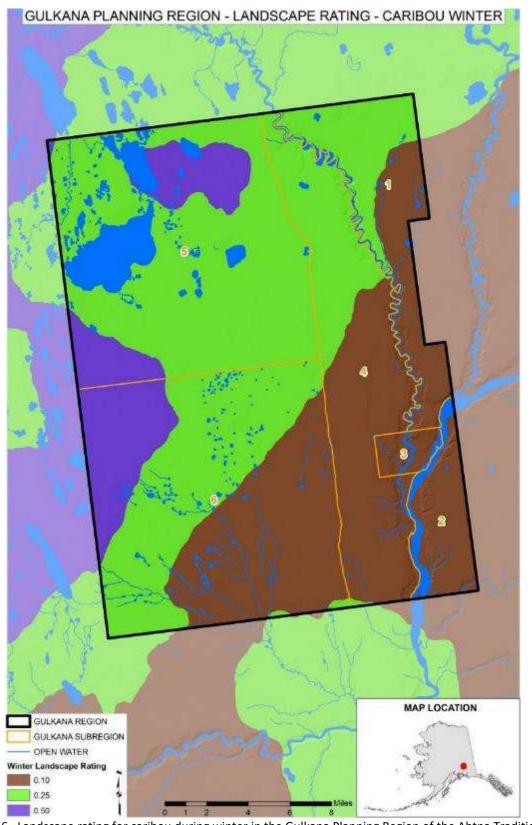


Figure 166. Landscape rating for caribou during winter in the Gulkana Planning Region of the Ahtna Traditional Use Territory.

Gulkana Site Improvement Areas

Potential treatment sites identified in the Gulkana area are displayed in figure 167. Based on all of the stated criteria, a number of sites were located for designated treatments. Figures 168-175 depict the locations of these areas. Appendix C provides a description of each of these sites. Sites ranging from approximately 9-53 acres in size were identified and are listed in Table 28. This table gives the site size and the treatment goal of primarily timber stand improvement and biomass production or moose browse enhancement. The table also displays an estimate of the total biomass of all standing above ground biomass. This estimate is not tonnage of usable biomass for energy but the total of all stems regardless of size. However, stands on the higher end of the total amount will typically have larger trees and are a candidate for biomass energy production.

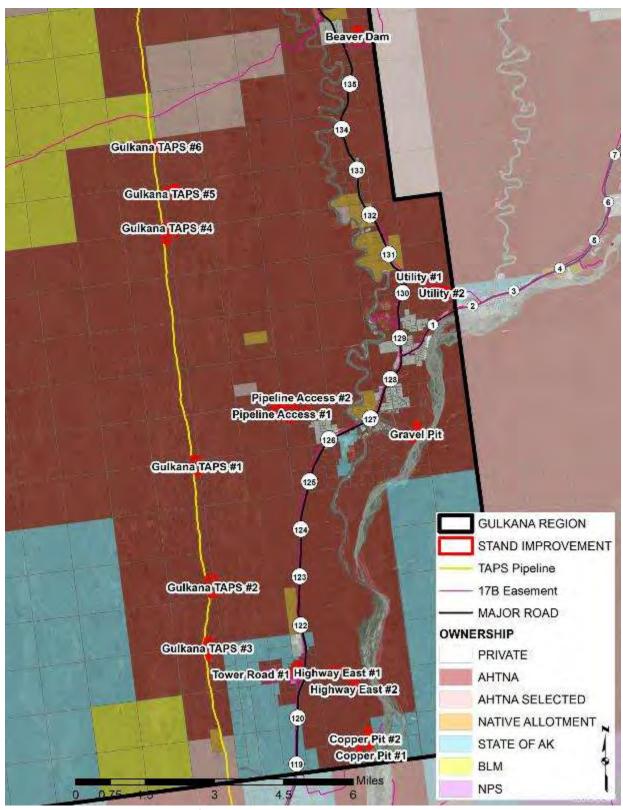


Figure 167. Overview of recommended treatment sites in the Gulkana management area.



Figure 168. Map of proposed treatment areas (Utility #1 and #2, Gravel Pit, and Pipeline Access #1 and #2) in the Gulkana Planning Region showing surface ownership and aerial imagery.

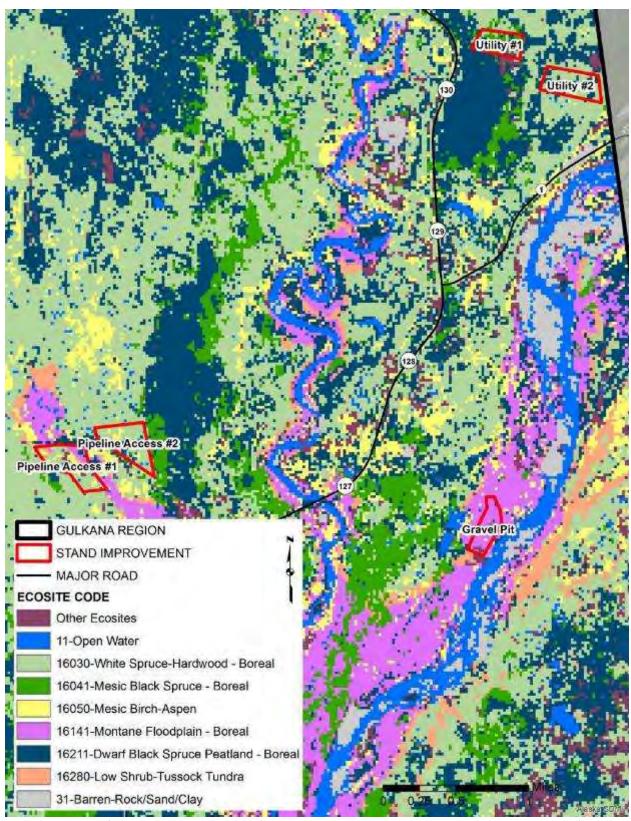


Figure 169. Map of proposed treatment areas (Utility #1 and #2, Gravel Pit, and Pipeline Access #1 and #2) in the Gulkana Planning Region showing ecological sites.

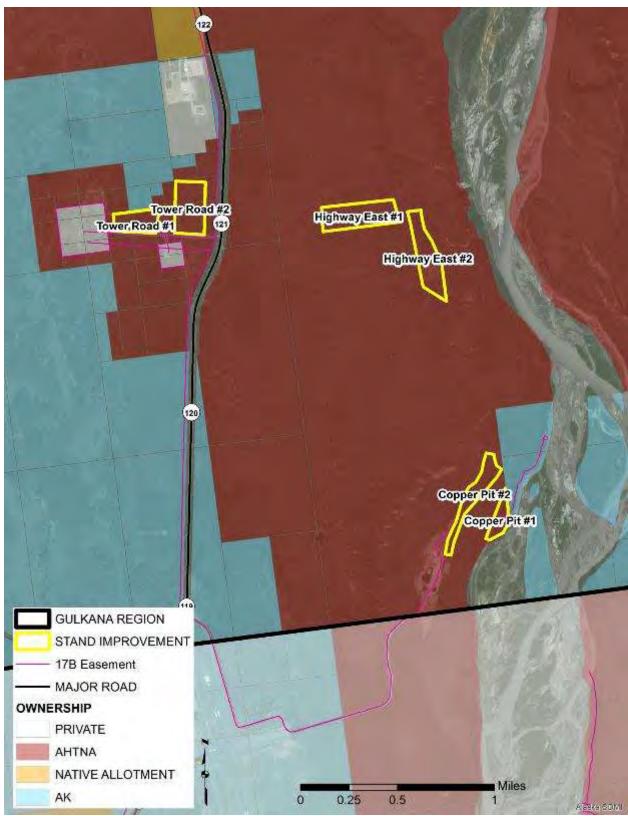


Figure 170. Map of proposed habitat improvement areas (Tower Road #1 and #2, Copper Pit #1 and #2, and Highway East #1 an #2) in the Gulkana Planning Region showing surface ownership and aerial imagery.

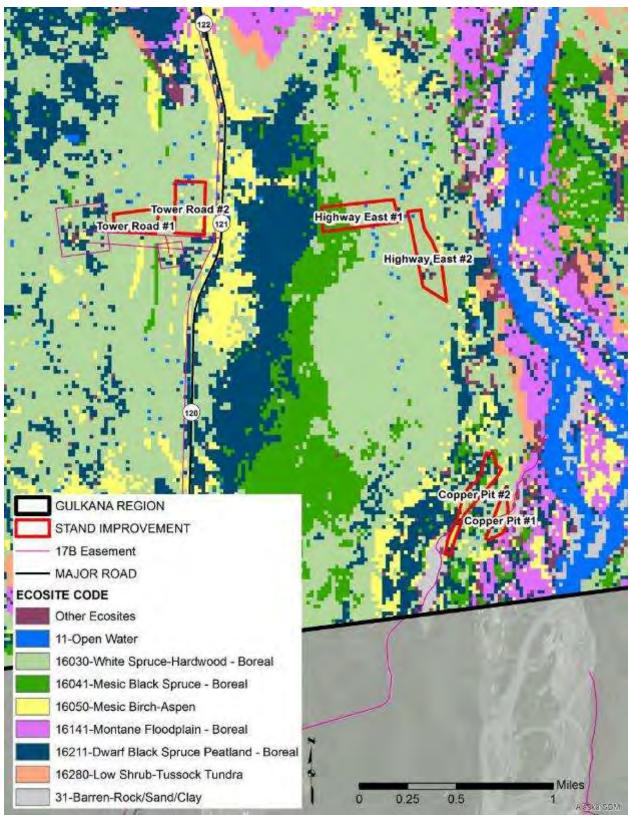


Figure 171. Map of proposed habitat improvement areas (Tower Road #1 and #2, Copper Pit #1 and #2, and Highway East #1 an #2) in the Gulkana Planning Region showing ecological sites.

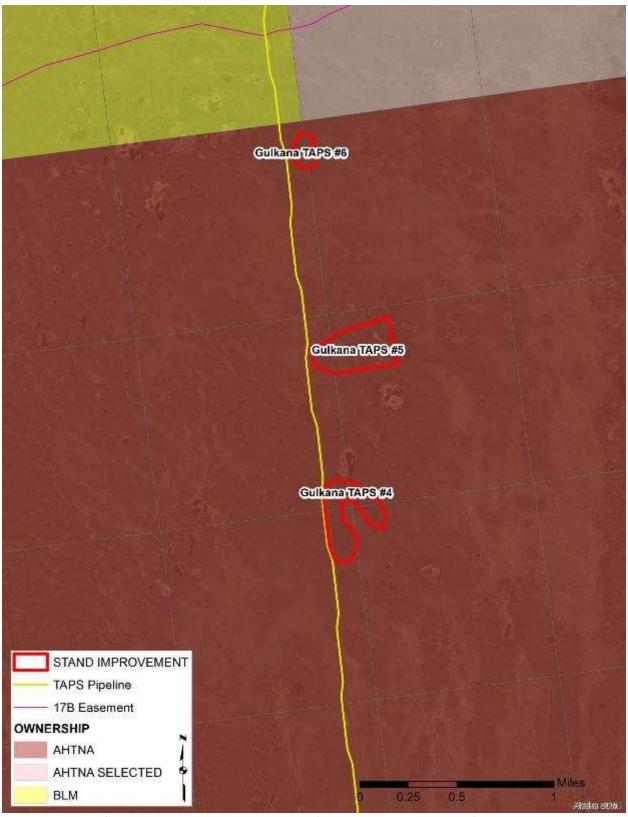


Figure 172. Map of proposed habitat improvement areas (Gulkana TAPS #4, #5, and #6) in the Gulkana Planning Region showing surface ownership and aerial imagery.

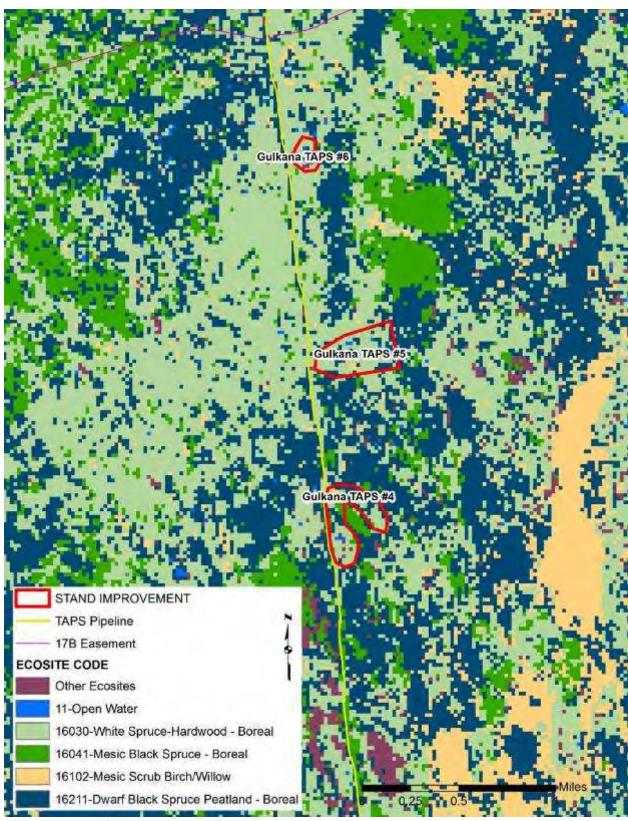


Figure 173. Map of proposed habitat improvement areas (Gulkana TAPS #4, #5, and #6) in the Gulkana Planning Region showing ecological sites.



Figure 174. Map of proposed habitat improvement areas (Gulkana TAPS #1, #1, and #3) in the Gulkana Planning Region showing surface ownership and aerial imagery.

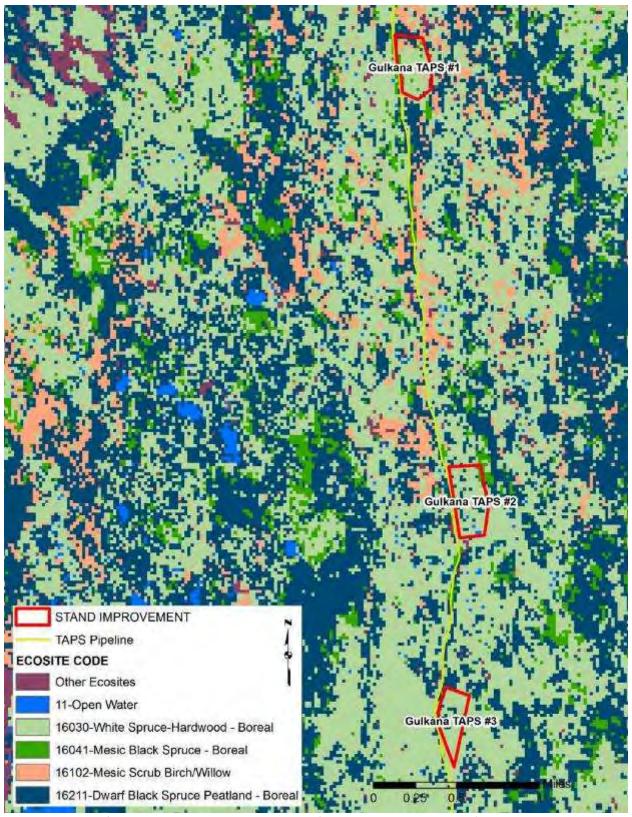


Figure 175. Map of proposed habitat improvement areas (Gulkana TAPS #1, #2, and #3) in the Gulkana Planning Region showing ecological sites.

Table 28. Treatment site with treatment goal, site size, total tons of biomass and the BpS code and disturbance class for each site in the Gulkana Planning Region.

Site Name	Treatment Goal	Acres	Biomass - Total	BpS Code
Site Name			(tons)	_Dist Code
Beaver Dam	Moose Browse	23.1	857.2	16030_C
Copper Pit #1	Moose Browse	8.8	299.3	16030_C
Copper Pit #2	Moose Browse	19.5	454.2	16030_C
Gulkana Gravel Pit	Biomass/Browse	30.3	78.9	16141_A
Gulkana TAPS #1	Moose Browse	39.9	1645.6	16030_C
Gulkana TAPS #2	Moose Browse	51.1	1922.6	16030_C
Gulkana TAPS #3	Moose Browse	32.6	1258.2	16030_C
Gulkana TAPS #4	Moose Browse	45.2	1372.1	16030_C
Gulkana TAPS #5	Moose Browse	53.4	1467.0	16030_C
Gulkana TAPS #6	Moose Browse	10.9	396.0	16030_C
Highway East #1		30.3	1355.7	16030_C
Highway East #2		26.6	1034.3	16030_C
Pipeline Access #1	Biomass/Browse	50.1	1562.1	16030_C
Pipeline Access #2	Biomass/Browse	48.7	1751.3	16030_C
Tower Road #1	Biomass/Browse	26.5	1044.0	16030_C
Tower Road #2	Biomass/Browse	15.5	625.9	16030_C

Kluti-Kaah Planning Region

The Kluti-Kaah Planning Region encompasses an area of 390,361 acres. Figure 176 displays the overall planning area along with associated infrastructure. Ownership of this and the surrounding area is shown in Figure 177. As Figure 177 displays, land ownership patterns are varied in this area with Ahtna, Inc. owning 59.4% (231,824 acres) of the land.

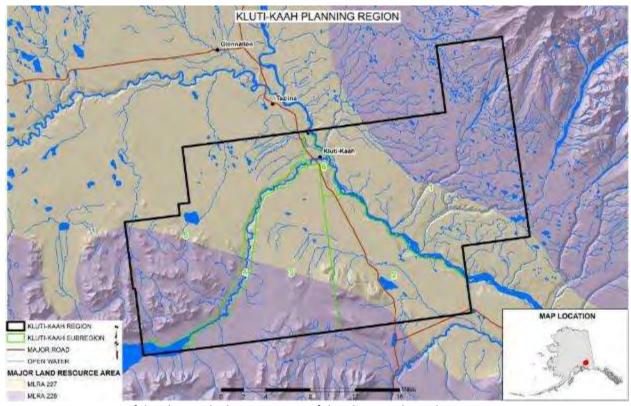


Figure 176. Overview of the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory.

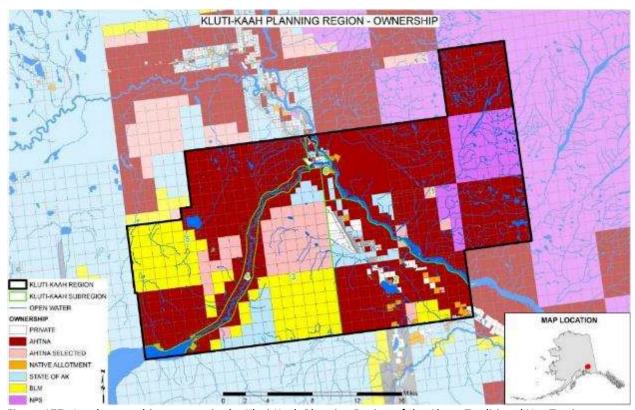


Figure 177. Land ownership patterns in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory.

A description of the general geology, climate, soils, permafrost, and vegetation is found in Chapter 1 of this report. Figures showing these features specific to the Kluti-Kaah Planning Region are displayed below. Soil texture in the Kluti-Kaah area is shown in Figure 178 and Figure 179 displays soil drainages. Permafrost in the Kluti-Kaah area is shown in Figure 180.

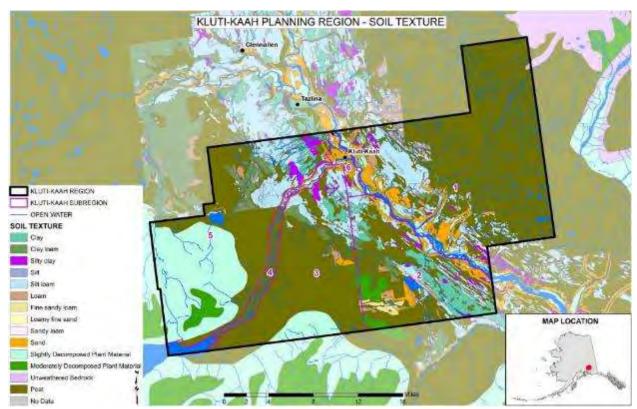


Figure 178. Soil texture in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

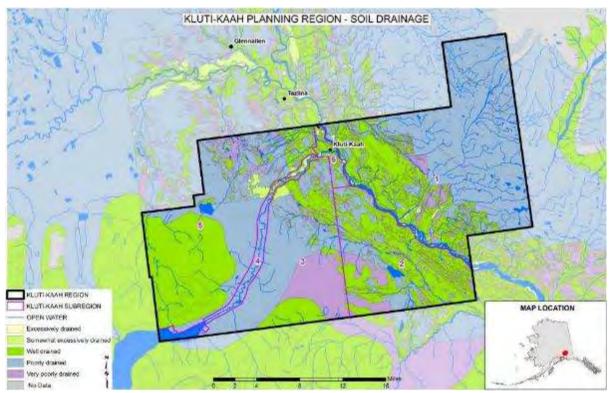


Figure 179. Soil drainage in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

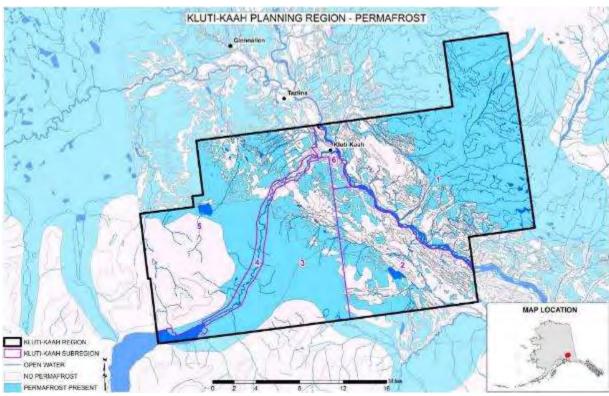


Figure 180. Permafrost in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

Disturbance Factors

Fire is the primary disturbance factor influencing the vegetation ecology in the Copper Basin. Figure 181 shows the fire history of the Kluti-Kaah area along with current fire protection zones. Although the level of fire occurring in this southcentral Alaska landscape is substantially less than that occurring in more interior areas of Alaska north of the Alaska Range, fire is still a significant disturbance when it occurs. When fire does occur it serves to set back succession. It can also burn off the organic material at the ground surface, including peat that can occur on many sites. This can influence the thermal layer protecting the underlying permafrost on some sites, causing the permafrost to melt (thermokarst) and changing the site conditions through this process. Riparian areas are also influenced by flooding and ice events. These serve to set back succession of vegetation in riparian areas, and can even shift site conditions, particularly in the case of significant flooding events.

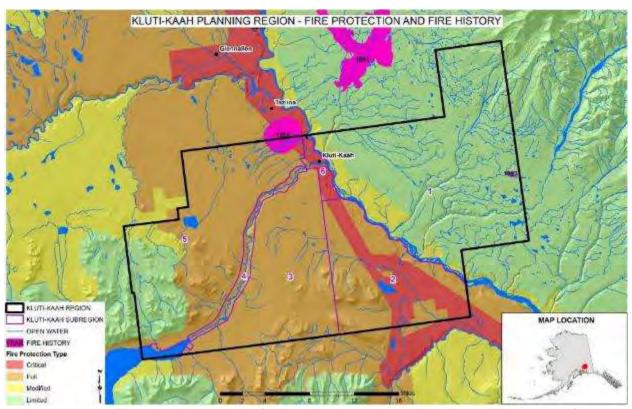


Figure 181. Current fire protection classes and fire history since 1940 in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory. Data from Alaska Interagency Coordination Center.

Ecosystem Diversity

The biophysical settings present within the Kluti-Kaah Planning region are displayed in Figure 182. Table 29 displays the acres for each biophysical setting and disturbance class. Figure 183 is a map of ecosystem diversity (represented by biophysical setting and disturbance class) in the Kluti-Kaah Planning Region.

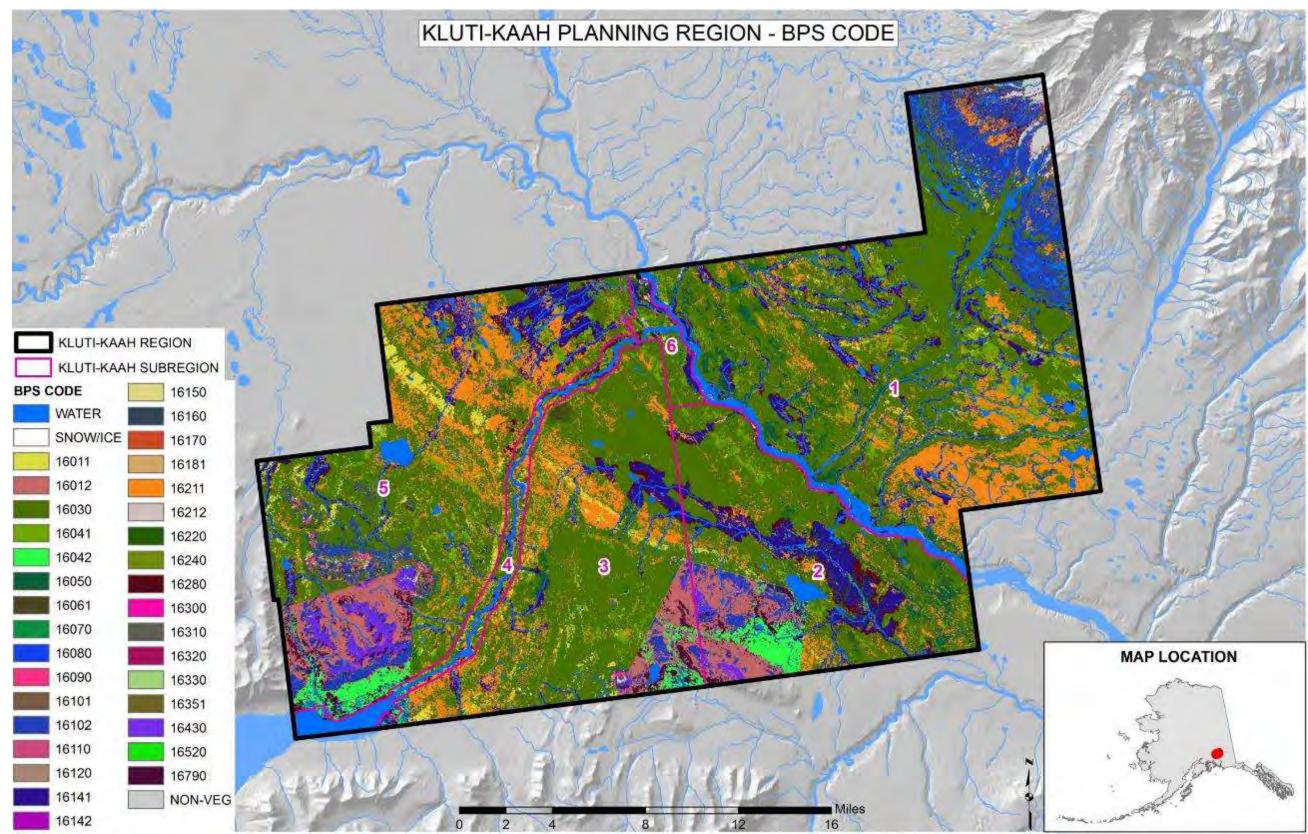


Figure 182. Biophysical settings (codes) in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory. Data from LANDFIRE.

Table 29. Acres by BpS code and disturbance class (A – E) in the Kluti-Kaah Planning Region. The BpS vegetation label is provided as well.

BpS Code _Dist Class	BpS Vegetation Label	Acres	BpS Code _Dist Class	BpS Vegetation Label	Acres
11	Open Water	9212.0	16142_A	Montane Floodplain-Subboreal	2.0
12	Perrennial Ice-Snow	146.3	16142_B	Montane Floodplain-Subboreal	0.4
16011_A	Treeline White Spruce-Boreal	737.7	16142_C	Montane Floodplain-Subboreal	0.4
16011_B	Treeline White Spruce-Boreal	6929.2	16142_D	Montane Floodplain-Subboreal	1.8
16011_C	Treeline White Spruce-Boreal	2.0	16150_A	Large River Floodplain	190.6
16012_A	Treeline White Spruce-SubBoreal	2510.4	16150_B	Large River Floodplain	30.2
16012_B	Treeline White Spruce-SubBoreal	12957.2	16150_C	Large River Floodplain	49.8
16012_C	Treeline White Spruce-SubBoreal	3.6	16150_D	Large River Floodplain	2.9
16030_A	White Spruce-Hardwood-Boreal	2304.7	16150_E	Large River Floodplain	1.1
16030_B	White Spruce-Hardwood-Boreal	18647.2	16160_A	Riparian Stringer	2.0
16030_C	White Spruce-Hardwood-Boreal	151152.8	16160_B	Riparian Stringer	11.1
16030_E	White Spruce-Hardwood-Boreal	257.1	16170_A	Shrub and Herbaceous Floodplain	153.0
16041_A	Mesic Black Spruce-Boreal	808.2	16170_B	Shrub and Herbaceous Floodplain	8.9
16041_B	Mesic Black Spruce-Boreal	3209.2	16170_C	Shrub and Herbaceous Floodplain	1.6
16041_C	Mesic Black Spruce-Boreal	1922.6	16170_D	Shrub and Herbaceous Floodplain	1.6
16041_D	Mesic Black Spruce-Boreal	15856.1	16170_E	Shrub and Herbaceous Floodplain	44.3
16041_E	Mesic Black Spruce-Boreal	5.1	16181_A	Herbaceous Fen	47.6
16042_A	Mesic Black Spruce-SubBoreal	992.8	16181_B	Herbaceous Fen	10.7
16042_B	Mesic Black Spruce-SubBoreal	5678.9	16181_C	Herbaceous Fen	5.8
16042_C	Mesic Black Spruce-SubBoreal	295.1	16181_D	Herbaceous Fen	76.3
16042_D	Mesic Black Spruce-SubBoreal	1.6	16211_A	Dwarf Black Spruce Peatland-Boreal	1503.4
16050_A	Mesic Birch-Aspen	11065.3	16211_B	Dwarf Black Spruce Peatland-Boreal	2887.8
16050_B	Mesic Birch-Aspen	2841.3	16211_C	Dwarf Black Spruce Peatland-Boreal	4110.3
16050_D	Mesic Birch-Aspen	4.2	16211_D	Dwarf Black Spruce Peatland-Boreal	46672.3
16050_E	Mesic Birch-Aspen	8.9	16212_A	Dwarf Black Spruce Peatland-Subboreal	1.8
16061_A	Dry Aspen-Steppe Bluff	39.1	16212_B	Dwarf Black Spruce Peatland-Subboreal	3.8
16061_B	Dry Aspen-Steppe Bluff	482.8	16212_C	Dwarf Black Spruce Peatland-Subboreal	14.5
16061_C	Dry Aspen-Steppe Bluff	580.2	16220_A	Black Spruce Wet-Mesic Slope	177.2
16061_D	Dry Aspen-Steppe Bluff	1305.7	16220_B	Black Spruce Wet-Mesic Slope	191.3
16070_A	Subalpine Balsam Poplar-Aspen	10.9	16220_C	Black Spruce Wet-Mesic Slope	80.3
16070_B	Subalpine Balsam Poplar-Aspen	39.1	16220_D	Black Spruce Wet-Mesic Slope	962.7
16080_A	Avalanche Slope Shrubland	47.1	16240_A	Deciduous Shrub Swamp	574.4
16080_B	Avalanche Slope Shrubland	66.5	16240_B	Deciduous Shrub Swamp	8.0
16090_A	Mesic Subalpine Alder	283.6	16280_A	Low Shrub-Tussock Tundra	7553.0
16090_B	Mesic Subalpine Alder	325.1	16280_B	Low Shrub-Tussock Tundra	1316.8
16102_A	Mesic Scrub Birch-Willow	5820.5	16280_C	Low Shrub-Tussock Tundra	278.4
16102_B	Mesic Scrub Birch-Willow	19204.9	16290_A	Tussock Tundra	0.4
16110_A	Mesic Bluejoint Meadow	17.6	16300_A	Wet Black Spruce-Tussock	35.4
16120_A	Dry Grassland	10.5	16300_B	Wet Black Spruce-Tussock	24.0
16141_A	Montane Floodplain-Boreal	3118.9	16300_C	Wet Black Spruce-Tussock	81.6
16141_B	Montane Floodplain-Boreal	3788.7	16310_A	Alpine Dwarf-Shrub Summit	72.9
16141_C	Montane Floodplain-Boreal	1591.7	16310_B	Alpine Dwarf-Shrub Summit	0.9
16141_D	Montane Floodplain-Boreal	19285.0	16320_A	Alpine Talus and Bedrock	1458.2
16141_E	Montane Floodplain-Boreal	1432.4	16320_B	Alpine Talus and Bedrock	22.0

BpS Code _Dist Class	BpS Vegetation Label	Acres	BpS Code _Dist Class	BpS Vegetation Label	Acres
16330_A	Alpine Mesic Herbaceous Meadow	326.9	16520_B	Subalpine Alder-Salmonberry	4.0
16351_A	Alpine Ericaceous Dwarf-Shrubland	2264.0	16550_A	Montane Floodplain	1.3
16351_B	Alpine Ericaceous Dwarf-Shrubland	70.9	16590_A	Mountain Hemlock Peatland	0.4
16372_B	Alpine Floodplain	0.7	16620_A	Emergent Marsh	0.7
16430_A	Alpine Dwarf Shrubland	3121.3	16790_A	White Spruce-Hardwood-SubBoreal	939.8
16430_B	Alpine Dwarf Shrubland	98.3	16790_B	White Spruce-Hardwood-SubBoreal	4760.1
16481_A	Mountain Hemlock	0.7	16790_C	White Spruce-Hardwood-SubBoreal	143.0
16500_A	Periglacial Woodland-Shrubland	0.7	16790_D	White Spruce-Hardwood-SubBoreal	8.2
16500_B	Periglacial Woodland-Shrubland	0.2	31	Barren-Rock-Sand-Clay	4797.1
16520_A	Subalpine Alder-Salmonberry	3.1			

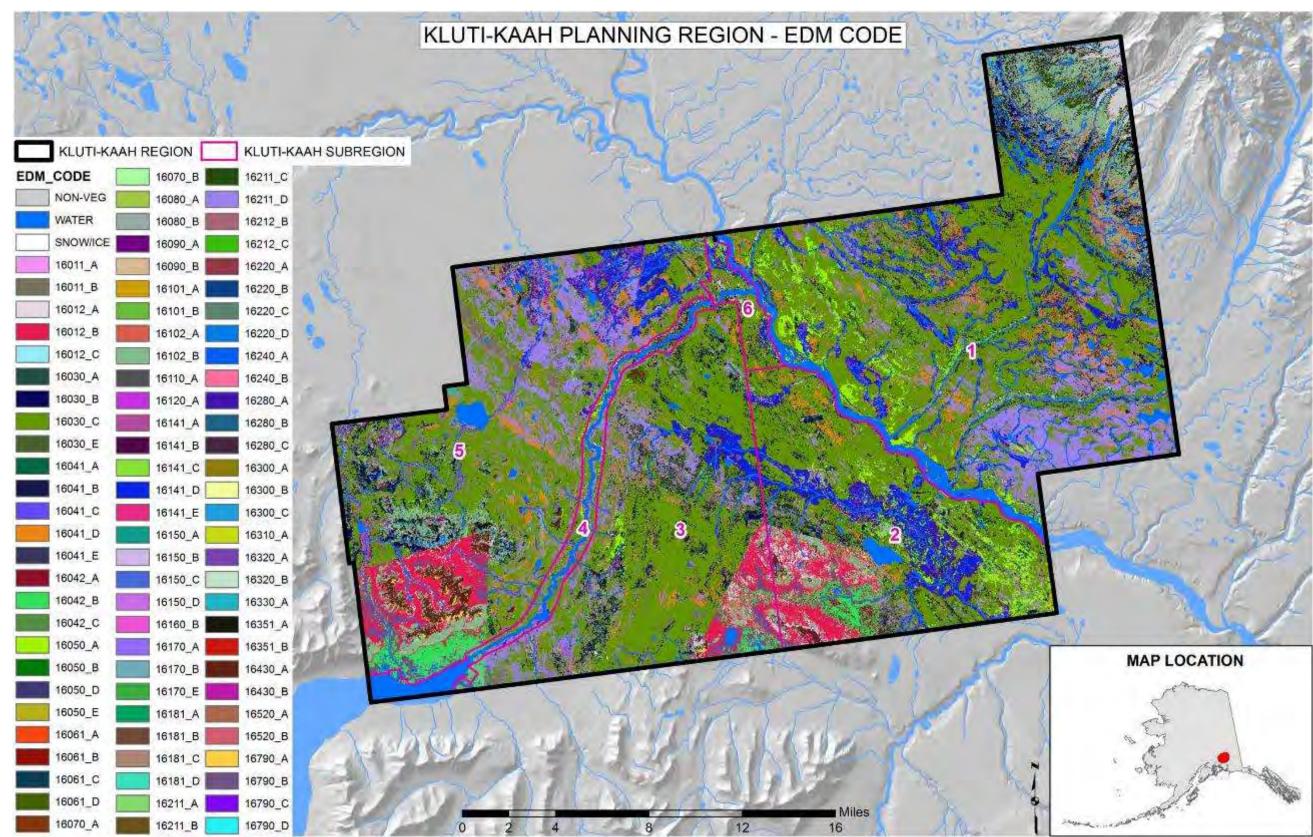


Figure 183. Map of ecosystem diversity in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory. Numbers in the legend refer to the BpS code, while letters (A-E) refer to the disturbance class.

Berry Production Areas

Ecological sites that had the potential for producing desirable species of berries were identified. These sites were then further identified as to the disturbance states most likely to support berry production. While numerous factors can influence where berry-producing vegetation occurs, the combination of ecological sites and disturbance states allows for the identification of areas with higher probabilities of finding good berry production. Figure 184 shows the areas identified with the highest potential for berry production. It may be desirable to avoid disturbances to these areas, but ground-truthing of the actual presence of good berry production is recommended.

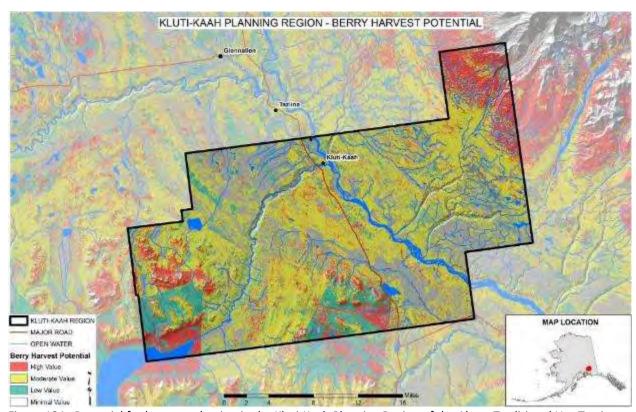


Figure 184. Potential for berry production in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory based on ecological sites and disturbance states derived from remotely sensed information.

Moose and Caribou Model Results

The habitat suitability figure for moose spring habitat is found in Figure 185 and the landscape rating for moose spring habitat is found in Figure 186. The habitat suitability figure for moose summer habitat is found in Figure 187 and the landscape rating for moose summer habitat is found in Figure 188. The habitat suitability figure for moose winter habitat is found in Figure 189 and the landscape rating for moose winter habitat is found in Figure 190. The habitat suitability figure for caribou summer/calving habitat are found in Figure 191 and the landscape rating for caribou summer/calving habitat are found Figure 192. The habitat suitability figure for caribou winter habitat are found in Figure 193 and the landscape rating for caribou winter habitat are found Figure 194. A complete description of the moose habitat model can be found in Appendix A. A complete description of the caribou habitat model can be found in Appendix B.

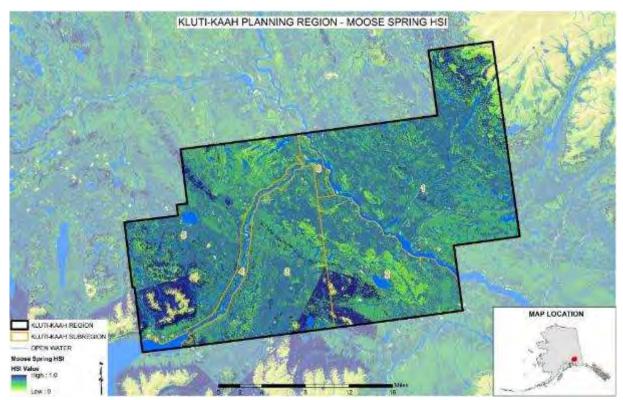


Figure 185. Habitat suitability for moose during spring in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory.

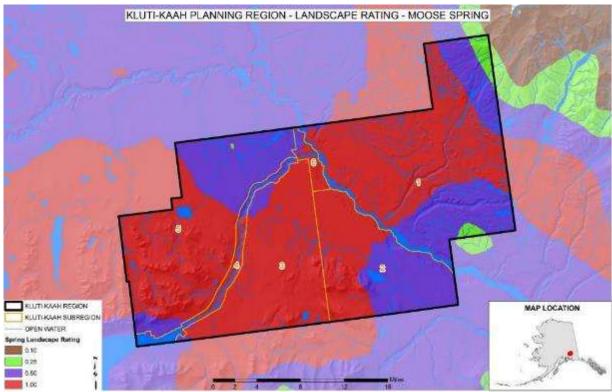


Figure 186. Landscape rating for moose during spring in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory.

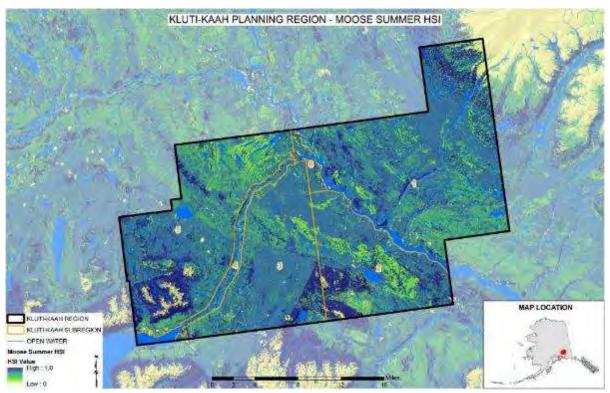


Figure 187. Habitat suitability for moose during summer in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory.

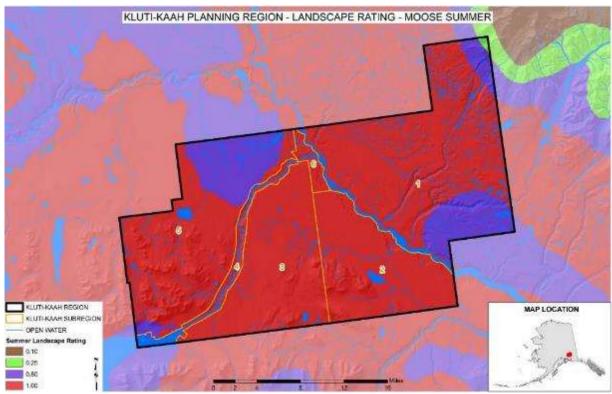


Figure 188. Landscape rating for moose during summer in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory.

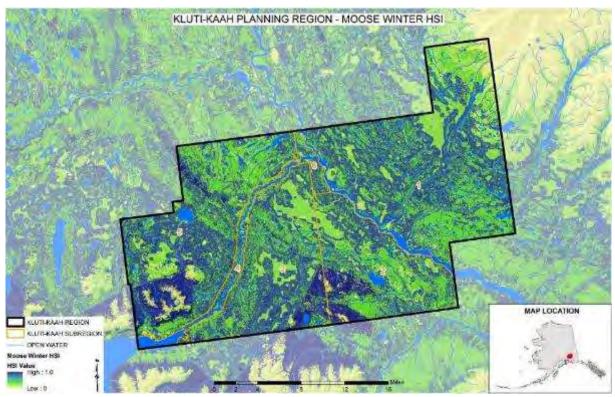


Figure 189. Habitat suitability for moose during winter in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory.

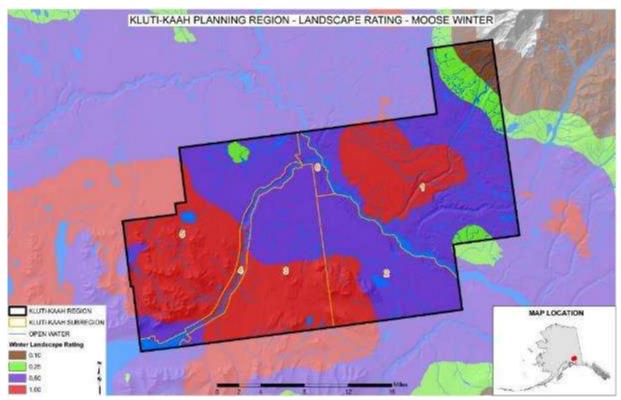


Figure 190. Landscape rating for moose during winter in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory.

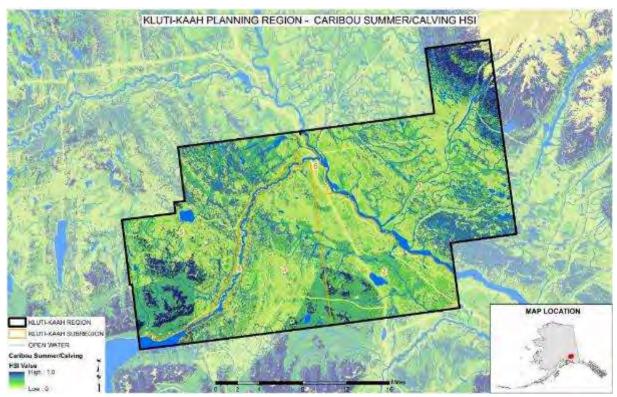


Figure 191. Habitat suitability for caribou during summer/calving in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory.

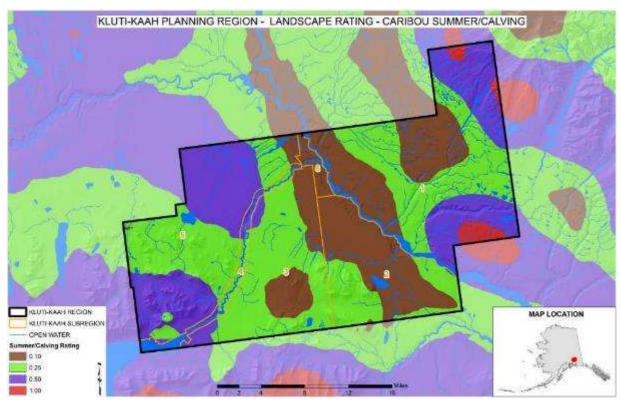


Figure 192. Landscape rating for caribou during summer/calving in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory.

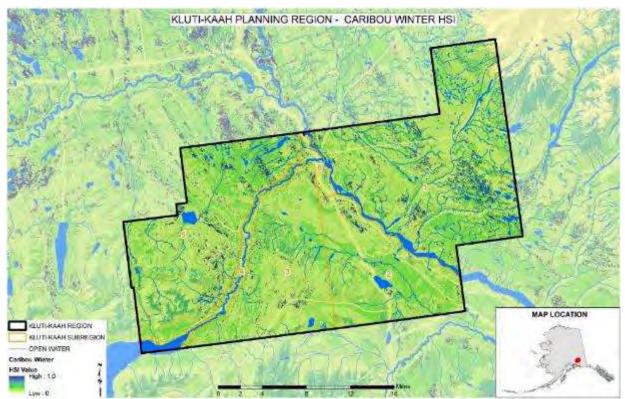


Figure 193. Habitat suitability for caribou during winter in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory.

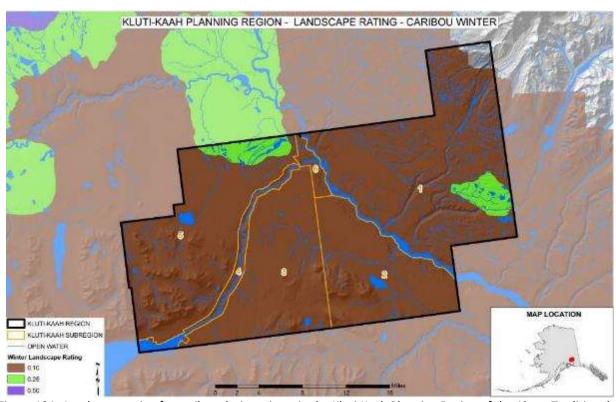


Figure 194. Landscape rating for caribou during winter in the Kluti-Kaah Planning Region of the Ahtna Traditional Use Territory.

Kluti-Kaah Site Improvement Areas

Potential treatment sites identified in the Kluti-Kaah area are displayed in figure 195. Based on all of the stated criteria, a number of sites were located for designated treatments. Figures 196-199 depict the locations of these areas. Appendix C provides a description of each of these sites. Sites ranging from approximately 35-52 acres in size were identified and are listed in Table 30. This table gives the site size and the treatment goal of primarily timber stand improvement and biomass production or moose browse enhancement. The table also displays an estimate of the total biomass of all standing above ground biomass. This estimate is not tonnage of usable biomass for energy but the total of all stems regardless of size. However, stands on the higher end of the total amount will typically have larger trees and are a candidate for biomass energy production.

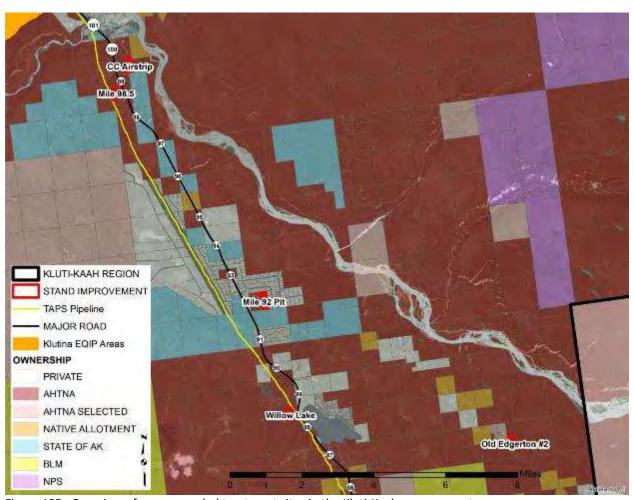


Figure 195. Overview of recommended treatment sites in the Kluti-Kaah management area.

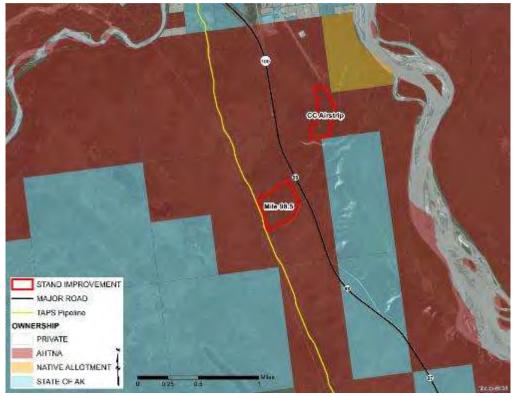


Figure 196. Map of proposed treatment areas (CC Airstrip and Mile 98.5) in the Kluti-Kaah Planning Region showing surface ownership and aerial imagery.

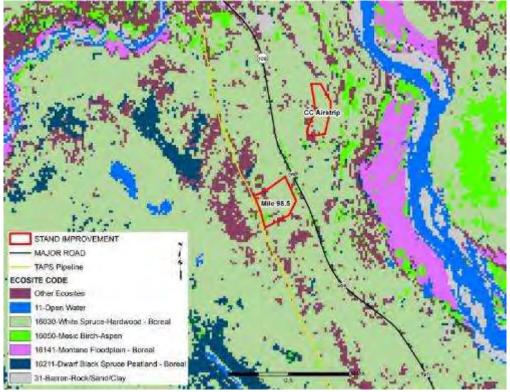


Figure 197. Map of proposed treatment areas (CC Airstrip and Mile 98.5) in the Kluti-Kaah Planning Region showing ecological sites.

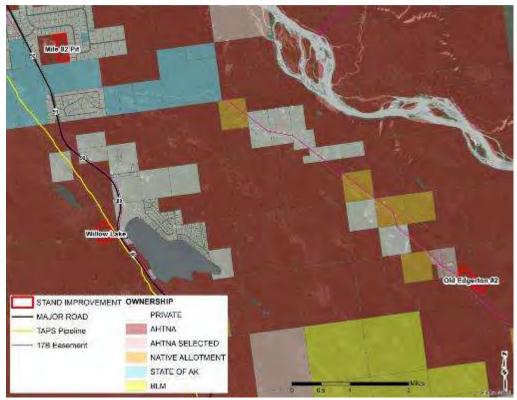


Figure 198. Map of proposed habitat improvement areas (Mile 92 Pit, Willow Lake, and Old Edgerton #2) in the Kluti-Kaah Planning Region showing surface ownership and aerial imagery.

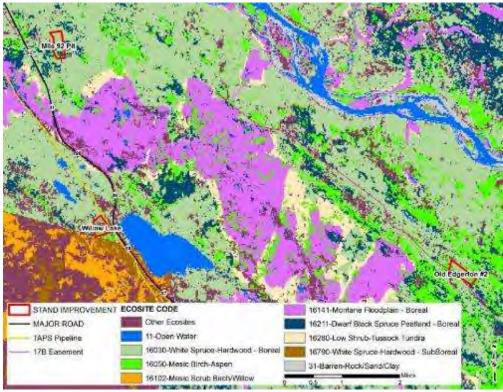


Figure 199. Map of proposed habitat improvement areas (Mile 92 Pit, Willow Lake, and Old Edgerton #2) in the Kluti-Kaah Planning Region showing ecological sites.

Table 30. Treatment site with treatment goal, site size, total tons of biomass and the BpS code and disturbance class for each site in the Kluti-Kaah Planning Region.

Site Name	Treatment Goal	Acres	Biomass - Total BpS Co	
Site Name	Treatment Goal Acre		(tons)	_Dist Code
CC Airstrip	Biomass/Browse	34.8	1668.0	16030_C
Mile 92 Pit	Biomass/Timber/Brow	51.2	1976.0	16030_C
Mile 98.5	Moose Browse	52.4	2426.6	16030_C
Old Edgerton #2	Moose Browse	49.9	1227.9	16030_C
Willow Lake	Moose Browse	44.9	784.3	16030_C

Mentasta Planning Region

The Mentasta area includes an area of 641,274 acres. Figure 200 displays the overall planning area along with associated infrastructure. Ownership of this and the surrounding area is shown in Figure 201. As Figure 201 displays, land ownership patterns are varied in this area with Ahtna owning 42.0% (269,100 acres) of the land.

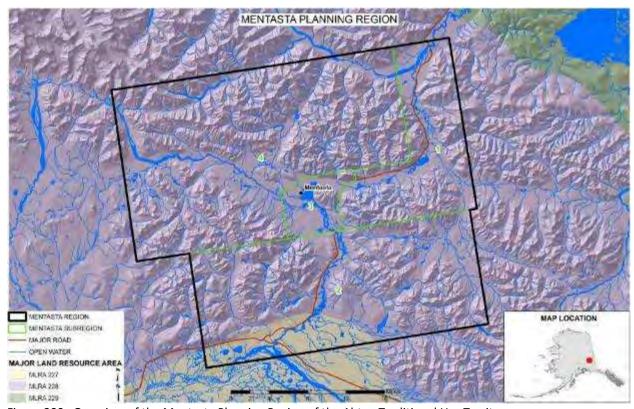


Figure 200. Overview of the Mentasta Planning Region of the Ahtna Traditional Use Territory.

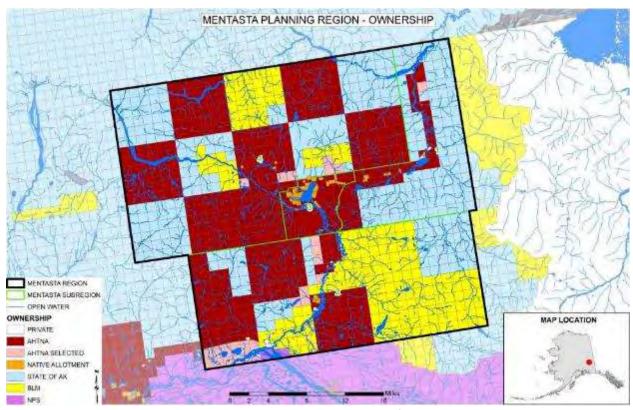


Figure 201. Land ownership patterns in the Mentasta Planning Region of the Ahtna Traditional Use Territory.

Climate

The Mentasta area falls within the Alaska Range and has a sub-arctic continental climate characterized by long cold winters and short warm summers. The average minimum temperature in January is -11.7 °F (-24.3 °C); daily low temperatures of -50 °F (-46 °C) or less occur frequently during the winter and may last for two or more weeks. The average maximum temperature in July is 68.5 °F (20.3 °C) and on occasion exceed 85 °F (30 °C). Although the daily minimum temperature in summer averages in the forties, freezing temperatures have been recorded in every month resulting in a varied growing season length each year. Mean annual precipitation is 15.4 inches (39.1 cm) with 45% being received as rain during the growing season (June-August). Average annual snowfall is 55.4 inches (140.7 cm). Continuous sunlight and twilight occur from early June through mid-July. Day length at the winter solstice is less than 5 hours long.

Soils

There is not detailed soil information for the Mentasta Planning Area. Large scale soil data are available for Alaska through STATSGO program. These data give information on certain soil characteristics. Figure 202 shows general soil textures and Figure 203 shows soil drainage for the Mentasta area.

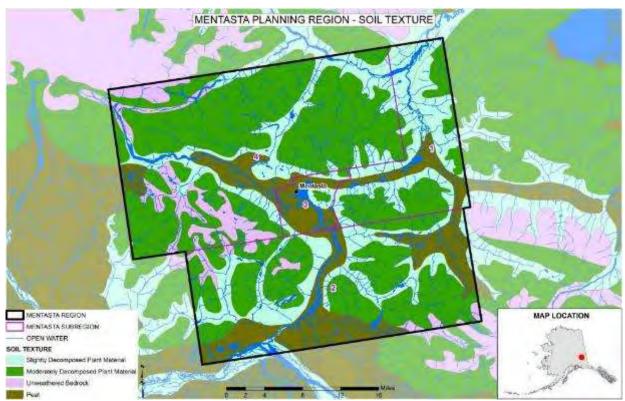


Figure 202. Soil texture in the Mentasta Planning Region of the Ahtna Traditional Use Territory. Data from NRCS STATSGO database for Alaska.

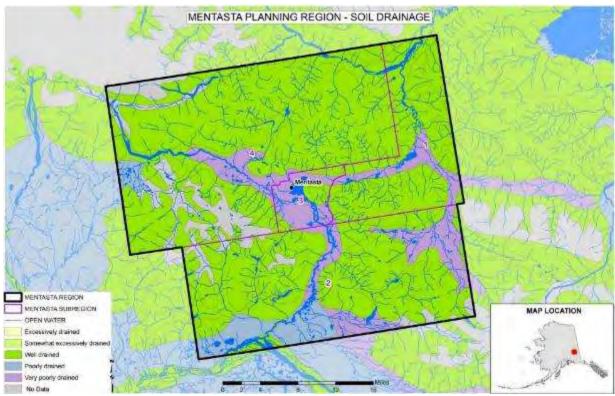


Figure 203. Soil drainage in the Mentasta Planning Region of the Ahtna Traditional Use Territory. Data from NRCS STATSGO database for Alaska.

Permafrost

Figure 204 displays the occurrence of permafrost in the Mentasta project area as interpreted from the soil survey information.

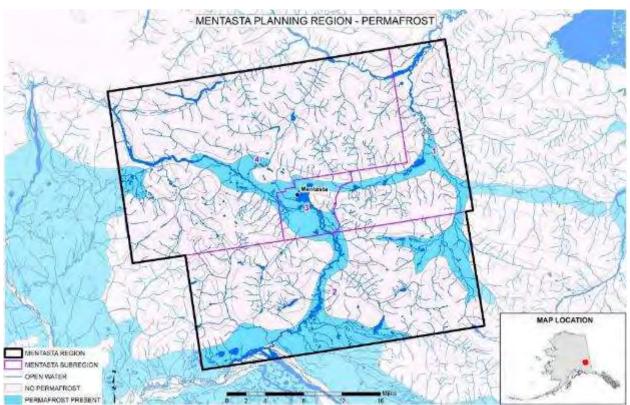


Figure 204. Permafrost in the Mentasta Planning Region of the Ahtna Traditional Use Territory. Data from NRCS STATSGO database for Alaska.

Vegetation Description

The <u>Denali National Park Area Soil Survey</u> provided a general description of the vegetation occurring higher elevation areas of central Alaska. While the region specifically described in the report is located approximately 200 miles (320 km) west of the Mentasta Planning Region it is relevant due to climatic, elevational, and biotic conditions. It stated: "General patterns of vegetation in the study area are the result of two major influences: the elevation gradient of the Alaska Range, and the different climactic regions north and south of the range. Much of the Denali Park is above tree line, and almost one-sixth is non-vegetated ice and rocky mountain slopes. In the vegetated zone, harsh conditions at high elevations limit plant communities to dwarf shrubs and herbaceous meadows in nivation hollows. Medium or tall shrubs are found lower down the slopes and these grade into forests or woodlands on well-drained substrates at lower elevations. Poor drainage at all elevations, because of glacial drift or permafrost, limits productivity. In lowlands, wet woodlands, shrubs, and herbaceous communities are found in a mosaic of fens, bogs, marshes and muskegs. Mountain vegetation of the Alaska Mountains Section is dominated by white mountain avens (Dryas octopetala) - dwarf ericaceous shrubs, which grade into medium-sized shrubs dominated by shrub birch and ericaceous shrubs such as blueberry (Vaccinium uliginosum), Labrador tea (Ledum palustre ssp. decumbens and L. groenlandicum) and crowberry (Empetrum hermaphroditum). On cooler, more northerly aspects these shrubs sometimes have high

percentages of sedge and other herbaceous vegetation. Warmer low slopes, especially in the Kantishna Hills and Park headquarters areas, support white spruce/mixed scrub woodlands."

Disturbance Factors

Fire is a major disturbance factor influencing the vegetation ecology in the project area. Figure 205 shows the fire history of the Mentasta area along with current fire protection zones. With Mentasta falling in the Alaska Range on the transition between the Copper Basin and the true Interior the rate of fire return is higher than areas further south in the Basin. Six large fires have occurred since 1940 in the project area, even though the most recent was in 1969. Although the level of fire occurring is substantially less than that occurring in more interior areas of Alaska north of the Alaska Range, fire is still a significant disturbance when it occurs. When fire does occur it serves to set back succession. It can also burn off the organic material at the ground surface, including peat that can occur on many sites. This can influence the thermal layer protecting the underlying permafrost on some sites, causing the permafrost to melt (thermokarst) and changing the site conditions through this process. Riparian areas are also influenced by flooding and ice events. These serve to set back succession of vegetation in riparian areas, and can even shift site conditions, particularly in the case of significant flooding events.

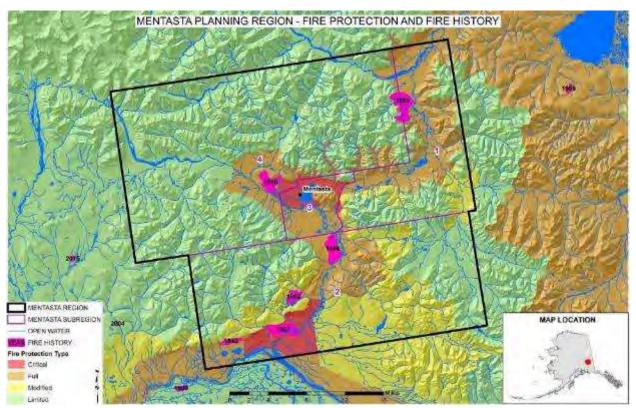


Figure 205. Current fire protection classes and fire history since 1940 in the Mentasta Planning Region of the Ahtna Traditional Use Territory. Data from Alaska Interagency Coordination Center.

Ecosystem Diversity

The biophysical settings present within the Mentasta Planning region are displayed in Figure 206. Table 31 displays the acres for each biophysical setting and disturbance class. Figure 207 is a map of ecosystem diversity (represented by biophysical setting and disturbance class) in the Mentasta Planning Region.

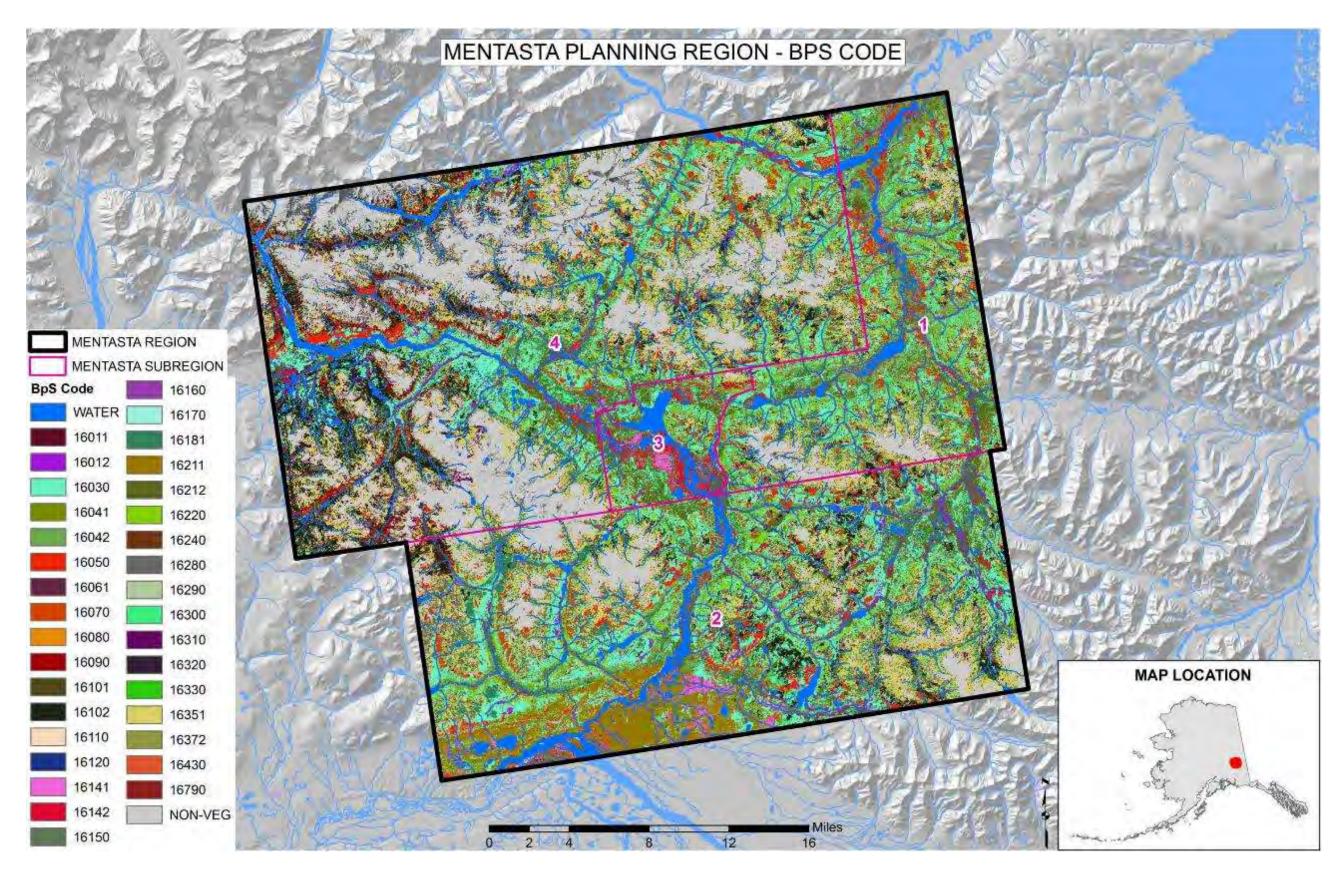


Figure 206. Biophysical settings (codes) in the Mentasta Planning Region of the Ahtna Traditional Use Territory. Data from LANDFIRE.

Table 31. Acres by BpS code and disturbance class (A - E) in the Mentasta planning region. The BpS vegetation label is provided as well.

BpS Code _Dist Class	BpS Vegetation Label	Acres	BpS Code _Dist Class	BpS Vegetation Label	Acres
11	Open Water	8237.7	16141_E	Montane Floodplain-Boreal	575.6
12	Perrennial Ice-Snow	15249.2	16142_A	Montane Floodplain-Subboreal	7370.6
16011_A	Treeline White Spruce-Boreal	145.9	16142_B	Montane Floodplain-Subboreal	3049.9
16011_B	Treeline White Spruce-Boreal	129.7	16142_C	Montane Floodplain-Subboreal	2120.1
16011_C	Treeline White Spruce-Boreal	0.4	16142_D	Montane Floodplain-Subboreal	2060.5
16012_A	Treeline White Spruce-SubBoreal	3531.0	16142_E	Montane Floodplain-Subboreal	335.1
16012_B	Treeline White Spruce-SubBoreal	3527.4	16150_A	Large River Floodplain	290.7
16012_C	Treeline White Spruce-SubBoreal	18.9	16150_B	Large River Floodplain	158.6
16030_A	White Spruce-Hardwood-Boreal	1797.0	16150_C	Large River Floodplain	136.1
16030_B	White Spruce-Hardwood-Boreal	40888.0	16150_D	Large River Floodplain	1.3
16030_C	White Spruce-Hardwood-Boreal	54815.5	16160_A	Riparian Stringer	64.0
16030_E	White Spruce-Hardwood-Boreal	332.9	16160_B	Riparian Stringer	350.7
16041_A	Mesic Black Spruce-Boreal	222.0	16160_C	Riparian Stringer	8.9
16041_B	Mesic Black Spruce-Boreal	1272.3	16170_A	Shrub and Herbaceous Floodplain	171.9
16041_C	Mesic Black Spruce-Boreal	1054.6	16170_B	Shrub and Herbaceous Floodplain	28.7
16041_D	Mesic Black Spruce-Boreal	4058.0	16170_C	Shrub and Herbaceous Floodplain	2.7
16041_E	Mesic Black Spruce-Boreal	24.2	16170_D	Shrub and Herbaceous Floodplain	8.9
16042_A	Mesic Black Spruce-SubBoreal	10569.8	16170_E	Shrub and Herbaceous Floodplain	105.9
16042_B	Mesic Black Spruce-SubBoreal	20690.1	16181_A	Herbaceous Fen	12295.3
16042_C	Mesic Black Spruce-SubBoreal	8081.2	16181_B	Herbaceous Fen	4218.6
16042_D	Mesic Black Spruce-SubBoreal	271.5	16181_C	Herbaceous Fen	709.2
16042_E	Mesic Black Spruce-SubBoreal	3.8	16181_D	Herbaceous Fen	20233.3
16050_A	Mesic Birch-Aspen	17742.5	16211_A	Dwarf Black Spruce Peatland-Boreal	1470.3
16050_B	Mesic Birch-Aspen	15857.9	16211_B	Dwarf Black Spruce Peatland-Boreal	1313.9
16050_D	Mesic Birch-Aspen	78.1	16211_C	Dwarf Black Spruce Peatland-Boreal	3840.3
16050_E	Mesic Birch-Aspen	123.7	16211_D	Dwarf Black Spruce Peatland-Boreal	15809.6
16061_A	Dry Aspen-Steppe Bluff	18.0	16212_A	Dwarf Black Spruce Peatland-Subboreal	3769.8
16061_B	Dry Aspen-Steppe Bluff	71.8	16212_B	Dwarf Black Spruce Peatland-Subboreal	10410.8
16061_C	Dry Aspen-Steppe Bluff	98.3	16212_C	Dwarf Black Spruce Peatland-Subboreal	28649.6
16061_D	Dry Aspen-Steppe Bluff	83.8	16220_A	Black Spruce Wet-Mesic Slope	6231.7
16070_A	Subalpine Balsam Poplar-Aspen	6483.5	16220_B	Black Spruce Wet-Mesic Slope	10907.8
16070_B	Subalpine Balsam Poplar-Aspen	7821.6	16220_C	Black Spruce Wet-Mesic Slope	140.6
16080_A	Avalanche Slope Shrubland	1213.8	16220_D	Black Spruce Wet-Mesic Slope	9207.6
16080_B	Avalanche Slope Shrubland	661.4	16240_A	Deciduous Shrub Swamp	2163.9
16090_A	Mesic Subalpine Alder	725.2	16240_B	Deciduous Shrub Swamp	142.1
16090_B	Mesic Subalpine Alder	3397.3	16280_A	Low Shrub-Tussock Tundra	4343.6
16102_A	Mesic Scrub Birch-Willow	6219.3	16280_B	Low Shrub-Tussock Tundra	3150.4
16102_B	Mesic Scrub Birch-Willow	45055.7	16280_C	Low Shrub-Tussock Tundra	808.2
16110_A	Mesic Bluejoint Meadow	42.9	16290_A	Tussock Tundra	321.1
16120_A	Dry Grassland	2.2	16290_B	Tussock Tundra	156.8
16141_A	Montane Floodplain-Boreal	1879.2	16300_A	Wet Black Spruce-Tussock	1526.1
16141_B	Montane Floodplain-Boreal	2207.9	16300_B	Wet Black Spruce-Tussock	1791.6
16141_C	Montane Floodplain-Boreal	891.8	16300_C	Wet Black Spruce-Tussock	4424.3
16141_D	Montane Floodplain-Boreal	3007.0	16310_A	Alpine Dwarf-Shrub Summit	3668.9

BpS Code _Dist Class	BpS Vegetation Label	Acres	BpS Code _Dist Class	BpS Vegetation Label	Acres
16310_B	Alpine Dwarf-Shrub Summit	29.8	16372_B	Alpine Floodplain	0.4
16320_A	Alpine Talus and Bedrock	8027.1	16430_A	Alpine Dwarf Shrubland	17.8
16320_B	Alpine Talus and Bedrock	129.9	16450_A	Alpine Mesic Herbaceous Meadow	0.7
16330_A	Alpine Mesic Herbaceous Meadow	283.3	16790_A	White Spruce-Hardwood-SubBoreal	11.3
16330_B	Alpine Mesic Herbaceous Meadow	53.2	16790_B	White Spruce-Hardwood-SubBoreal	3.8
16351_A	Alpine Ericaceous Dwarf-Shrubland	72099.4	16790_C	White Spruce-Hardwood-SubBoreal	2.2
16351_B	Alpine Ericaceous Dwarf-Shrubland	4605.8	31	Barren-Rock-Sand-Clay	99118.1
16372_A	Alpine Floodplain	1.8			

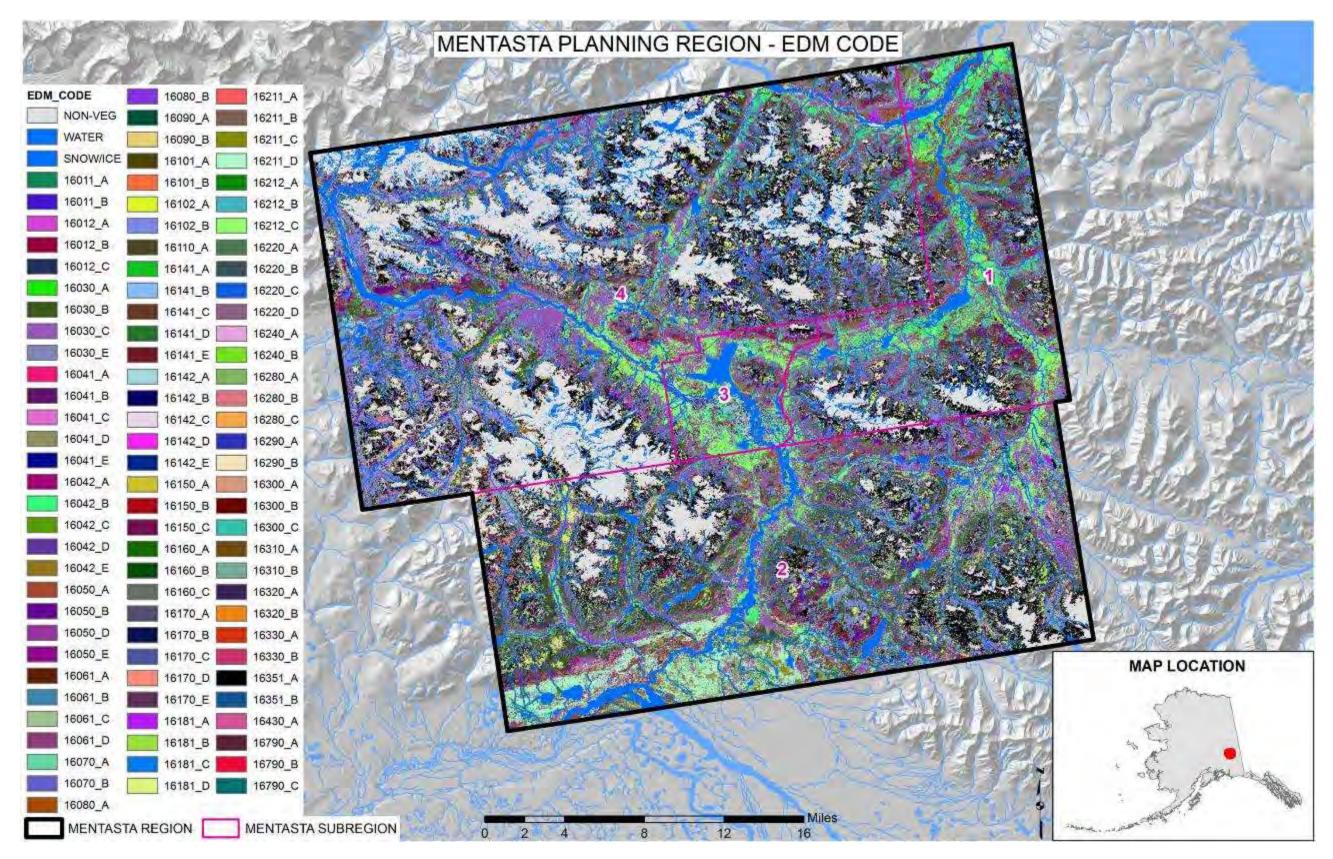


Figure 207. Map of ecosystem diversity in the Mentasta Planning Region of the Ahtna Traditional Use Territory. Numbers in the legend refer to the BpS code, while letters (A-E) refer to the disturbance class.

Berry Production Areas

Ecological sites that had the potential for producing desirable species of berries were identified. These sites were then further identified as to the disturbance states most likely to support berry production. While numerous factors can influence where berry-producing vegetation occurs, the combination of ecological sites and disturbance states allows for the identification of areas with higher probabilities of finding good berry production. Figure 208 shows the areas identified with the highest potential for berry production. It may be desirable to avoid disturbances to these areas, but ground-truthing of the actual presence of good berry production is recommended.

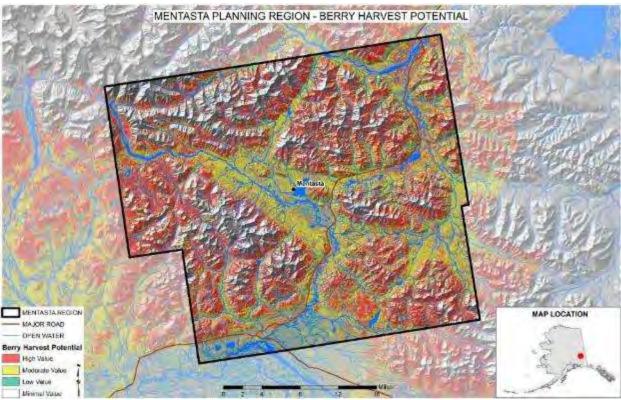


Figure 208. Potential for berry production in the Mentasta Planning Region of the Ahtna Traditional Use Territory based on ecological sites and disturbance states derived from remotely sensed information.

Moose and Caribou Model Results

The habitat suitability figure for moose spring habitat is found in Figure 209 and the landscape rating for moose spring habitat is found in Figure 210. The habitat suitability figure for moose summer habitat is found in Figure 211 and the landscape rating for moose summer habitat is found in Figure 212. The habitat suitability figure for moose winter habitat is found in Figure 213 and the landscape rating for moose winter habitat is found in Figure 214. The habitat suitability figure for caribou summer/calving habitat are found in Figure 215 and the landscape rating for caribou summer/calving habitat are found Figure 216. The habitat suitability figure for caribou winter habitat are found in Figure 217 and the landscape rating for caribou winter habitat are found Figure 218. A complete description of the moose habitat model can be found in Appendix A. A complete description of the caribou habitat model can be found in Appendix B.

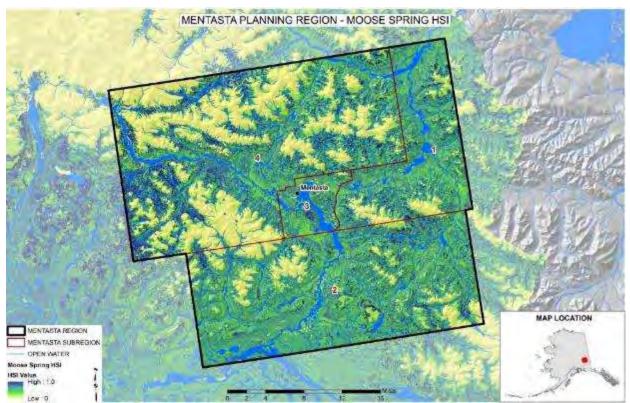


Figure 209. Habitat suitability for moose during spring in the Mentasta Planning Region of the Ahtna Traditional Use Territory.

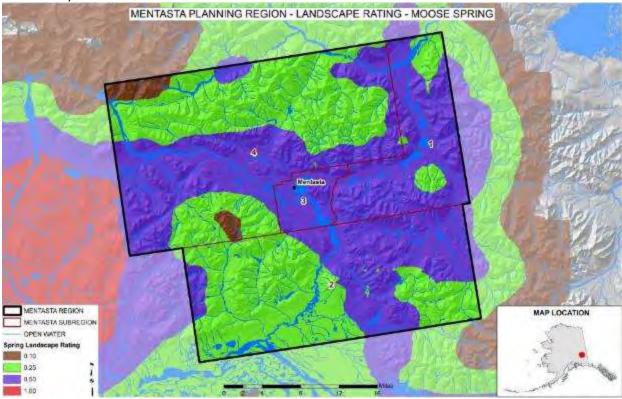


Figure 210. Landscape rating for moose during spring in the Mentasta Planning Region of the Ahtna Traditional Use Territory.

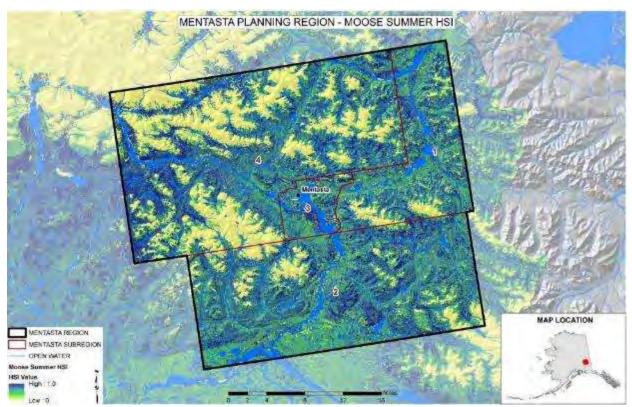


Figure 211. Habitat suitability for moose during summer in the Mentasta Planning Region of the Ahtna Traditional Use Territory.

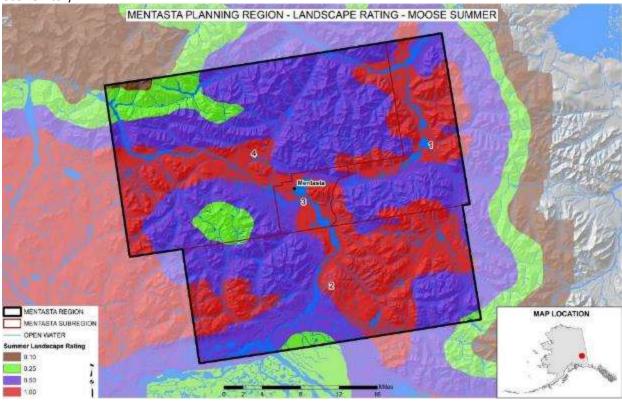


Figure 212. Landscape rating for moose during summer in the Mentasta Planning Region of the Ahtna Traditional Use Territory.

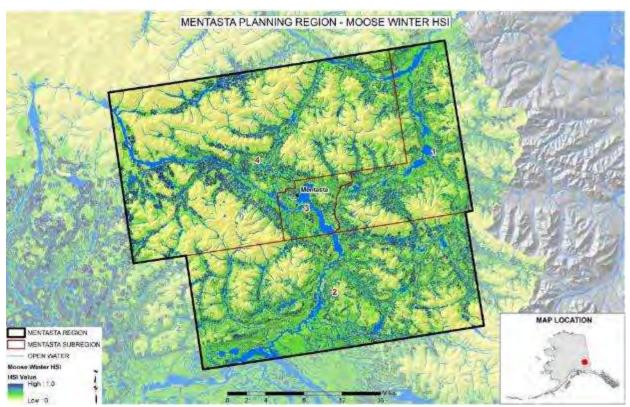


Figure 213. Habitat suitability for moose during winter in the Mentasta Planning Region of the Ahtna Traditional Use Territory.

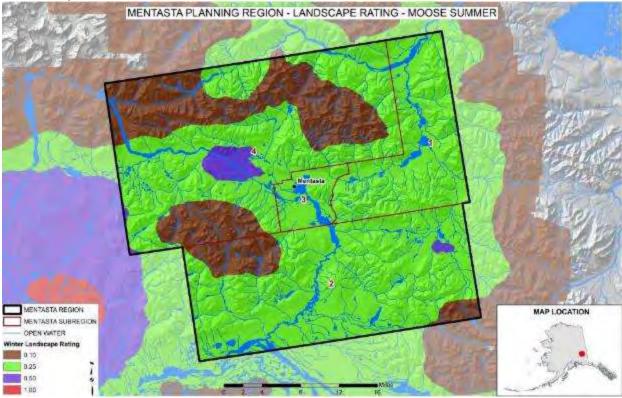


Figure 214. Landscape rating for moose during winter in the Mentasta Planning Region of the Ahtna Traditional Use Territory.



Figure 215. Habitat suitability for caribou during summer/calving in the Mentasta Planning Region of the Ahtna Traditional Use Territory.

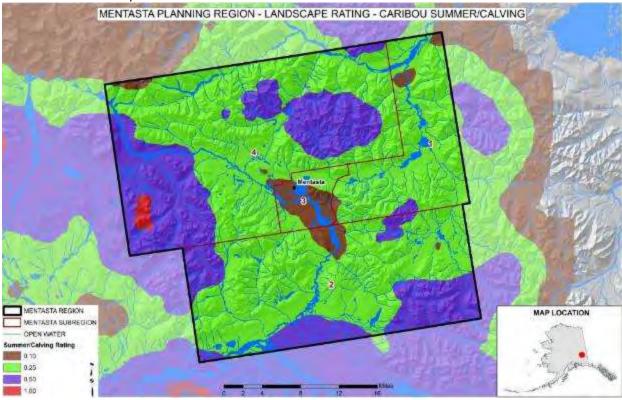


Figure 216. Landscape rating for caribou during summer/calving in the Mentasta Planning Region of the Ahtna Traditional Use Territory.

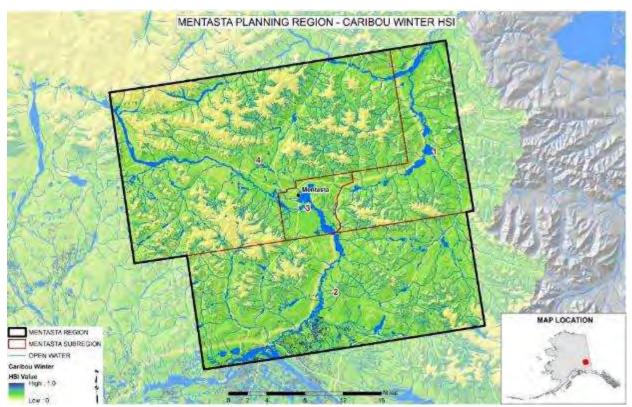


Figure 217. Habitat suitability for caribou during winter in the Mentasta Planning Region of the Ahtna Traditional Use Territory.

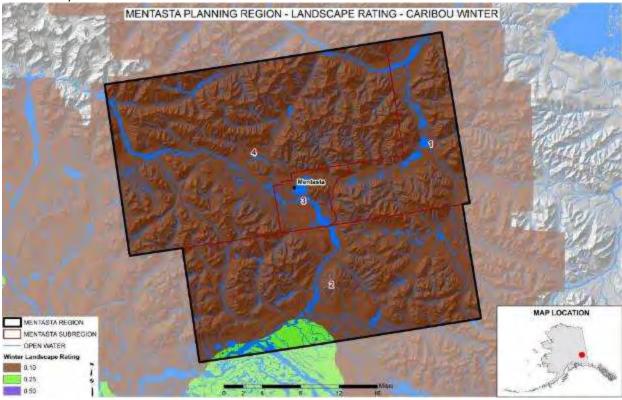


Figure 218. Landscape rating for caribou during winter in the Mentasta Planning Region of the Ahtna Traditional Use Territory.

Mentasta Site Improvement Areas

Potential treatment sites identified in the Mentasta area are displayed in figure 219. Based on all of the stated criteria, a number of sites were located for designated treatments. Figures 220-228 depict the locations of these areas. Appendix C provides a description of each of these sites. Sites ranging from approximately 13-51 acres in size were identified and are listed in Table 32. This table gives the site size and the treatment goal of primarily timber stand improvement and biomass production or moose browse enhancement. The table also displays an estimate of the total biomass of all standing above ground biomass. This estimate is not tonnage of usable biomass for energy but the total of all stems regardless of size. However, stands on the higher end of the total amount will typically have larger trees and are a candidate for biomass energy production.

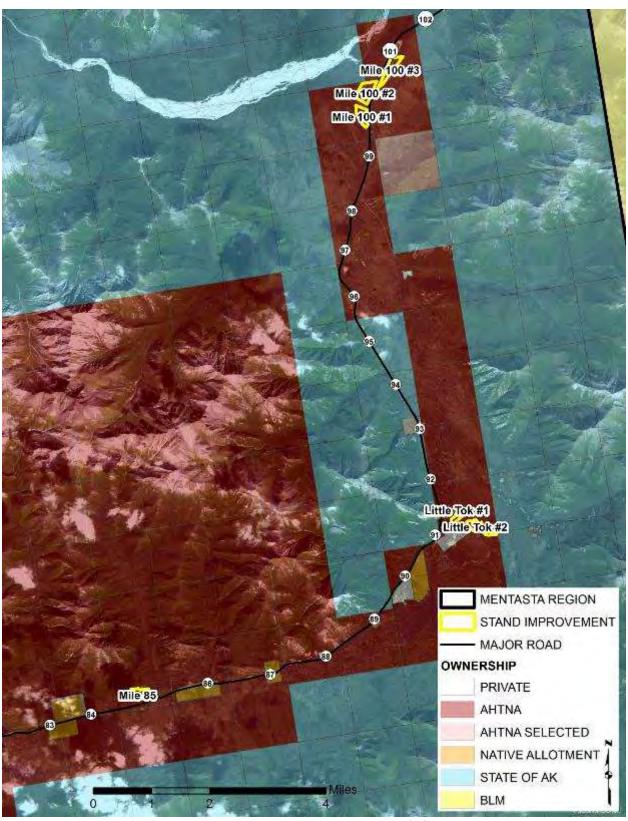


Figure 219. Overview of recommended treatment sites in the northern half of the Mentasta management area.

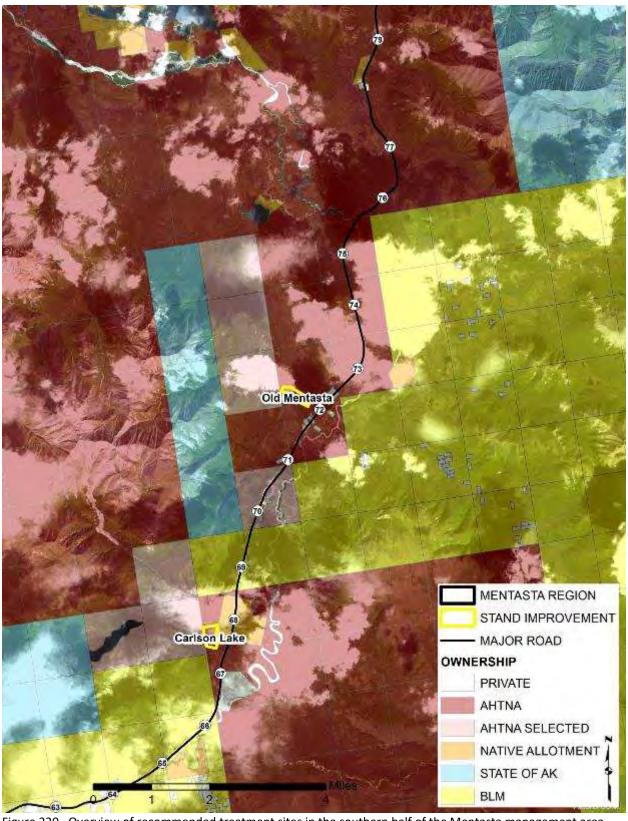


Figure 220. Overview of recommended treatment sites in the southern half of the Mentasta management area.

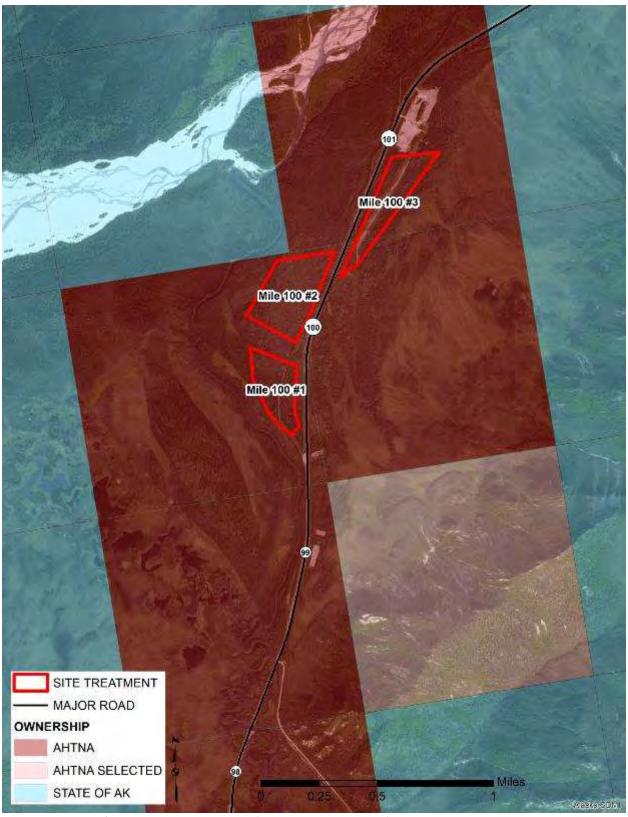


Figure 221. Map of proposed treatment areas (Mile 100 #1, #2, and #3) in the Mentasta Planning Region showing surface ownership and aerial imagery.

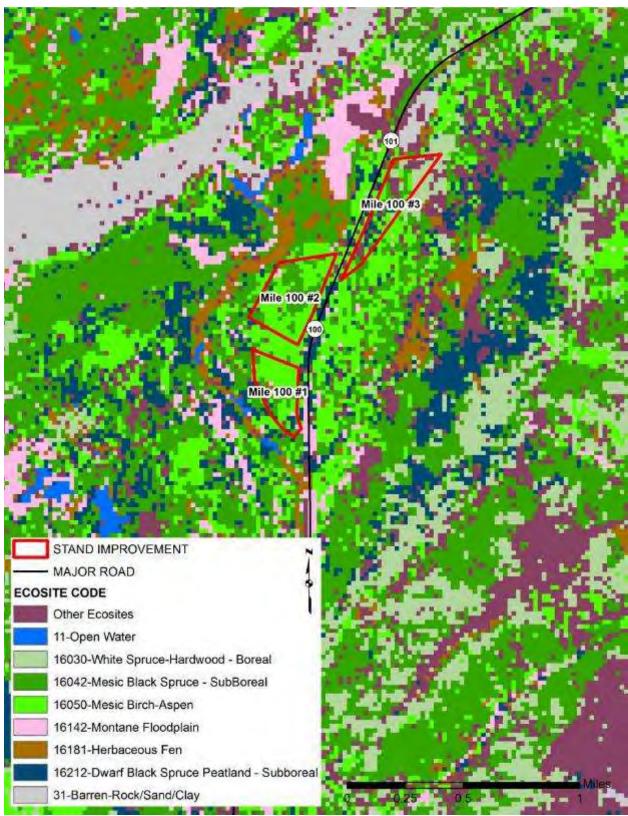


Figure 222. Map of proposed treatment areas (Mile 100 #1, #2, and #3) in the Mentasta Planning Region showing ecological sites.

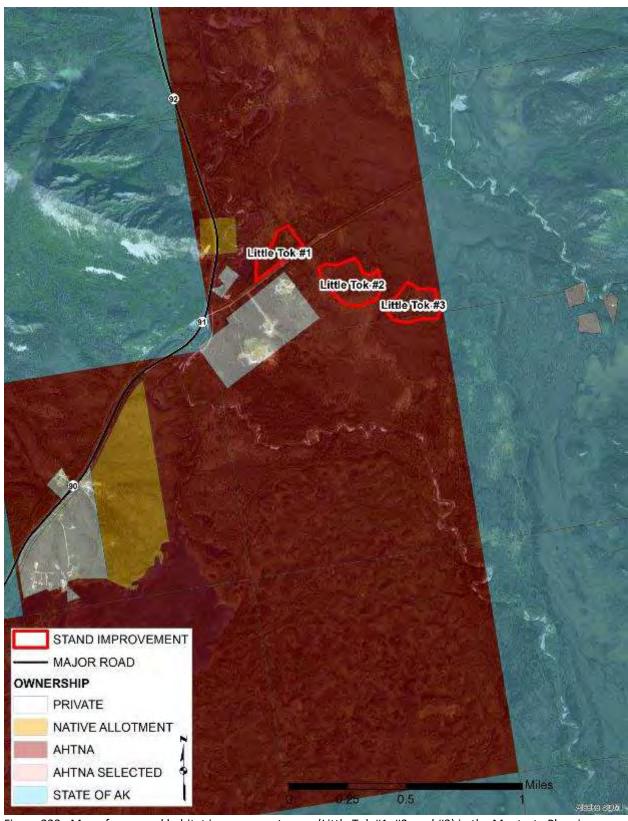


Figure 223. Map of proposed habitat improvement areas (Little Tok #1, #2, and #3) in the Mentasta Planning Region showing surface ownership and aerial imagery.

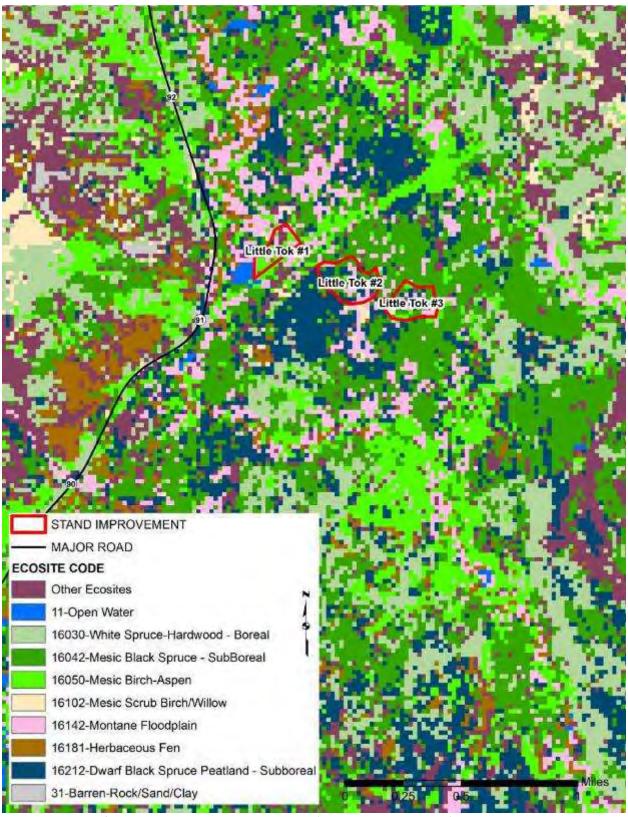


Figure 224. Map of proposed habitat improvement areas (Little Tok #1, #2, and #3) in the Mentasta Planning Region showing ecological sites.

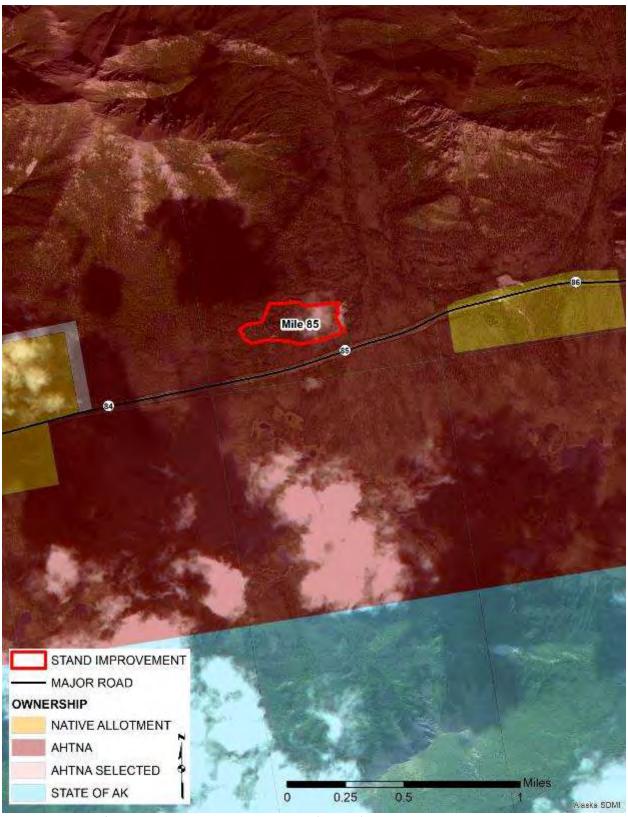


Figure 225. Map of proposed habitat improvement areas (Mile 85) in the Mentasta Planning Region showing surface ownership and aerial imagery.

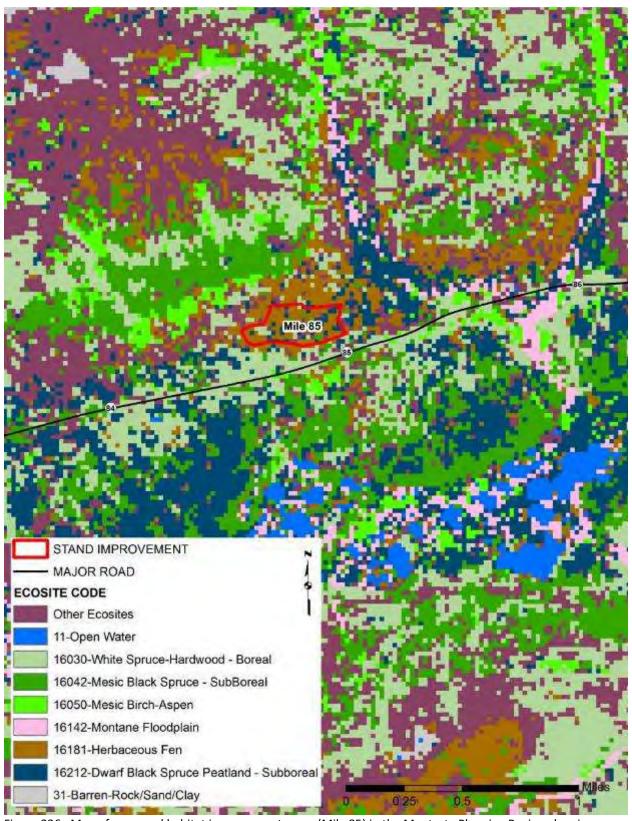


Figure 226. Map of proposed habitat improvement areas (Mile 85) in the Mentasta Planning Region showing ecological sites.

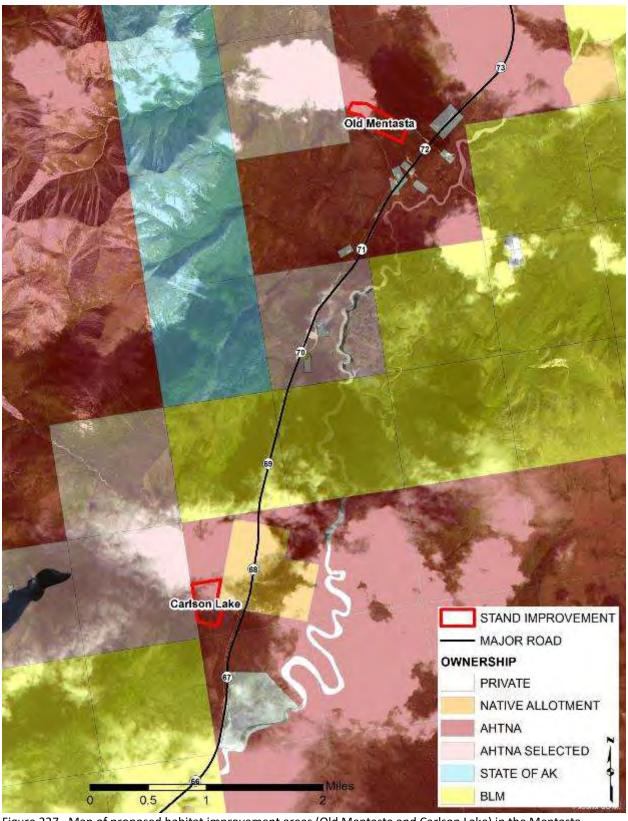


Figure 227. Map of proposed habitat improvement areas (Old Mentasta and Carlson Lake) in the Mentasta Planning Region showing surface ownership and aerial imagery.

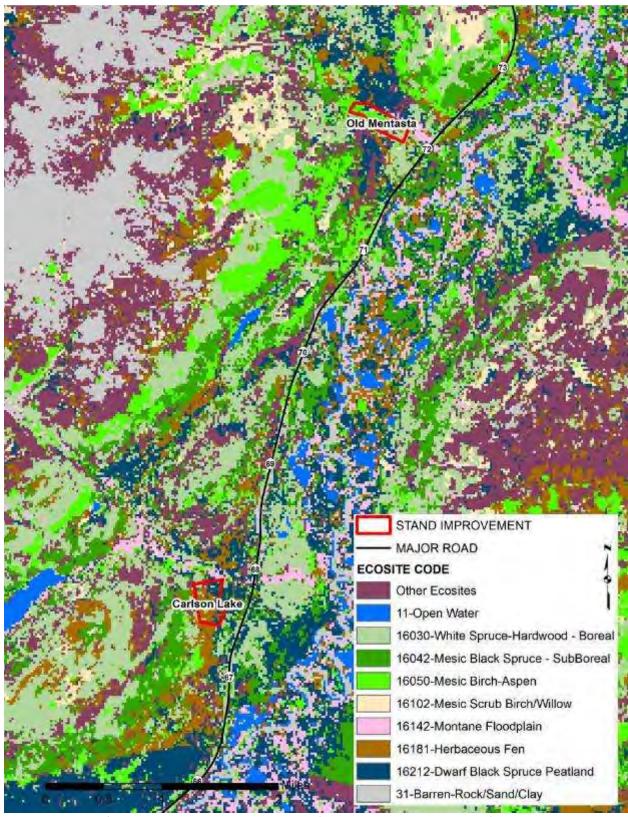


Figure 228. Map of proposed habitat improvement areas (Old Mentasta and Carlson Lake) in the Mentasta Planning Region showing ecological sites.

Table 32. Treatment site with treatment goal, site size, total tons of biomass and the BpS code and disturbance class for each site in the Mentasta Planning Region.

Site Name	Treatment Goal	Acres	Biomass - Total	BpS Code
Site Name	rreatment doar	Acres	(tons)	_Dist Code
Carlson Lake	Timber/Browse	48.8	634.5	16181_A
Little Tok #1	Moose Browse	12.8	116.1	16050_A
Little Tok #2	Timber Management	23.1	535.3	16212_C
Little Tok #3	Timber Management	17.6	397.7	16042_B
Mile 100 #1	Timber Management	30.3	459.3	16042_B
Mile 100 #2	Moose Browse	50.8	990.0	16042_B
Mile 100 #3	Moose Browse	37.2	847.9	16030_C
Mile 85		36.6	311.8	16181_A
Old Mentasta	Timber Improvement	44.5	733.2	16030_C

Tazlina Planning Region

The Tazlina Planning Region encompasses an area of 250,843 acres. Figure 229 displays the overall planning area along with associated infrastructure. Ownership of this and the surrounding area is shown in Figure 230. As Figure 230 displays, land ownership patterns are varied in this area with Ahtna, Inc. owning 48.3% (121,213 acres) of the land.

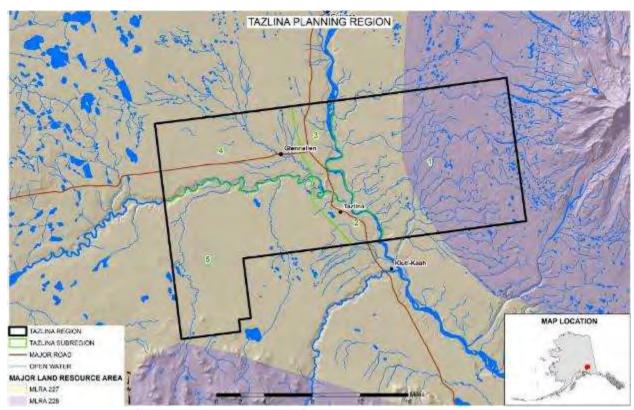


Figure 229. Overview of the Tazlina Planning Region of the Ahtna Traditional Use Territory.

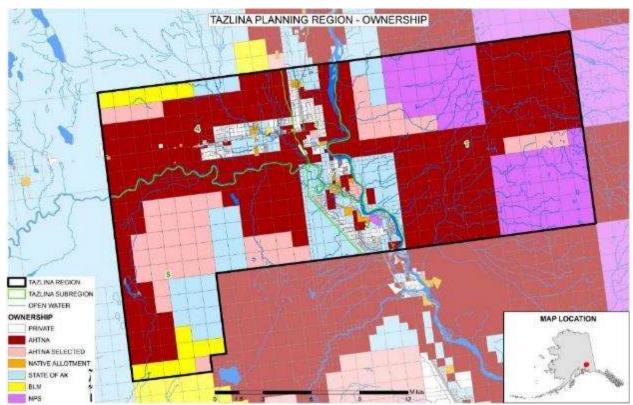


Figure 230. Land ownership patterns in the Tazlina Planning Region of the Ahtna Traditional Use Territory.

A description of the general geology, climate, soils, permafrost, and vegetation is found in Chapter 1 of this report. Figures showing these features specific to the Tazlina Planning Region are displayed below. Soil texture in the Tazlina area is shown in Figure 231 and Figure 232 displays soil drainages. Permafrost in the Tazlina area is shown in Figure 233.

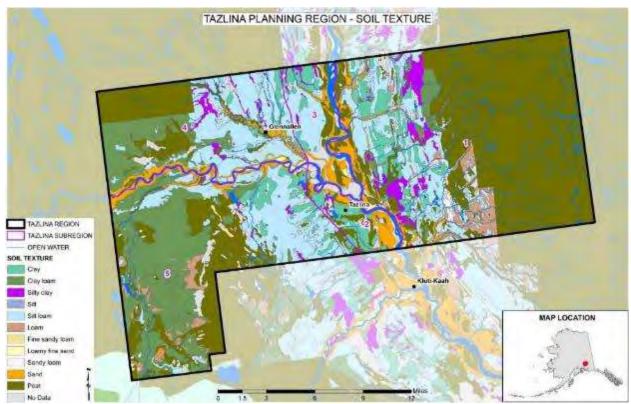


Figure 231. Soil texture in the Tazlina Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

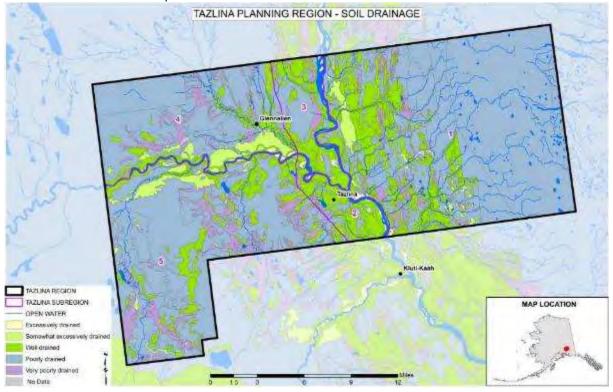


Figure 232. Soil drainage in the Tazlina Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

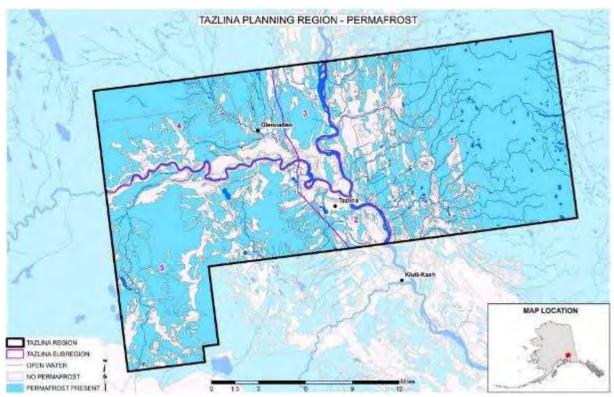


Figure 233. Permafrost in the Tazlina Planning Region of the Ahtna Traditional Use Territory. Data from NRCS SURRGO database for Soil Map Unit AK612 and NRCS STATSGO database for Alaska.

Disturbance Factors

Fire is the primary disturbance factor influencing the vegetation ecology in the Copper Basin. Figure 234 shows the fire history of the Tazlina area along with current fire protection zones. Although the level of fire occurring in this southcentral Alaska landscape is substantially less than that occurring in more interior areas of Alaska north of the Alaska Range, fire is still a significant disturbance when it occurs. When fire does occur it serves to set back succession. It can also burn off the organic material at the ground surface, including peat that can occur on many sites. This can influence the thermal layer protecting the underlying permafrost on some sites, causing the permafrost to melt (thermokarst) and changing the site conditions through this process. Riparian areas are also influenced by flooding and ice events. These serve to set back succession of vegetation in riparian areas, and can even shift site conditions, particularly in the case of significant flooding events.

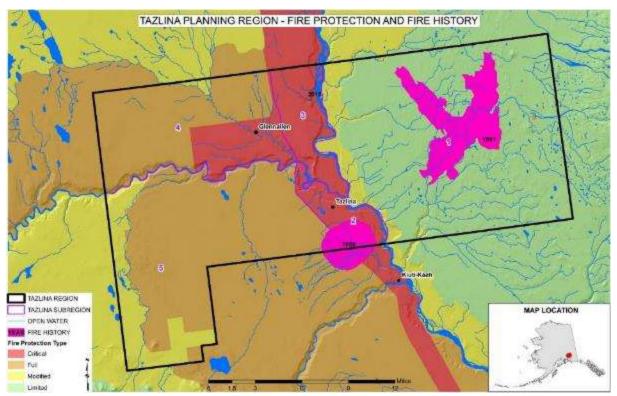


Figure 234. Current fire protection classes and fire history since 1940 in the Tazlina Planning Region of the Ahtna Traditional Use Territory. Data from Alaska Interagency Coordination Center.

Ecosystem Diversity

The biophysical settings present within the Tazlina Planning region are displayed in Figure 235. Table 33 displays the acres for each biophysical setting and disturbance class. Figure 236 is a map of ecosystem diversity (represented by biophysical setting and disturbance class) in the Tazlina Planning Region.

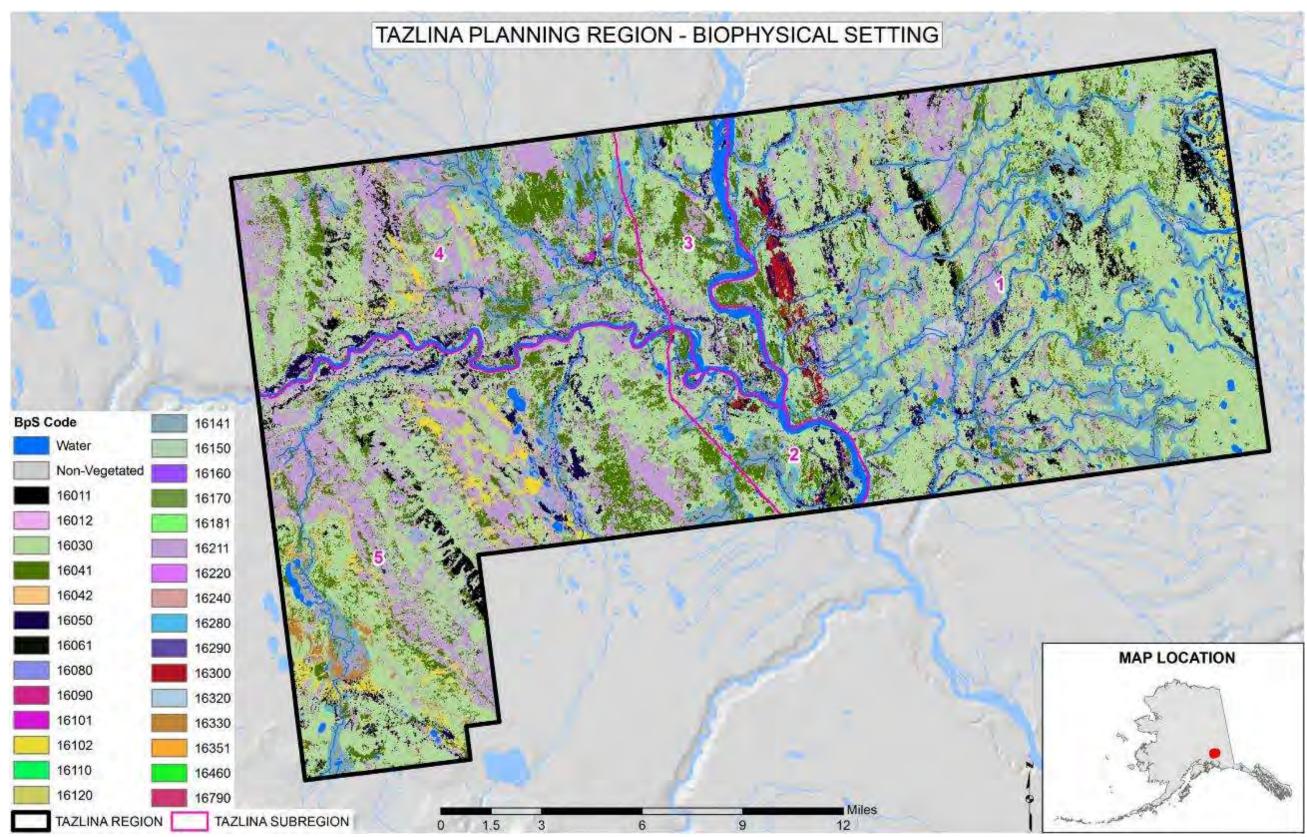


Figure 235. Biophysical settings (codes) in the Tazlina Planning Region of the Ahtna Traditional Use Territory. Data from LANDFIRE.

Table 33. Acres by BpS code and disturbance class (A - E) in the Tazlina Planning Region. The BpS vegetation label is provided as well.

BpS Code _Dist Class	BpS Vegetation Label	Acres	BpS Code Dist Class	BpS Vegetation Label	Acres
11	Open Water	4230.8	16150_B	Large River Floodplain	28.5
12	Perrennial Ice/Snow	22.5	16150_C	Large River Floodplain	57.2
16011_A	Treeline White Spruce – Boreal	1576.1	16150_D	Large River Floodplain	2.0
_ 16011_B	Treeline White Spruce – Boreal	4757.0	16150_E	Large River Floodplain	0.7
_ 16012_A	Treeline White Spruce – SubBoreal	15.4	16160_A	Riparian Stringer	1.8
16012_B	Treeline White Spruce – SubBoreal	30.9	16160_B	Riparian Stringer	6.0
16030_A	White Spruce-Hardwood – Boreal	3360.4	16170_A	Shrub and Herbaceous Floodplain	175.0
16030_B	White Spruce-Hardwood – Boreal	14292.9	16170_B	Shrub and Herbaceous Floodplain	9.1
16030_C	White Spruce-Hardwood – Boreal	93422.8	_ 16170_C	Shrub and Herbaceous Floodplain	1.8
16030_E	White Spruce-Hardwood – Boreal	236.4	16170_D	Shrub and Herbaceous Floodplain	1.8
_ 16041_A	Mesic Black Spruce – Boreal	1136.9	16170_E	Shrub and Herbaceous Floodplain	52.0
16041_B	Mesic Black Spruce – Boreal	2603.4	16181_A	Herbaceous Fen	51.2
16041_C	Mesic Black Spruce – Boreal	1820.8	16181_B	Herbaceous Fen	4.5
16041 D	Mesic Black Spruce – Boreal	21569.9	16181_C	Herbaceous Fen	0.9
16041 E	Mesic Black Spruce – Boreal	12.9	16181 D	Herbaceous Fen	13.8
16042 A	Mesic Black Spruce – SubBoreal	8.0	16211 A	Dwarf Black Spruce Peatland – Boreal	2374.1
16042 B	Mesic Black Spruce – SubBoreal	37.6	16211_B	Dwarf Black Spruce Peatland – Boreal	2930.7
16042_C	Mesic Black Spruce – SubBoreal	6.2	16211_C	Dwarf Black Spruce Peatland – Boreal	3109.3
16042 D	Mesic Black Spruce – SubBoreal	0.2	16211 D	Dwarf Black Spruce Peatland – Boreal	41100.2
16050 A	Mesic Birch-Aspen	5859.4	16212 A	Dwarf Black Spruce Peatland – Subboreal	3.6
16050_B	Mesic Birch-Aspen	1140.4	16212_B	Dwarf Black Spruce Peatland – Subboreal	5.8
16050_D	Mesic Birch-Aspen	11.1	16212_C	Dwarf Black Spruce Peatland – Subboreal	14.5
16050_E	Mesic Birch-Aspen	18.9	16220_A	Black Spruce Wet-Mesic Slope	81.4
16061_A	Dry Aspen-Steppe Bluff	86.1	16220_B	Black Spruce Wet-Mesic Slope	50.9
16061_B	Dry Aspen-Steppe Bluff	353.6	16220_C	Black Spruce Wet-Mesic Slope	221.5
16061_C	Dry Aspen-Steppe Bluff	296.2	16240_A	Deciduous Shrub Swamp	62.5
16061_D	Dry Aspen-Steppe Bluff	1623.5	16280_A	Low Shrub-Tussock Tundra	6032.5
16080_A	Avalanche Slope Shrubland	0.4	16280 B	Low Shrub-Tussock Tundra	793.5
16080 B	Avalanche Slope Shrubland	2.0	16280 C	Low Shrub-Tussock Tundra	149.9
16090 A	Mesic Subalpine Alder	30.5	16290 A	Tussock Tundra	0.2
16090 B	Mesic Subalpine Alder	6.9	16290 B	Tussock Tundra	0.2
16102 A	Mesic Scrub Birch/Willow	2483.7	16300 A	Wet Black Spruce-Tussock	767.0
16102_B	Mesic Scrub Birch/Willow	3422.4	16300_B	Wet Black Spruce-Tussock	127.2
_ 16110_A	Mesic Bluejoint Meadow	13.3	16300_C	Wet Black Spruce-Tussock	359.4
_ 16120_A	Dry Grassland	3.1	16320_A	Alpine Talus and Bedrock	706.8
_ 16141_A	, Montane Floodplain – Boreal	21082.6	16320_B	Alpine Talus and Bedrock	8.9
_ 16141_B	Montane Floodplain – Boreal	1434.2	16330_A	Alpine Mesic Herbaceous Meadow	919.2
_ 16141_C	Montane Floodplain – Boreal	650.1	16351_A	Alpine Ericaceous Dwarf-Shrubland	166.8
16141_D	Montane Floodplain – Boreal	58.9	16481_C	Mountain Hemlock	0.2
16141 E	Montane Floodplain – Boreal	32.9	16790_A	White Spruce-Hardwood – SubBoreal	26.0
16142_A	Montane Floodplain – Subboreal	3.6	16790_B	White Spruce-Hardwood – SubBoreal	43.1
16142 B	Montane Floodplain – Subboreal	2.0	16790 C	White Spruce-Hardwood – SubBoreal	24.7
16142_C	Montane Floodplain – Subboreal	0.7	16790_D	White Spruce-Hardwood – SubBoreal	0.9
_ 16150_A	Large River Floodplain	192.6	31	Non-Vegetated	2326.7

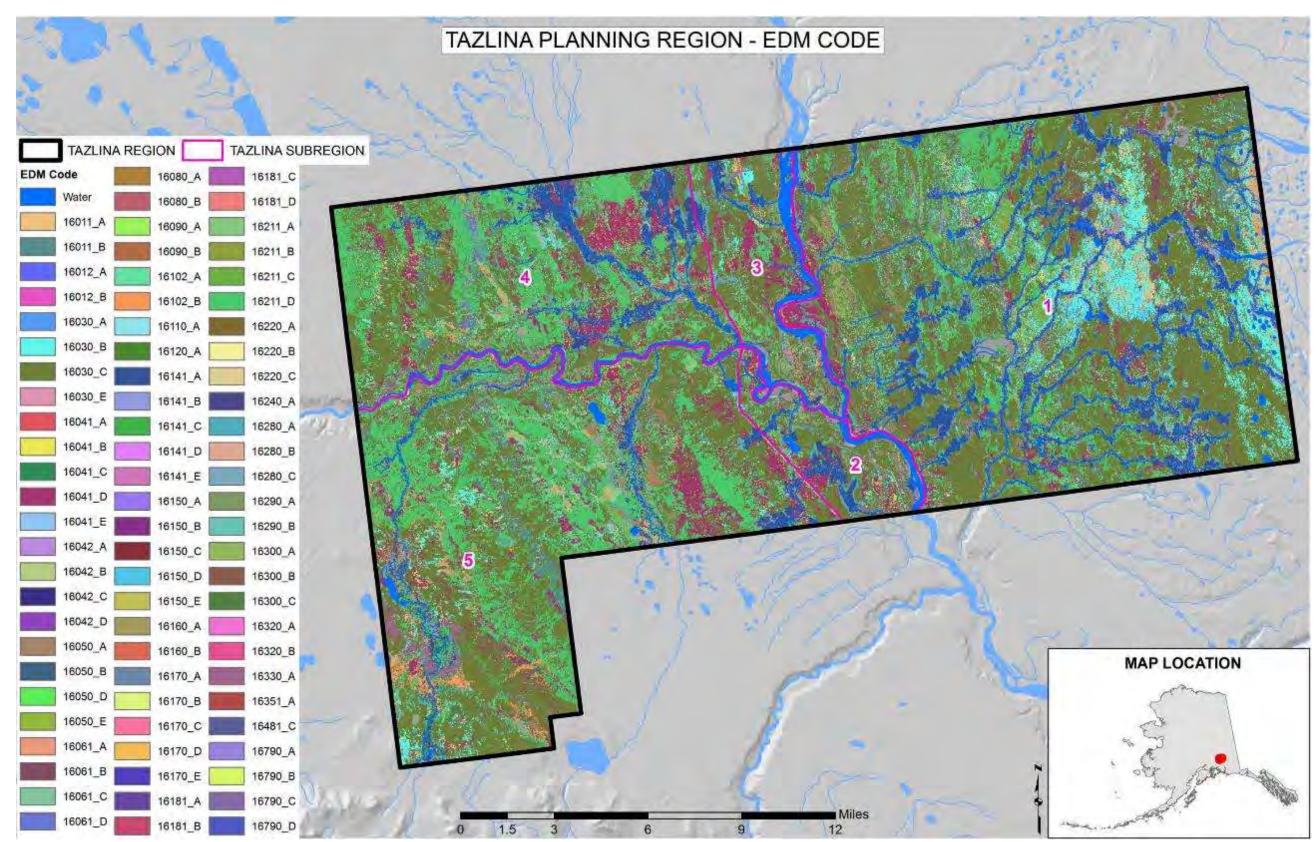


Figure 236. Map of ecosystem diversity in the Tazlina Planning Region of the Ahtna Traditional Use Territory. Numbers in the legend refer to the BpS code, while letters (A-E) refer to the disturbance class.

Berry Production Areas

Ecological sites that had the potential for producing desirable species of berries were identified. These sites were then further identified as to the disturbance states most likely to support berry production. While numerous factors can influence where berry-producing vegetation occurs, the combination of ecological sites and disturbance states allows for the identification of areas with higher probabilities of finding good berry production. Figure 237 shows the areas identified with the highest potential for berry production. It may be desirable to avoid disturbances to these areas, but ground-truthing of the actual presence of good berry production is recommended.

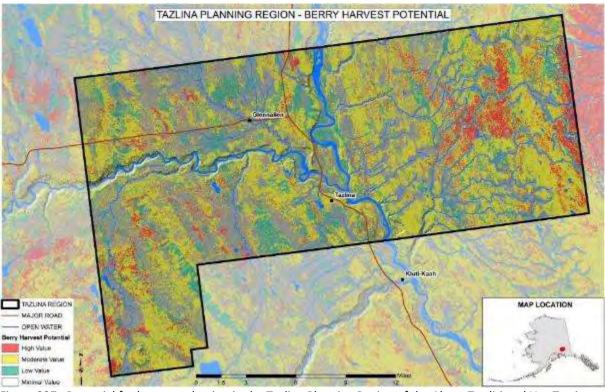


Figure 237. Potential for berry production in the Tazlina Planning Region of the Ahtna Traditional Use Territory based on ecological sites and disturbance states derived from remotely sensed information.

Moose and Caribou Model Results

The habitat suitability figure for moose spring habitat is found in Figure 238 and the landscape rating for moose spring habitat is found in Figure 239. The habitat suitability figure for moose summer habitat is found in Figure 240 and the landscape rating for moose summer habitat is found in Figure 241. The habitat suitability figure for moose winter habitat is found in Figure 242 and the landscape rating for moose winter habitat is found in Figure 243. The habitat suitability figure for caribou summer/calving habitat are found in Figure 244 and the landscape rating for caribou summer/calving habitat are found Figure 245. The habitat suitability figure for caribou winter habitat are found in Figure 246 and the landscape rating for caribou winter habitat are found Figure 247. A complete description of the moose habitat model can be found in Appendix A. A complete description of the caribou habitat model can be found in Appendix B.

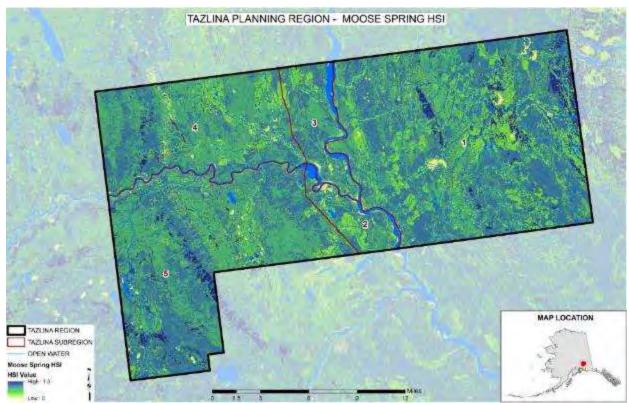


Figure 238. Habitat suitability for moose during spring in the Tazlina Planning Region of the Ahtna Traditional Use Territory.

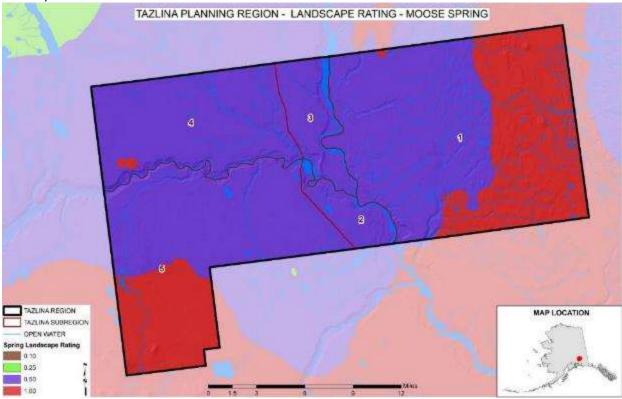


Figure 239. Landscape rating for moose during spring in the Tazlina Planning Region of the Ahtna Traditional Use Territory.

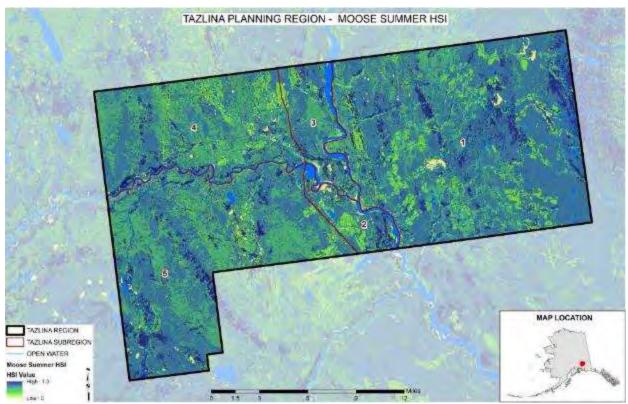


Figure 240. Habitat suitability for moose during summer in the Tazlina Planning Region of the Ahtna Traditional Use Territory.

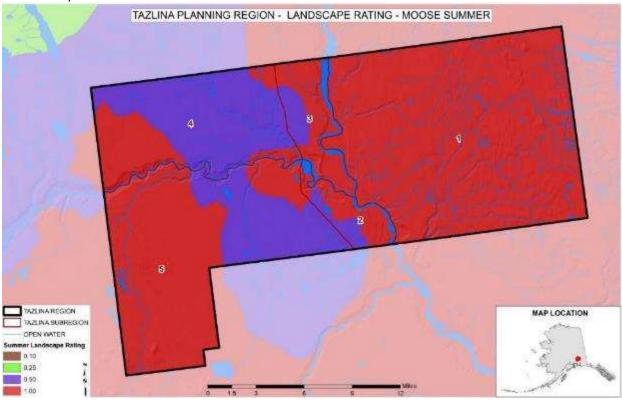


Figure 241. Landscape rating for moose during summer in the Tazlina Planning Region of the Ahtna Traditional Use Territory.

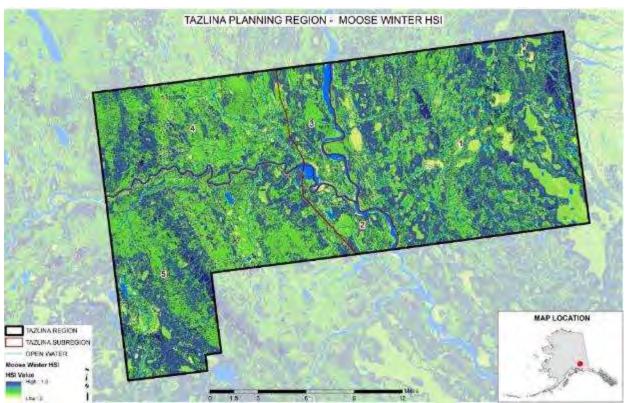


Figure 242. Habitat suitability for moose during winter in the Tazlina Planning Region of the Ahtna Traditional Use Territory.

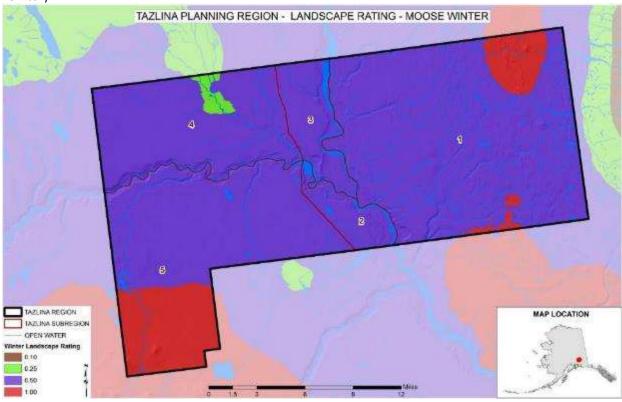


Figure 243. Landscape rating for moose during winter in the Tazlina Planning Region of the Ahtna Traditional Use Territory.

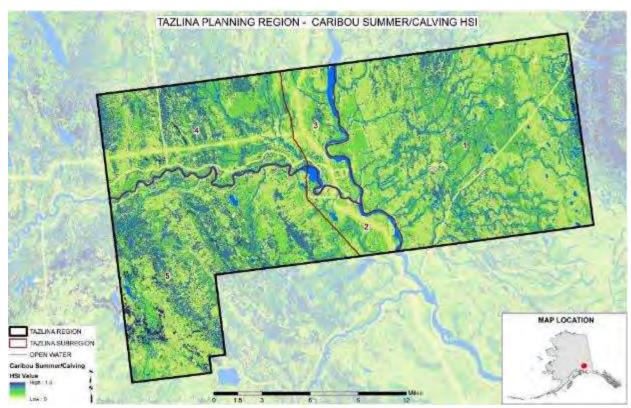


Figure 244. Habitat suitability for caribou during summer/calving in the Tazlina Planning Region of the Ahtna Traditional Use Territory.

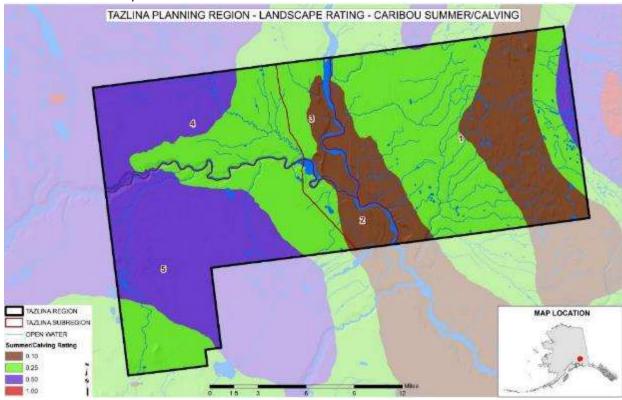


Figure 245. Landscape rating for caribou during summer/calving in the Tazlina Planning Region of the Ahtna Traditional Use Territory.

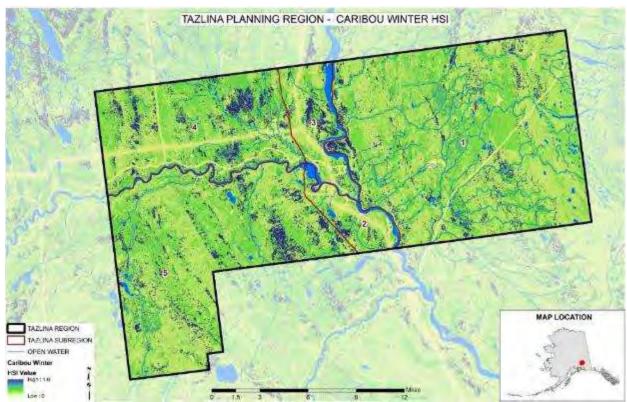


Figure 246. Habitat suitability for caribou during winter in the Tazlina Planning Region of the Ahtna Traditional Use Territory.

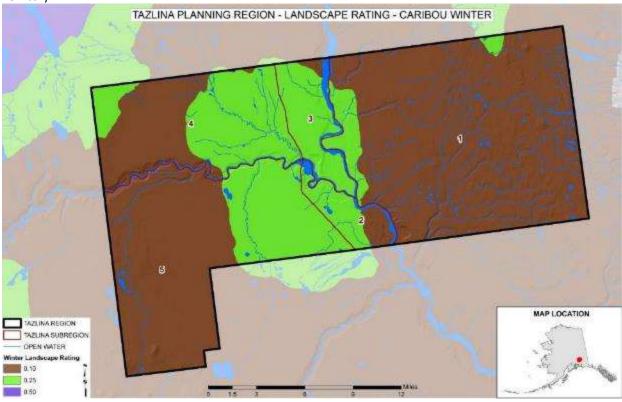


Figure 247. Landscape rating for caribou during winter in the Tazlina Planning Region of the Ahtna Traditional Use Territory.

Tazlina Site Improvement Areas

Site selection for improvement areas in the Tazlina project area identified a number of sites for designated treatments. Figure 248 provides an overview of the treatment site locations and figures 249-255 depict the locations and conditions of these areas. Appendix C provides a description of each of these sites.

Treatments for moose habitat improvement could use a number of possible mechanical treatments designed to stimulate growth of preferred moose browse species such as willow, or potentially prescribed fire. Treatments for fuel mitigation would use similar methods, but are designed to reduce the amounts of flammable material in the primary lines of defense and provide a location where defensive actions can be taken to counter an approaching fire. Specific treatments need to be determined for each selected site. Site characteristics are listed in Table 34.

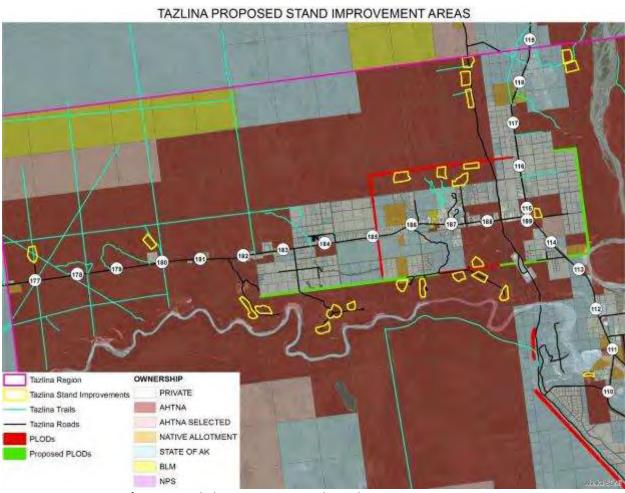


Figure 248. Overview of recommended treatment sites in the Tazlina management area.



Figure 249. Vegetation conditions at a proposed site that could be treated by mechanical treatments to open up the stand and stimulate production of willow or other browse species. This is site category 16030-C.



Figure 250. Vegetation conditions at the Logging Road N1 proposed treatment site showing the presence of aspen that could be stimulated to produce new shoots and increase browse production through mechanical treatment of the site. This is in site category 16030-C.

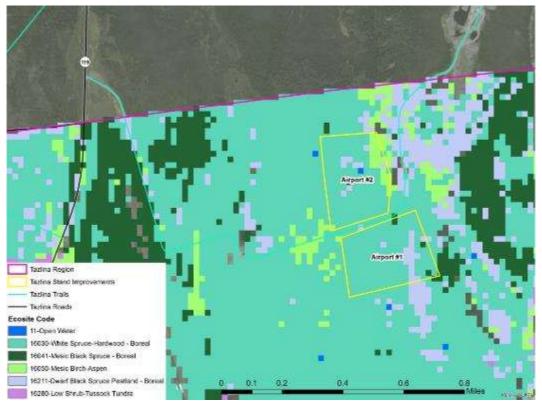


Figure 251. Map of two proposed habitat improvement sites labeled Airport 1 and 2 showing the BpS designations for these areas.

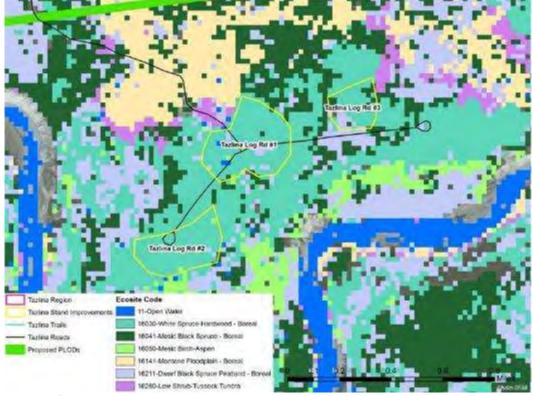


Figure 252. Map of three proposed habitat improvement sites labeled Tazlina Log Road 1, 2, and 3 showing the BpS designations for these areas.



Figure 253. Map of aerial imagery of treatment areas in the North Stand Improvement area termed TAPS North 1, 2, and 3 showing access roads and trails to these sites.



Figure 254. Map showing ownership of proposed treatment areas in the TAPS North Stand Improvement area.

TAZLINA EXISTING AND PROPOSED PRIMARY LINES OF DEFENSE 113 (112) OWNERSHIP Tazlina Region PRIVATE Tazlina Stand Improvements AHTNA Tazlina Trails AHTNA SELECTED Tazlina Roads NATIVE ALLOTMENT PLODs Proposed PLODs STATE OF AK BLM

Figure 255. Map of proposed primary lines of defense (PLODS) surrounding the Tazlina area for wildfire defense.

NPS

Table 34. Treatment site with treatment goal, site size, total tons of biomass and the BpS code and disturbance class for each site.

Site Name	Treatment Goal	Acres	Biomass - Total	BpS Code_ Dist
	Treatment Goal	Aues	(tons)	Class
Ahtna Office #1	Moose Browse	20.0	642.8	16030_C
Airport #1	Moose Browse	38.4	1296.3	16030_C
Airport #2	Moose Browse	37.5	1250.0	16030_C
Fishers Pit #1	Moose Browse	43.6	1361.3	16030_C
Fishers Pit #2	Moose/Timber	35.3	1048.1	16030_C
North Fireline #1	Moose Browse	26.4	316.1	16141_A
North Fireline #2	Moose Browse	30.8	805.9	16211_D
North Fireline #3	Moose Browse	48.6	650.7	16211_D
Taz West Trails #5	Moose Browse	34.7	1254.3	16030_C
Taz West Trails #6	Moose Browse	40.6	1007.1	16211_D
Tazlina Fireline #1	Moose/Timber	31.6	826.3	16030_C
Tazlina Fireline #2	Timber Improvement	31.0	476.4	16211_D
Tazlina Fireline #3	Moose Browse	30.3	1112.5	16030_C
Tazlina Fireline #4	Moose/Timber	28.8	767.3	16030_C
Tazlina Fireline #5	Moose/Timber	29.8	942.1	16030_C
Tazlina Fireline #6	Moose/Timber	33.0	1017.4	16211_D
Tazlina Log Rd #1	Timber Improvement	47.9	1758.6	16030_C
Tazlina Log Rd #2	Moose Browse	34.4	1366.6	16030_C
Tazlina Log Rd #3	Timber Improvement	18.2	647.0	16030_C
Tazlina Pit	Moose Browse	9.9	175.8	16030_C
Tazlina TAPS North #1	Moose Browse	37.3	1282.1	16030_C
Tazlina TAPS North #2	Moose Browse	43.1	1503.6	16030_C
Tazlina TAPS North #3	Moose Browse	32.4	205.0	16280_A
Terrace Drive	Timber Improvement	43.1	1582.4	16030_C

Climate Change

Alaska is experiencing significant impacts from climate change with more extreme changes being noted than many other parts of the United States. Future planning for natural resource management must should consider the potential effects of these changes. While projections of future climate conditions have considerable variance around their estimates, the modelled outputs represent the best science-based assessments of likely conditions. We used SNAP (Scenarios Network for Alaska and Arctic Planning) as the primary source for climate change information (https://www.snap.uaf.edu/tools-and-data/all-analysis-tools). We examined climate projections for 3 locations that span the Ahtna Traditional

Use Territory; Cantwell, Glennallen, and Chitina). Figures 256-261 display the projected changes in climate (monthly temperature and precipitation means) that are predicted to occur under a continued high worldwide production of greenhouse gases (8.5 scenario). While these levels may not be achieved if aggressive response actions are adopted by human society, to date such actions have not been initiated.

Climate projections reveal that South Central Alaska is expected to experience an increase in annual temperatures. This will be especially true in the summer months. Increases in winter temperatures are also expected as indicated by the trends in future temperatures, but the ranges in these estimated projections include potential overlap with historical temperatures. Precipitation shows trends for increases, but these are relatively small and ranges overlap with existing levels. Even if precipitation levels increase, increases in temperatures will result in greater evapotranspiration, especially in the summer, likely producing a drying effect across the landscape.

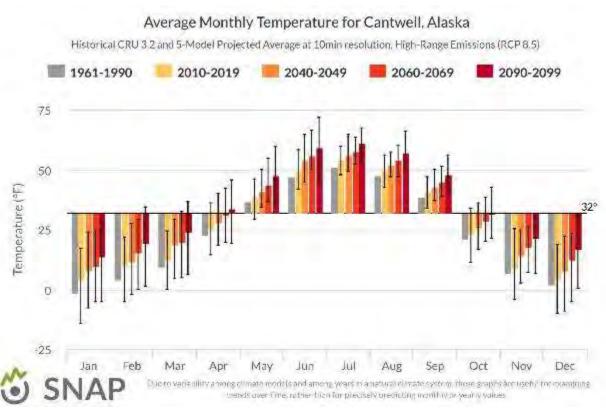


Figure 256. Average monthly temperature projections for Cantwell, Alaska, 2010-2099 (SNAP).

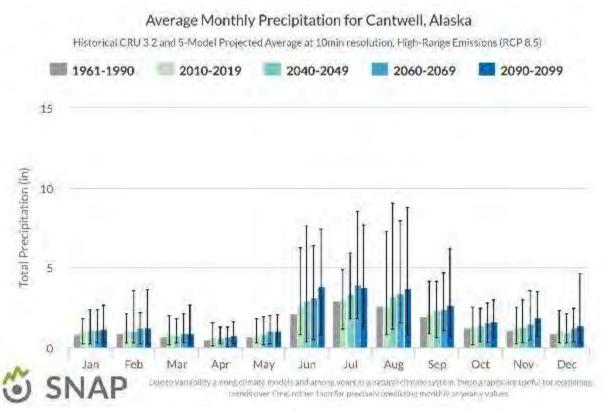


Figure 257. Average monthly precipitation projections for Cantwell, Alaska, 2010-2099 (SNAP).

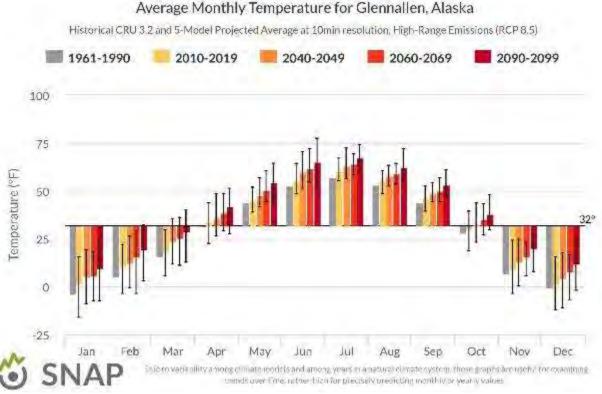


Figure 258. Average monthly temperature projections for Glennallen, Alaska, 2010-2099 (SNAP).

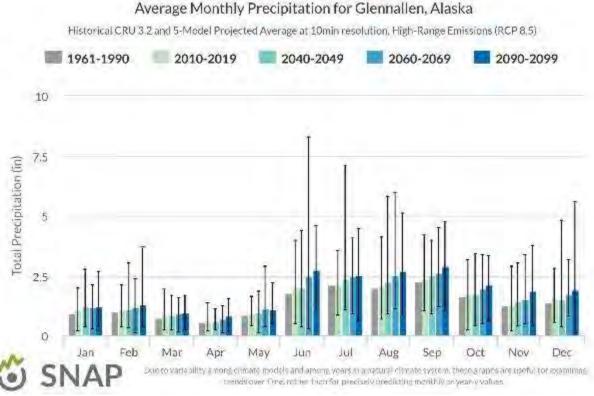


Figure 259. Average monthly precipitation projections for Glennallen, Alaska, 2010-2099 (SNAP).

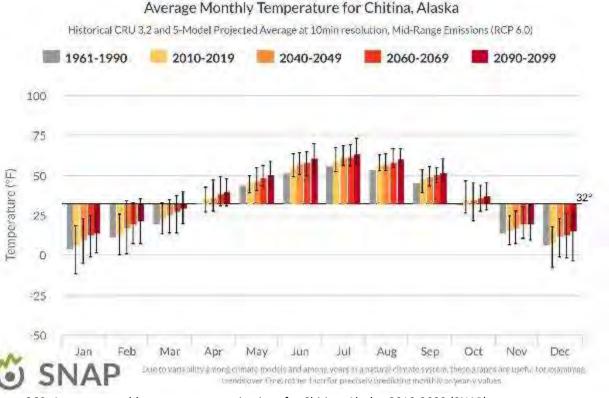


Figure 260. Average monthly temperature projections for Chitina, Alaska, 2010-2099 (SNAP).

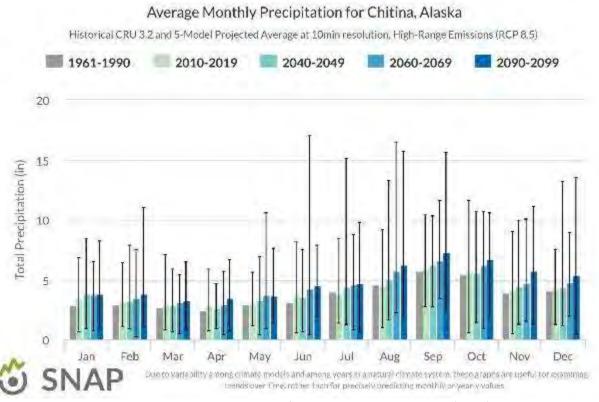


Figure 261. Average monthly precipitation projections for Chitina, Alaska, 2010-2099 (SNAP).

Ecosystem Diversity Trends

Short-term

The interior boreal forest of the Copper River Basin has been shaped by disturbance in the form of fire and influenced by the presence of discontinuous permafrost. In the short term these effects will likely remain as the primary forces shaping ecosystems and influencing succession. The presence of permafrost dictates possible vegetation types starting with open water or emergent wetlands and culminating in dwarf black spruce or tall shrubs. Over short time frames these sites are likely to remain static. Increased summer temperatures and/or increased disturbance in the form of fire could convert sites as the depth to permafrost increases or the permafrost melts entirely. Upland sites are likely to see short term changes due to increased prevalence of insects and disease and increased fire return intervals.

Long-term

The driving force behind long term ecosystem diversity trends is climate change. While this continues to be a developing field of study, current research indicates that over time both summer and winter mean temperatures will rise. Warmer summer temperatures could result in increased fire frequency in forested sites, shrub encroachment in alpine sites, increased flood frequency in riparian sites, and melting of permafrost. Warmer winter temperatures could result in more frequent rain and/or icing events, increased avalanche frequency, increased pests that are normally controlled by cold winter temperatures, and longer growing seasons which would result in increased fine fuel loads and longer fire seasons. Warmer temperatures could also result in decreased surface area of glaciers which may

provide new habitat for forest and shrub ecosystems. Thermokarst is expected to increase significantly, changing many sites with the removal of permafrost and associated inundation and successional processes described previously.

Landscape Level Planning

In addition to planning site treatments around each of the 8 villages, potential broader scale planning objectives were considered. The site treatments in each village had the objectives of improving moose browse production, harvesting biomass, improving stand conditions, or creating a primary line of defense from fire. All of these are envisioned to use mechanical treatments. Improvement of moose habitat from these treatments is unlikely to have any significant influence on moose populations other than to shift their distributions slightly to take advantage of areas with higher browse availability and increase opportunities for subsistence hunting by the villages. Limited amounts of biomass will be produced from these treatments, but can help to provide the villages with wood for fuel. The primary lines of defense will be a factor in community wildfire protection planning. However, when viewed from a broader landscape perspective (Figure 262), it is apparent that the scale of these treatments will not have significant effects on such things as improvement of moose habitat.

At the broader landscape level, other objectives are considered. If increases in overall moose numbers are desired through habitat improvements, larger areas must be treated than those conducted at the village planning level. Mechanical treatments can play a role by providing fuel breaks or fire management lines, but are not thought to be practical for large scale treatments. Large scale treatments will rely on either effects from wildfire managed through various fire response actions, or from use of prescribed fire. Existing fire protection designations (Figure 263) should be reviewed and adjusted if appropriate for desired future management actions

An additional landscape level objective was added in 2016, that of carbon sequestration. Ahtna, Inc. has entered into a carbon sequestration agreement to provide carbon offsets for the California carbon market. This means that amounts of carbon on designated lands owned by Ahtna must be managed to maintain or increase amounts of carbon into the future. Areas included in the carbon agreement are displayed in Figure 264. Immediate objectives for these lands recommend that full wildfire suppression is desired. However, this status will change after 2 years allowing for new management to occur.

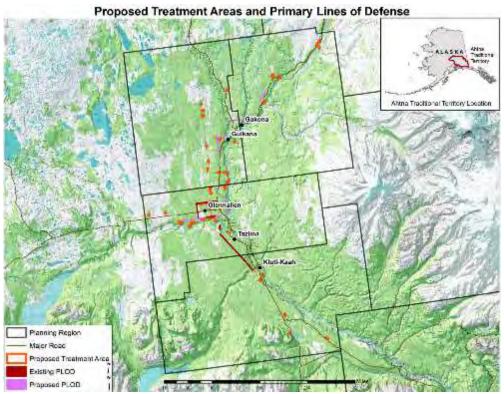


Figure 262. Map of proposed village treatments for the Gakona, Gulcana, Tazlina, and Kluti-Kaah village planning areas.

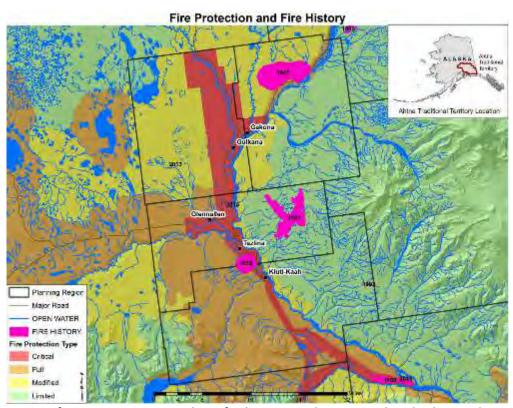


Figure 263. Current fire protection status and past fire locations in the primary Ahtna land ownership portion of the Ahtna Traditional Use Territory.

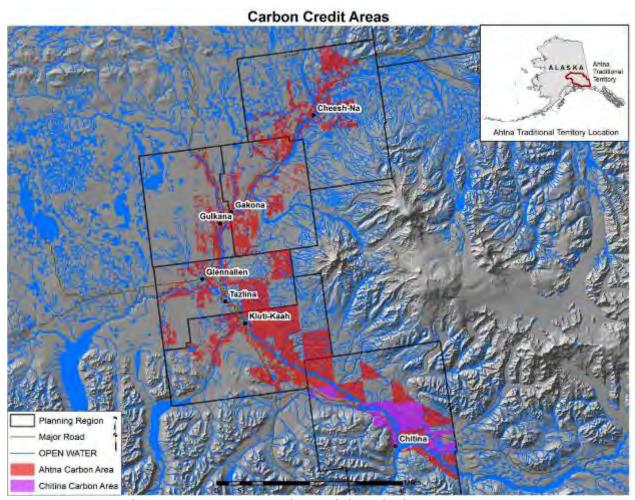


Figure 264. Current carbon sequestration areas on Ahtna and Chitina lands.

Recommendations for Landscape Management

One of the primary objectives of this project was to recommend ways to increase subsistence supply of moose for native villages while maintaining caribou habitat. Moose occur at relatively low densities, so efforts to increase overall population sizes requires improvements to large areas. The habitat treatments recommended for the villages will improve small patches of habitat. These can help a few moose by providing improved foraging opportunities, but will have very limited effects on overall population sizes. The primary function of these treatments is to increase harvest opportunities on existing moose populations by concentrating moose in accessible locations on Ahtna lands. To increase moose populations over larger areas will require much larger scales of treatments. This largely precludes mechanical treatments such as timber harvests or roller chopping from being effective tools except when used in conjunction with other disturbances. Primary recommended tools are selective let burn areas for wildfires and prescribed burning. Use of these tools must integrate with protection of human infrastructure, carbon sequestration goals, maintenance of caribou habitat, agreement from adjacent landowners, and economic viability.

Carbon sequestration can be compatible with moose habitat improvement and biomass harvests when properly coordinated. Some lands contain decadent stands of spruce that hold current carbon in the biomass present on these sites, but are losing this carbon through tree mortality over time. Additional carbon can be sequestered by disturbing some types of sites and encouraging higher site productivity

and sequestration rates. In particular, those BpS's that support productive white or black spruce or aspen hardwood sites but that are currently in late seral, decadent stands can be improved through either mechanical treatments or fire. This can not only result in greater long term carbon sequestration, but can improve moose habitat and in some locations be sources of biomass. Figure 265 displays some areas that may have this potential, that are mapped as BpS's and current successional conditions that may benefit from future treatments to set back succession. If these can be targeted for a combination of mechanical treatments that can produce fuel breaks or defensive lines for fire, areas can then be designated to allow wildfires to burn or for application of prescribed burning. The fire protection zones assigned to such areas should be reviewed to determine if adjustments to these zones are needed to integrate with the potential treatment zones. Figure 266 provides a closer look at potential sites near the village of Kluti-Kaah. Figures 267-270 show potential areas to increase the quality of moose habitat and also show other resource values that may be impacted by management activities in those areas. Figures 268-270 focus on one potential area as an example.

It is beyond the scope of this project to propose and evaluate specific landscape level treatment zones. However, the GIS data and maps provide starting points for identifying potential treatment areas as shown in Figures 265-270. Such areas should be checked for their other resource values, such as high quality berry production areas or caribou habitat (Figure 270), and then considered for further treatments. All proposed treatment zones should receive ground verification before assuming the mapped existing conditions are accurate.

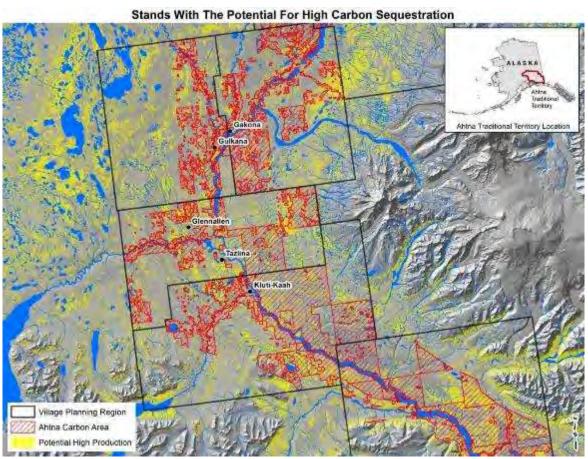


Figure 265. Potential areas for improving moose winter habitat and increasing carbon sequestration. Potential high production means a stand is currently in a late successional state, but could be returned to high production through management activities.

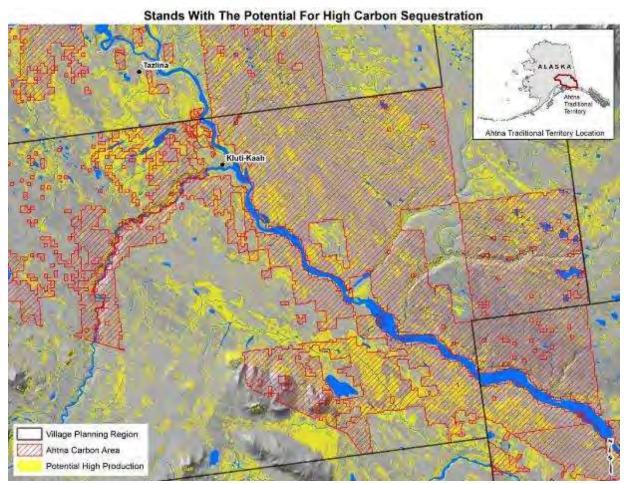


Figure 266. Potential areas for improving moose winter habitat and increasing carbon sequestration. Potential high production means a stand is currently in a late successional state, but could be returned to high production through management activities. This figure provides a closer view of the region south of Kluti-Kaah.

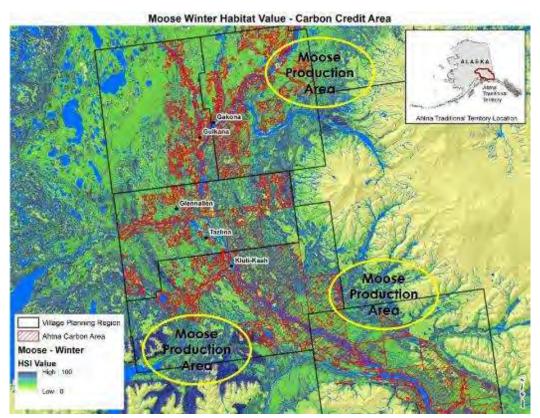


Figure 267. Potential areas for improving moose winter habitat along with carbon sequestration boundary.

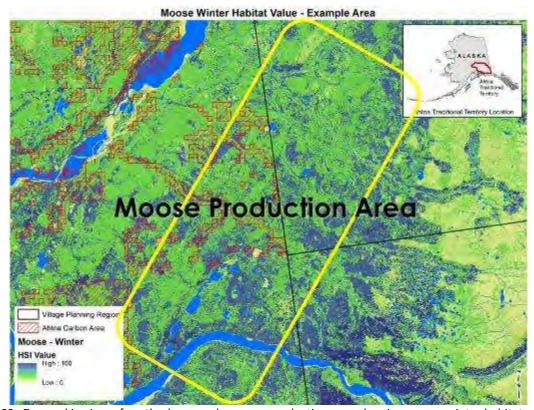


Figure 268. Zoomed in view of northerly example moose production area showing moose winter habitat quality and carbon sequestration stands.

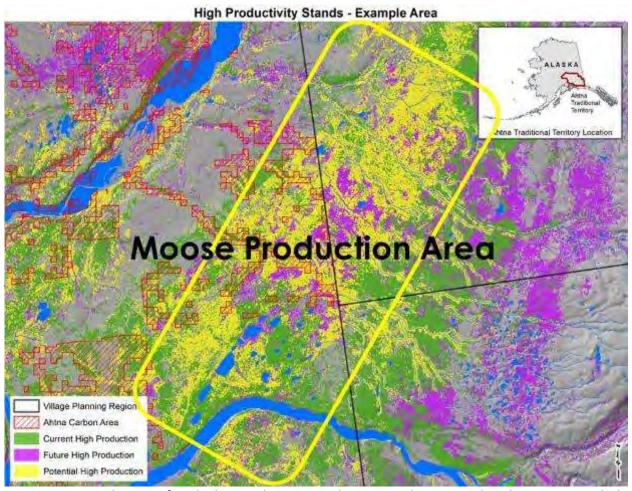


Figure 269. Zoomed in view of northerly example moose production area showing stand productivity. Future high production means the stand will increase in productivity as it ages. Potential high production means the stand could be returned to high production through management activities.

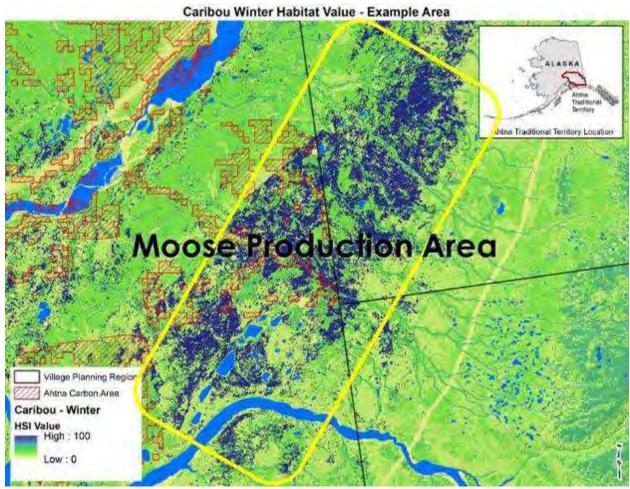


Figure 270. Zoomed in view of northerly example moose production area showing caribou winter habitat quality and carbon sequestration stands. Improving moose habitat quality is likely to reduce caribou habitat quality, so these conflicting objectives will need to be integrated in treatment plans.

Literature Cited

- Barney, R. J. 1969. Interior Alaska wildfires, 1956-1965. USDA Forest Service, Pacific NW Forest and Range Experiment Station, Institute of Northern Forestry, Juneau, Alaska, USA.
- Barrett, K., A. D. McGuire, E. E. Hoy, and E. S. Kasischke. 2011. Potential shifts in dominant forest cover in interior Alaska driven by variations in fire severity. Ecological Applications 21:2380-2396.
- Bernhardt, E. L., T. N. Hollingsworth, and F. S. Chapin, III. 2011. Fire severity mediates climate-driven shifts in understorey community composition of black spruce stands of interior Alaska. Journal of Vegetation Science 22:32-44.
- Boggs, K. 2000. Classification of community types, successional sequences, and landscapes of the Copper River Delta, Alaska. USDA Forest Service, Pacific Northwest Research Station. PNW-GTR-469. Portland, Oregon, USA.
- Clark, M. H., and D. R. Kautz. 1999. Soil Survey of Copper River Area, Alaska. USDA Natural ResourceConservation Service.
- Clark, M. H., and M. S. Duffy. 2006. Soil Survey of Denali National Park, Alaska. USDANatural Resource Conservation Service.
- DeVelice, R. L., C. J. Hubbard, C.J., K. Boggs, S. Boudreau, M. Potkin, T. Boucher, and C. Wertheim. 1999. Plant community types of the Chugach National Forest: southcentral Alaska. USDA Forest Service, Technical Publication R10-TP-76, Anchorage, Alaska, USA.
- Ferrians, O. J., D. R. Nichols, and J. R. Williams. 1983. Copper River Basin, résumé of quaternary geology. Page 137 in T. L. Péwé and R. D. Reger, editors. Guidebook to permafrost and quaternary geology along the Richardson and Glenn Highways between Fairbanks and Anchorage, Alaska. Volume 1. Alaska Dept. of Natural Resources.
- Hollingsworth, T. N., M. D. Walker, F. S. Chapin, and A. L. Parsons. 2006. Scale-dependent environmental controls over species composition in Alaskan black spruce communities. Canadian Journal of Forest Research-Revue Canadienne De Recherche Forestiere 36:1781-1796. Johnstone, J. F., T. N. Hollingsworth, F. S. Chapin, III, and M. C. Mack. 2009. Changes in fire regime break the legacy lock on successional trajectories in Alaskan boreal forest. Global Change Biology 16:1281-1295.
- Johnstone, J. F., T. S. Rupp, M. A. Olson, and D. Verbyla. 2011. Modeling impacts of fire severity on successional trajectories and future fire behavior in Alaskan boreal forests. Landscape Ecology 26:487-500.
- Jorgenson, M. T., J. E. Roth, S. F. Schlentner, E. R. Pullman, M. Macander, and C. H Racine. 2003. An ecological land survey for Fort Richardson, Alaska. ERDC/CRREL TR-03019. Cold Regions Research and Engineering Laboratory, Hanover, New Hampshire, USA.
- Jorgenson, M. T., J. E. Roth, M. Raynolds, M. D. Smith, W. Lentz, A. Zusi-Cobb and C. H. Racine. 1999. Anecological land survey for Fort Wainwright, Alaska. CRREL Report 99-9. U.S. Army Cold Regions Research and Engineering Laboratory, Hanover, New Hampshire, USA.
- LANDFIRE Vegetation Dynamics Models. (2009, November last update). Homepage of the LANDFIRE Project, U.S. Department of Agriculture, Forest Service; U.S. Department of Interior, [Online]. Available: http://www.landfire.gov/index.php [2015, August 17].
- Lynch, J. A., J. L. Hollis, and F. S. Hu. 2004. Climatic and landscape controls of the boreal forest fire regime: Holocene records from Alaska. Ecology 92:477-489.
- Mann, D. H., T. S. Rupp, M. A. Olson, and P. A. Duffy. 2012. Is Alaska's boreal forest now crossing a major ecological threshold? Arctic, Antarctic, and Alpine Research 44:319-331.
- Murry, D. F., B. M. Murray, B. A. Yurtsev, and R. Rowenstein. 1983. Biogeographic significance of steppe vegetation in subarctic Alaska. Pages 883-888 *in* Proceedings of Permafrost Fourth International Conference. 17-22 July, 1983. Fairbanks, Alaska, USA.
- NatureServe. 2008. International ecological classification standard: terrestrial ecological classifications. Draft Ecological Systems Description for Alaska Boreal and Sub-boreal Regions. Arlington, Virginia, USA.
- Patric, J. H., and P. E. Black. 1968. Potential evapotranspiration and climate in Alaska by Thornthwaite's Classification. General Technical Report PNW-71. USDA Forest Service, Juneau, Alaska, USA.

- Pegau, R.E. 1972. Caribou investigations-analysis of range. *In* R. E. Pegau and J.E. Hemming, editors.Caribou report. Volume 12. Progress report. Federal Aid in Wildlife Restoration, Alaska Dept. of Fish and Game, Juneau
- Peters, V. S., S. E. Macdonald and M. R. T. Dale. 2005. The interaction between masting and fire is key to white spruce regeneration. Ecology 86: 1744-1750.
- Scheffer, M., M. Hirota, M. Holmgren, E. H. Van Nes, and F. S. Chapin III. 2012. Thresholds for boreal biome transitions. Proceedings of the National Academy of Sciences of the United States of America 109:21384-21389.
- Scott, R. W. 1974. Successional patterns on moraines and outwash of the Frederika Glacier, Alaska. Pages 319-329 in V. C. Bushnell and M.G. Marcus, editors. Icefield Ranges Research Project, Scientific Results, Volume 4.

 American Geographical Society and Arctic Institute of North America, New York, USA.
- Shenoy, A., J. F. Johnstone, E. S. Kasischke, and K. Kielland. 2011. Forest ecology and management 261:381-390.
- Shephard, M.E. 1995. Plant community ecology and classification of the Yakutat Foreland, Alaska. TP-56. USDA Forest Service, Tongass National Forest, Sitka AK.
- Tarr, R. S., and L. Martin. 1913. Glacial deposits of the continental type in Alaska. The Journal of Geology 21: 289-300.
- Thilenius, J. F. 1990. Woody plant succession on earthquake-uplifted coastal wetlands of the Copper River Delta, Alaska. Forest Ecology and Management 33/34: 439-462.
- Turetsky, M. R., M. C. Mack, T. N. Hollingsworth, and J. W. Harden. 2010. The role of mosses in ecosystem succession and function in Alaska's boreal forest. Canadian Journal of Forestry Research 40:1237-1264.
- Van Cleve, K., C. T. Dyrness, L. A. Viereck, J. Fox, F. S. Chapin, III, and W. Oechel. 1983. Taiga ecosystems in Interior Alaska. BioScience 33:39-44.
- Viereck L. A. 1966. Plant succession and soil development on gravel outwash on the Muldrow Glacier, Alaska. Ecological Monographs 36:181-199.
- Viereck, L. A. 1979. Characteristics of treeline plant communities in Alaska. Holarctic Ecology 2:228-238.
- Viereck, L. A., and L. A. Schandlemeier. 1980. Effects of fire in Alaska and adjacent Canada a literature review. BLM-AK Tech. Rep. 6. U.S. Department of the Interior, Bureau of Land Management.
- Viereck, L. A., C. T. Dyrness, A. R. Batten, K. J. Wenzlick. 1992. The Alaska vegetation classification. USDA Forest Service, Pacific Northwest Research Station. PNW-GTR-286. Portland, Oregon.

Appendix A. Moose Habitat Model for South Central Alaska Introduction

The Copper River-Ahtna Intertribal Conservation District (CRITR) has initiated a landscape scale project funded through the Conservation Innovation Grant program of the Natural Resources Conservation Service. The purpose of the project is to develop innovative tools and to increase the technical capacity for planning by CRITR. The specific objectives include:

- Conduct an ecologically based resource assessment of Ahtna lands; develop ecological site classification, and develop site-specific vegetation treatments;
- Develop moose and caribou habitat models to support habitat management of these species;
- Develop an innovative 10-year management plan for Ahtna's 1.7 million acres to increase moose for food and biomass for energy while maintaining or improving caribou habitat; and
- Train local technicians to conduct habitat treatments and monitor results.

A moose habitat model will be an important tool to help identify sites with the best potential for improving moose habitat and to incorporate management of these sites into an overall plan that considers a landscape context.

Moose Distributions

Alaskan moose (*Alces alces gigas*) are the largest in size of 4 subspecies of moose in North America. Moose distributions in south-central Alaska, as determined by Alaska Department of Fish and Game, are shown in Figure 1.

Moose Habitat Requirements

Moose habitat requirements vary throughout the year with greatest consideration given to winter, spring, and summer habitat. All three seasons are influenced by the availability of preferred foods as well as avoidance of predation risk and disturbance from human activities, and selection of thermal cover in both winter and summer.

Moose select areas providing them with a mix of food and cover (Maier et al. 2005). Moose rely heavily on willows throughout the year (MacCracken et al. 1997). Other foods include sweetgale (*Myrica gale*), Sitka alder (*Alnus sinuata*), and emergent aquatic plants like marsh fivefinger (*Potentilla palustris*), horsetails (*Equisetum* spp.) and buckbean (*Menyanthes trifoliate*) (MacCracken et al. 1997). Habitat selection by moose has been shown to be influenced by the scale of analysis. For example, a study in Norway showed that at large scales, moose selected areas that contained higher percentages of preferred habitat types, while at the scale of the home range, smaller home ranges contained higher percentages of the preferred habitat types than larger home ranges (Herfindal et al. 2009). However, they found at the home range scale that the preference for preferred habitat types was not found because the home ranges were selected in areas containing these preferred types.

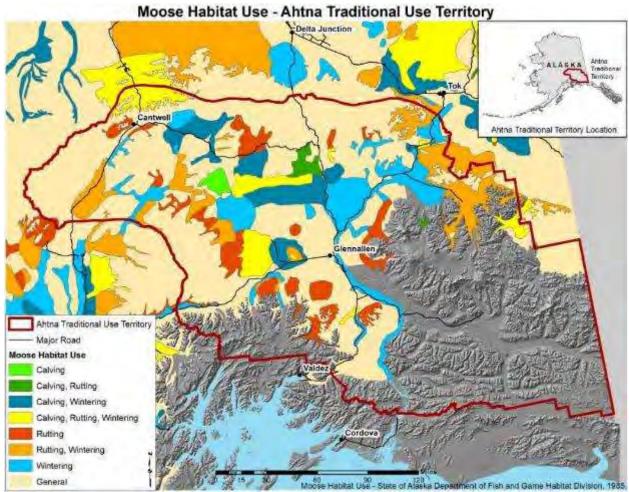


Figure 1. Map of existing moose distribution in southcentral Alaska (from Alaska Department of Fish and Game).

Winter Habitat

Collins and Helm (1997) studied moose winter habitat selection in a floodplain in south-central Alaska and found that moose selected areas that provided high amounts of browse with feltleaf willow (*Salix alaxensis*) the preferred food. Highbush cranberry (*Viburnum edule*) and rose (*Rosa acicularis*) were also present on selected sites, but were not available as browse when snow depths exceeded 110 cm. They noted that flood action in the floodplain was the primary disturbance that maintained preferred early succession conditions. They recommended that moose habitat management focus on upland areas where fire or other disturbances were needed to produce increased amounts of early successional habitat. LeResche and Davis (1973) reported moose using burned areas for up to 50 years with greatest use occurring 20-25 years post burn. Maier et al. (2005) found that moose utilized areas 11-30 year post burn, and had higher densities in areas with mixed vegetation conditions, and avoided areas of mixed terrain or lacking vegetation. Puttock et al. (1996) found that moose selected stands that were 1-20 years of age as well as stands with <30% canopy cover in Ontario. Moose in southeast Alaska used riparian shrub and high volume coniferous forests during thick snow, and were found to use clearcuts <30 years old more heavily than unlogged forest (Doerr 1983).

Maier et al. (2005) looked at moose distributions in early winter, and found that moose selected areas that had a diversity of vegetation conditions but did not select areas that were variable in terrain. They particularly selected areas that had burned 11-30 years previously.

Moose have been found to respond to snow depths. Nietfeld et al. (1985) reported that moose avoided areas where snow depths exceeded 65-75 cm. Eastman (1977) reported that moose increased use of mature conifer stands when snow depths approached 80 cm. Schwab and Pitt (1991) reported that conifer stands should have >70% cover to provide optimal snow reductions and thermal cover in the winter. Poole and Stuart-Smith (2006) reported that moose occurred at lower elevations in their study area in British Columbia as the winter progressed which they attributed to moose selecting areas with lower snow depths. They also found that moose selected areas with more gentle terrain and that offered solar radiation. Dussault et al. (2005b) studied moose in Quebec and found that in winter moose avoided areas with the lowest snow depths, a response they attributed to avoiding wolves. In their study, moose selected areas that provided shelter from snow in close proximity to foraging areas. In a related analysis, Dussault et al. (2005a) found that moose home ranges were smaller in winter where good food resources occurred, but they did not observe this in summer. Moose movements were less in areas with good supply of food in both winter and summer (Dussault et al. 2005a). Leclerc et al. (2012b) also studied moose in Quebec and found that moose selected harvested areas in winter where forage was available, but needed a mix of cover and foraging areas.

Spring/Calving Habitat

Chekchak et al. (1998) identified moose calving sites in Quebec. They found that moose selected sites on tops of hills with gentle slopes. Bowyer et al. (1999) similarly looked at calving sites of moose in Alaska. They found that moose selected sites higher on hills with dry southerly exposures as preferred calving locations where females had some visibility so that predators could be observed and avoided, but that also supported some willow, thus balancing risk of predation with food availability.

Several studies found that females with calves selected areas in the spring/early summer that contained forest cover to provide predator avoidance (Miquelle et al. 1992, White and Berger 2001, Oehlers et al. 2011). In contrast, both of these studies found that males selected areas that maximized forage production at this time of the year without needing forest cover, as predation was apparently not a driver of habitat selection at this time of the year for this segment of the population.

Summer Habitat

Forage, escape cover, and thermal cover are habitat needs of moose in summer. Moose diets have been noted to expand in the summer with the availability of additional foods such as aquatic plants (MacCracken et al. 1993). Areas providing both forage and cover were selected by moose in Norway, with moose using areas containing more preferred forage and cover types having smaller home ranges (Bjorneraas et al. 2012).

Demarchi and Bunnel (1995) found that moose selected thermal cover during times of heat stress in British Columbia. They found that moose selected areas with greater forest cover when conditions of heat stress during certain times of the day existed in open areas. Dussault et al. (2004) reported that moose selected thermal cover in summer when air temperatures got high, and switched their activity patterns to occur more at night than during the day. McCann et al. (2013) monitored captive moose responses to summer temperatures and found that moose became stressed at around 17° C in calm conditions, but this increased to 24° C when wind provided some cooling. They noted that moose selected shady areas during hot weather when the sun was out, indicating that solar radiation is also a

factor in heat stress. Melin et al. (2014) reported that there was a strong selection by moose for thermal cover containing higher and denser canopies when day time temperatures exceeded 20° C. Similarly, van Beest et al. (2012) found that moose selected closed canopy conifer cover when temperatures got higher in summer, reducing use of more open areas and thus reducing their ability to find forage. However, Lowe et al. (2010) did not find habitat selection by moose to avoid higher temperatures in a study in southern Ontario.

Anthropogenic Disturbance and Habitat Changes

Eldegard et al. (2012) found that moose in Norway appeared to balance their selection for high levels of browse with proximity to roads, noting that moose preferred sites with high amounts of browse as well as sites farther from roads. They also found that moose moved closer to smaller roads with lower amounts of traffic than larger roads with more traffic. Beyer et al. (2013) investigated moose occurrence and movements in relation to roads in Ontario. They found that moose occurred in areas with roads at a landscape scale, as roads were linked with timber harvests that produced early successional habitat favored by moose, but that moose avoided roads based on analysis at a finer scale. Leclerc et al. (2012a) found that moose abundance decreased with road density. Dussault et al. (2007) found that moose movement rates increased when crossing roads and that the area around roads was perceived as low quality habitat by moose.

Laurian et al. (2012) examined moose locations in relation to roads in Quebec and found that moose searched for areas containing high amounts of forage while avoiding highways and roads, but that this avoidance was only for 100-250m. In a related article (Laurian et al. 2008) that tracked nearly 200,000 moose movement segments, only 328 crossed highways and 1,172 crossed forests roads which were 16 and 10 times less than by chance. They did note that while moose generally avoided highways, some moose did use areas along highways for foraging which they attributed to selection for sodium-rich foods. In fact, browsing rates along highways were equal to those more distant from highways even though use by moose as measured by time or occurrence was substantially less.

Harris et al. (2014) evaluated the effects of winter recreation on moose, focusing on snowmobile use. They found that snowmobile use affects moose when it is unpredictable, spans large areas, is long in duration, has a large spatial footprint, involves non-motorized use, and when animals are displaced into poor quality habitat. Shanley and Pyare (2011) evaluated moose distributions in summer and fall in relation to OHV use and found that even relatively low levels of OHV use (<0.25 km travel/km²/day) elicited a response with males affected up to 1000m away from a trail and females affected even more than 1000m away.

Moose Mortality Factors

The relationship of moose population sizes as influenced by habitat qualities and mortality factors, in particular predation by wolves and bears and human harvest, has been and continues to be evaluated. Ballard et al. (1991) studied moose in southcentral Alaska from 1976-1986 found high pregnancy rates (81%) and twinning rates (38%) in this population, but only a 39% rate of calf survival through 5 months with 83% of loss caused by predation of which 96% occurred during the first 6 weeks of life and 73% of this loss attributed to brown bears. McCracken et al. (1997) found calf production on the Copper River Delta to average 1 calf/cow, but fall cow/calf ratios averaged 30 calves/100 cows. Mortality of calves was due to cold spring weather and predation by brown bears.

Existing Moose Habitat Models

Moose habitat requirements have shown many similarities across the range of the species. However, Mabille et al. (2012) cautioned that functional habitat selection by moose can vary locally depending on the types of conditions occurring at the location, so extrapolations about habitat selection or use from other areas must be treated cautiously. Similarly, McLoughlin et al. (2011) found differences in seasonal habitat selection (in response to hunting) and road avoidance between two nearby study areas that had substantially different management programs, while Osko et al. (2004) found different selection of habitat by two groupings of moose based on availability of habitat classes. Further, sex and reproductive status have been shown to influence seasonal habitat selection by moose (Miquelle et al. 1992). They found that females with calves selected areas in early summer that provided forest cover, apparently as a predator avoidance for calves, while males selected areas specifically for high forage production in Denali National Park. Some additional sexual separation was noted in the Denali study at other times of the year, but it was not as noticeable as during early summer. A study conducted in Tongass National Forest (Oehlers et al. 2011) found similar differences in habitat selection by males and females with calves during the spring. Thus, a generic habitat model for moose should generally characterize habitat for the species, but some differences in habitat selection may be missed or masked, especially when considering spring (calving) habitat requirements. Dettki et al. (2003) modeled moose habitat quality related to vegetation and compared this to an empirically driven model that computed a number of environmental variables and found substantial differences, stressing the importance of incorporating environmental variables (e.g., elevation) into models.

Various habitat models have been developed for moose, but need to be carefully evaluated for their application to moose in southcentral Alaska. A habitat suitability model was developed for moose in Quebec (Dussault et al. 2006). They included two variables in the model, food and an interspersion measurement, and evaluated the model at different scales. The model that integrated the two variables worked well for wintering males at a scale of 500 ha, but not for females. Females responded to the food variable, but not the interspersion variable. Female habitat use was found to correspond better to forage at home range scales measured in 100 ha polygons than compared to the 500 ha scale.

A habitat suitability model for moose in the Lake Superior area was developed by Allen et al. (1987). This model incorporated measures of summer and winter browse, wetland areas for summer food, and winter cover as variables. It evaluated winter based on the provision of food as modified by proximity to cover (within 100 m of cover). This model was then put into a GIS framework using remotely sensed mapping (Hepinstall et al. 1996). This application suggested that a 50% overlap in a moving window analysis was sufficient to capture landscape variation.

Several models have been developed as part of impact evaluations for oil and gas or other developments in Canada. These models have generally used vegetation classifications and associated rankings of moose habitat that have very limited application to a landscape-scale for south-central Alaska.

Habitat Model for Moose in Southcentral Alaska Landscape Assessment

At a landscape scale, moose habitat will be considered to be important within mapped moose range (Figure 1). Within existing moose range, moose habitat quality will be rated according to the procedures outlined below. A moving window analysis is used to evaluate the quality of an area of approximately 10,000 ha (24,000 ac) surrounding each pixel as indicated in Table 1.

Table 1. Landscape scale rating of habitat quality based on aggregate quality of winter, spring and summer habitat within a 10,000 ha area.

Percentage of 10,000 ha area with	Area weighted HSI value
HSI > 0.75	For winter, spring and summer habitat
>50%	1.0
25-50%	0.5
10-25%	0.25
<10%	0.1

Stand Assessment

The existing moose range was mapped for each biophysical setting/structural class combination which is then assigned a habitat quality value for winter, spring, and summer foraging habitat and thermal/escape cover value (Table 2). This value will then be further modified based on an interspersion variable for edges. In addition, an adjustment for terrain effects on snow depth as an influence on winter habitat quality could be added, but accurately assessing potential snow depths may not be feasible, so this variable has not been included in the current model. Table 2 lists the value of each biophysical setting/structural state. Table 3 shows the complete name of each biophysical setting as represented by the BpS code.

Table 2. HSI Scores by biophysical setting (BpS) and Structure class for moose in the Ahtna Traditional Use Region.

	Structure	Winter	Spring	Summer	Winter	Spring	Summer
BpS ¹	Class ²	Foraging	Foraging	Foraging	Cover	Cover	Cover
16030	Α	0.25	0.25	0.75	0.10	0.25	0.25
16030	В	0.25	0.25	0.75	0.10	0.25	0.25
16030	С	0.75	0.75	0.75	0.25	0.50	0.50
16030	D	0.25	0.50	0.50	0.50	0.75	0.75
16030	Е	0.10	0.10	0.10	1.00	0.50	0.75
16790	Α	0.25	0.25	1.00	0.10	0.10	0.10
16790	В	0.75	0.75	0.75	0.25	0.50	0.75
16790	С	0.50	0.50	0.50	0.10	0.25	0.50
16790	D	0.25	0.50	0.50	0.50	0.75	1.00
16790	Е	0.10	0.25	0.25	0.75	1.00	1.00
16041	Α	0.75	0.75	1.00	0.10	0.25	0.25
16041	В	0.75	0.75	1.00	0.10	0.25	0.25
16041	С	0.75	0.75	0.75	0.25	0.50	0.50
16041	D	0.25	0.50	0.50	0.75	0.75	0.75
16041	Е	0.25	0.50	0.50	0.75	0.50	0.75
16042	Α	0.25	0.50	0.10	0.10	0.10	0.10
16042	В	0.25	0.50	0.50	0.50	0.50	0.50
16042	С	0.75	0.75	0.75	0.25	0.50	0.75
16042	D	0.25	0.50	0.50	0.50	0.75	1.00
16042	E	0.10	0.10	0.10	1.00	1.00	1.00
16011	Α	1.00	1.00	1.00	0.10	0.25	0.25

16011	В	1.00	1.00	1.00	0.50	0.75	0.75
16011	С	1.00	0.25	0.25	0.75	0.75	0.75
16012	Α	1.00	1.00	1.00	0.10	0.25	0.25
16012	В	1.00	1.00	1.00	0.50	0.75	0.75
16012	С	1.00	0.25	0.25	0.75	0.75	0.75
16050	Α	0.25	0.50	1.00	0.10	0.25	0.25
16050	В	0.25	0.50	1.00	0.10	0.25	0.25
16050	С	0.75	0.75	0.75	0.25	0.50	1.00
16050	D	0.75	0.75	0.75	0.25	0.50	0.75
16050	E	0.50	0.50	0.50	0.25	0.75	1.00
16070	Α	1.00	1.00	1.00	0.10	0.25	0.25
16070	В	0.50	0.50	0.75	0.25	0.50	1.00
16061	Α	0.75	0.75	1.00	0.10	0.25	0.25
16061	В	0.75	0.75	1.00	0.10	0.25	0.25
16061	С	0.25	0.25	0.50	0.25	0.50	0.75
16061	D	0.10	0.10	0.25	0.50	0.75	0.75
16141	Α	0.25	0.75	1.00	0.10	0.10	0.10
16141	В	1.00	1.00	1.00	0.25	0.50	0.75
16141	С	0.50	0.50	0.75	0.50	0.75	1.00
16141	D	0.25	0.25	0.25	0.50	0.50	0.50
16141	E	0.10	0.10	0.10	1.00	0.75	1.00
16142	Α	0.25	0.75	1.00	0.10	0.10	0.10
16142	В	1.00	1.00	1.00	0.25	0.50	0.75
16142	С	0.50	0.50	0.75	0.50	0.75	1.00
16142	D	0.25	0.25	0.25	0.50	0.50	0.50
16142	E	0.10	0.10	0.10	1.00	0.75	1.00
16150	Α	0.25	0.75	1.00	0.10	0.10	0.10
16150	В	1.00	1.00	1.00	0.25	0.50	0.75
16150	С	0.50	0.50	0.75	0.50	0.75	1.00
16150	D	0.25	0.25	0.25	0.50	0.50	0.50
16150	E	0.10	0.10	0.10	1.00	0.75	1.00
16160	Α	1.00	1.00	1.00	0.10	0.50	0.75
16160	В	0.75	0.75	0.75	0.50	0.75	0.75
16160	С	0.25	0.25	0.25	0.75	0.75	1.00
16211	Α	0.10	0.75	1.00	0.10	0.50	0.75
16211	В	0.10	0.50	0.75	0.10	0.50	0.75
16211	С	0.50	0.50	0.50	0.10	0.10	0.25
16211	D	0.25	0.50	0.50	0.50	0.50	0.50
16212	Α	0.10	0.50	0.75	0.10	0.50	0.75
16212	В	0.10	0.50	0.75	0.10	0.50	0.75
16212	С	0.25	0.25	0.50	0.10	0.10	0.25
16220	Α	0.25	0.50	0.50	0.25	0.50	0.50
16220	В	0.25	0.50	0.50	0.25	0.50	0.50

16220	С	0.10	0.10	0.10	1.00	0.50	1.00
16220	D	0.25	0.25	0.25	0.75	0.75	0.75
16300	Α	0.10	0.25	0.25	0.10	0.10	0.10
16300	В	0.10	0.25	0.25	0.10	0.10	0.10
16300	С	0.10	0.25	0.25	0.25	0.25	0.25
16102	Α	0.10	0.25	0.25	0.10	0.10	0.10
16102	В	1.00	1.00	1.00	0.10	0.25	0.75
16280	Α	0.10	0.25	0.25	0.10	0.10	0.10
16280	В	0.75	0.75	0.75	0.25	0.50	0.75
16280	С	0.50	0.50	0.50	0.10	0.25	0.50
16351	Α	0.10	0.50	0.50	0.10	0.10	0.10
16310	Α	0.10	0.50	0.50	0.10	0.10	0.10
16290	Α	0.10	0.10	0.10	0.10	0.10	0.10
16290	В	0.10	0.25	0.25	0.10	0.10	0.10
16330	Α	0.10	0.50	0.50	0.10	0.10	0.10
16110	Α	0.10	0.25	0.25	0.10	0.10	0.10
16120	Α	0.10	0.25	0.25	0.10	0.10	0.10
16080	Α	0.10	0.25	0.25	0.10	0.10	0.10
16080	В	0.50	0.50	0.50	0.10	0.25	0.75
16090	Α	0.10	0.25	0.25	0.10	0.10	0.10
16090	В	0.50	0.50	0.50	0.10	0.50	0.75
16520	Α	0.10	0.25	0.25	0.10	0.10	0.10
16520	В	0.50	0.50	0.50	0.10	0.25	0.75
16430	Α	0.10	0.50	0.50	0.10	0.10	0.10
16170	Α	0.10	0.50	1.00	0.10	0.25	1.00
16170	В	0.10	0.50	1.00	0.10	0.25	0.75
16170	С	0.25	0.50	0.75	0.10	0.25	0.25
16170	D	0.50	0.75	0.75	0.10	0.25	0.25
16170	E	0.10	0.25	0.25	0.10	0.10	0.10
16181	Α	0.10	0.50	1.00	0.10	0.25	0.75
16181	В	0.10	0.50	1.00	0.10	0.25	0.50
16181	С	0.25	0.50	0.75	0.10	0.25	0.25
16181	D	0.50	0.75	0.75	0.10	0.25	0.25
16372	Α	0.10	0.25	0.50	0.10	0.10	0.10
16372	В	0.75	0.75	0.75	0.10	0.25	0.25
16372	С	0.75	0.75	0.75	0.10	0.10	0.10
16240	Α	0.75	0.75	0.75	0.10	0.50	1.00
16481	Α	0.25	0.25	0.50	0.10	0.10	0.10
16481	В	0.10	0.10	0.25	0.50	0.75	1.00
16481	С	0.10	0.10	0.10	1.00	1.00	1.00
16460	Α	0.25	0.25	0.50	0.10	0.10	0.10
16460	В	0.10	0.10	0.25	0.10	0.25	0.50
16460	С	0.10	0.10	0.10	0.75	1.00	1.00

16460	D	0.10	0.10	0.10	0.75	1.00	1.00
16440	Α	0.10	0.10	0.25	0.50	0.75	0.75
16500	Α	0.25	0.25	0.50	0.10	0.25	0.50
16500	В	0.25	0.25	0.50	0.10	0.25	0.50
16500	С	0.75	0.75	0.75	0.50	0.75	0.75
16500	D	0.10	0.10	0.25	0.50	0.50	1.00
16550	Α	0.10	0.10	0.25	0.25	0.50	0.75
16550	В	0.10	0.10	0.25	0.25	0.50	0.75
16550	С	0.10	0.10	0.25	0.50	0.50	0.75
16550	D	0.10	0.10	0.25	0.75	0.75	0.75
16550	E	0.10	0.10	0.25	1.00	1.00	1.00
16590	Α	0.10	0.10	0.25	0.25	0.25	0.25
16450	Α	0.10	0.50	0.50	0.10	0.10	0.10
16800	Α	0.10	0.50	0.50	0.10	0.10	0.10
16800	В	0.50	0.50	0.50	0.10	0.25	0.75
16620	Α	0.10	0.50	1.00	0.10	0.10	0.10

¹BpS code names

Table 3. Biophysical setting (BpS) setting names in the Ahtna Traditional Use Region of Southcentral Alaska.

BpS	EDM_Name
16011	Treeline White Spruce - Boreal
16012	Treeline White Spruce - SubBoreal
16030	White Spruce-Hardwood - Boreal
16041	Mesic Black Spruce - Boreal
16042	Mesic Black Spruce - SubBoreal
16050	Mesic Birch-Aspen
16061	Dry Aspen-Steppe Bluff
16070	Subalpine Balsam Poplar-Aspen
16080	Avalanche Slope Shrubland
16090	Mesic Subalpine Alder
16102	Mesic Scrub Birch/Willow
16110	Mesic Bluejoint Meadow
16120	Dry Grassland
16141	Montane Floodplain - Boreal
16142	Montane Floodplain - Subboreal
16150	Large River Floodplain
16160	Riparian Stringer
16170	Shrub and Herbaceous Floodplain
16181	Herbaceous Fen
16211	Dwarf Black Spruce Peatland - Boreal
16212	Dwarf Black Spruce Peatland - Subboreal
16220	Black Spruce Wet-Mesic Slope

16240	Deciduous Shrub Swamp
16280	Low Shrub-Tussock Tundra
16290	Tussock Tundra
16300	Wet Black Spruce-Tussock
16310	Alpine Dwarf-Shrub Summit
16330	Alpine Mesic Herbaceous Meadow
16351	Alpine Ericaceous Dwarf-Shrubland
16372	Alpine Floodplain
16430	Alpine Dwarf Shrubland
16440	Sitka Spruce
16450	Alpine Mesic Herbaceous Meadow
16460	Western Hemlock
16481	Mountain Hemlock
16500	Periglacial Woodland-Shrubland
16520	Mesic Scrub Birch/Willow
16550	Montane Floodplain - Subboreal
16590	Mountain Hemlock Peatland
16620	Shrub and Herbaceous Floodplain
16790	White Spruce-Hardwood - SubBoreal
16800	Avalanche Slope Shrubland

²Structure classifications- see attached landscape classification matrices for specific structure classes and moose habitat ratings for each category.

The following describe some of the assumptions and initial values to be used in the model.

Winter

BPS and structure categories were used to rate forage quality and thermal cover quality of each pixel or stand. Ratings for each BPS/structural class category were assigned as 0.1, 0.25, 0.5, 0.75, or 1, with poor quality habitat ranked as 0.1. Similarly, thermal cover values were assigned, but a minimum size of 2 acres was set for a stand to qualify as thermal/escape cover.

Spring

BPS and structure categories were used to rate each category for spring foods and escape cover. Forage values were rated as 0.1, 0.25, 0.5, 0.75, or 1, with poor quality habitat ranked as 0.1. As with winter thermal cover, escape cover needed to be at least 2 acres in size to be rated as escape cover.

Summer

BPS and structural categories were rated for summer habitat in terms of providing desired foraging or thermal/escape cover. As with winter and spring, each pixel or stand was assigned a rating based on its BPS and structure category as 0.1, 0.25, 0.5, 0.75, or 1, with poor quality habitat ranked as 0.1. Similar to winter and spring, thermal cover needed to be at least 2 acres in size to be considered functional.

Interspersion Evaluation

Foraging is influenced by proximity to thermal or escape cover. This may vary by season. For example, summer thermal cover may only need to be within 500m, while winter thermal escape cover is preferred within 100m and use may be minimal beyond 300m away. The following interspersion values were used as modifiers such that they reduced the rated value of foraging areas by moose. Winter foraging: Forage quality based on the forage value assigned to the BPS/structure class discussed above was further modified based on its interspersion with thermal cover. Forage within 300m of thermal cover that was rated at least 0.5 in value remained at its assigned forage quality rating. Beyond 300m, the forage quality rating was multiplied by 0.25 to reduce its value. Alternatively, forage qualities could be further refined based on distance from thermal cover: within 100m of thermal/escape cover rated as the full forage value, 100-200m rated as 0.75 times the forage value, 200-300m away rated as 0.5 times the forage value, and >300m rated as 0.25 times the forage value. These banded modifiers have not been incorporated into the current model.

Spring forage quality values were reduced in the same manner as winter forage values based on distance from escape cover. Forage quality ratings within 300m of escape cover that was rated as 0.5 or greater in value received its full foraging value, while areas >300m from escape cover were reduced in value as 0.25 times the forage value.

Summer foraging: Summer forage quality values that were located within 500m of thermal cover at least 0.5 in value received full value. Areas >500m away from thermal cover were reduced in value by a multiplier of 0.25 times the forage value.

Summer thermal cover may also be provided by wetlands/ponds. These have not currently been mapped or included in the model.

Additional Considerations

Habitat quality is clearly a driver of moose population status and dynamics, as high quality habitat is necessity for populations to have high levels of recruitment and survival. However, given the relatively low productivity of moose populations even in high quality habitat, mortality factors can play a significant role as well. Severe winters can impact populations (Ballard et al. 1991). Wolves have been identified as a major predator of moose (Ballard et al. 1991). Bears, both brown and black have also been found to be significant predators on calves (Ballard et al. 1991). Balancing habitat quality, predator populations, human impacts on habitat, and human harvest of moose is a challenging management issue. Boertje et al. (2010) reviewed information on moose population dynamics in relation to predator control programs. They found that predator control can be an effective tool to increase available human harvest of moose, but stressed that nutrient-based management and consideration of other factors is essential to make both politically and biologically correct decisions. Similarly, Crete and Courtois (1997) noted that limiting factors to moose populations need to be assessed prior to making management decisions relative to mortality factors.

Model Results

Figures 2-7 display the outputs of the moose habitat model for the Ahtna Traditional Use Territory.

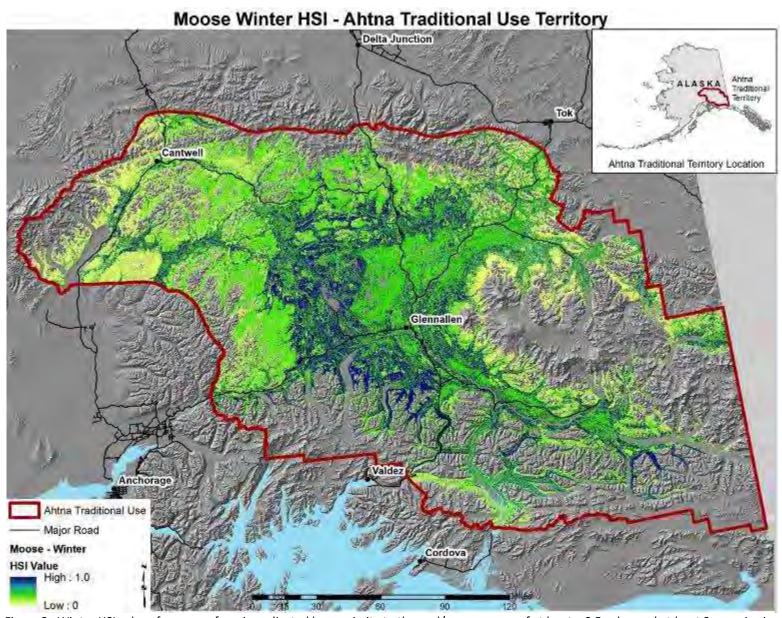


Figure 2. Winter HSI values for moose foraging adjusted by proximity to thermal/escape cover of at least a 0.5 value and at least 2 acres in size.

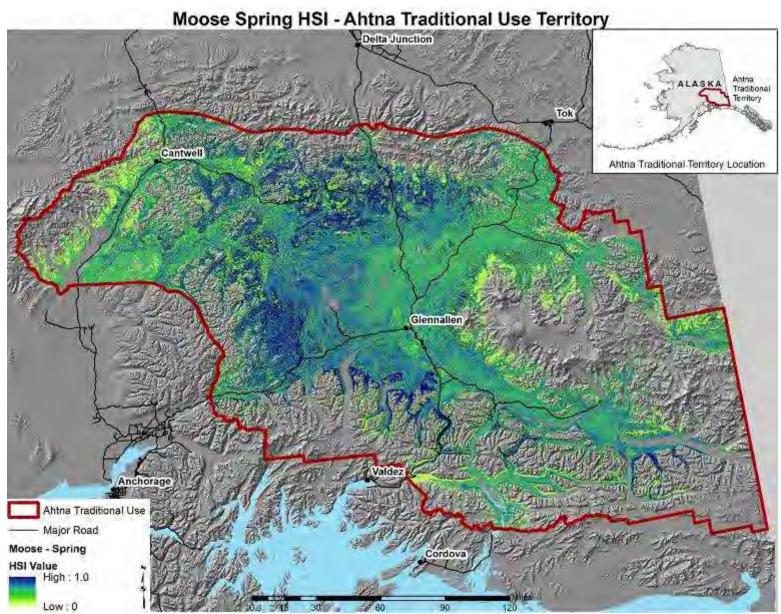


Figure 3. Spring HSI values for moose foraging adjusted by proximity to escape cover of at least a 0.5 value and at least 2 acres in size.

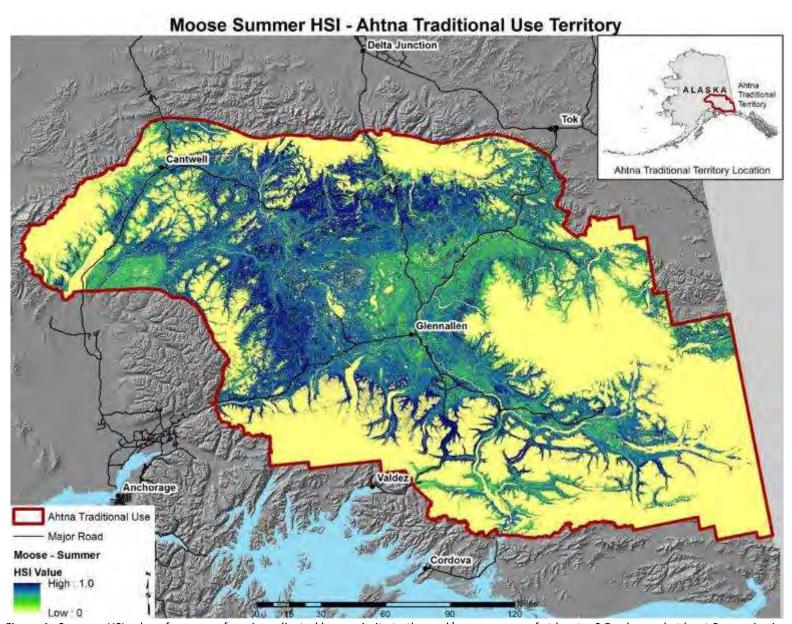


Figure 4. Summer HSI values for moose foraging adjusted by proximity to thermal/escape cover of at least a 0.5 value and at least 2 acres in size.

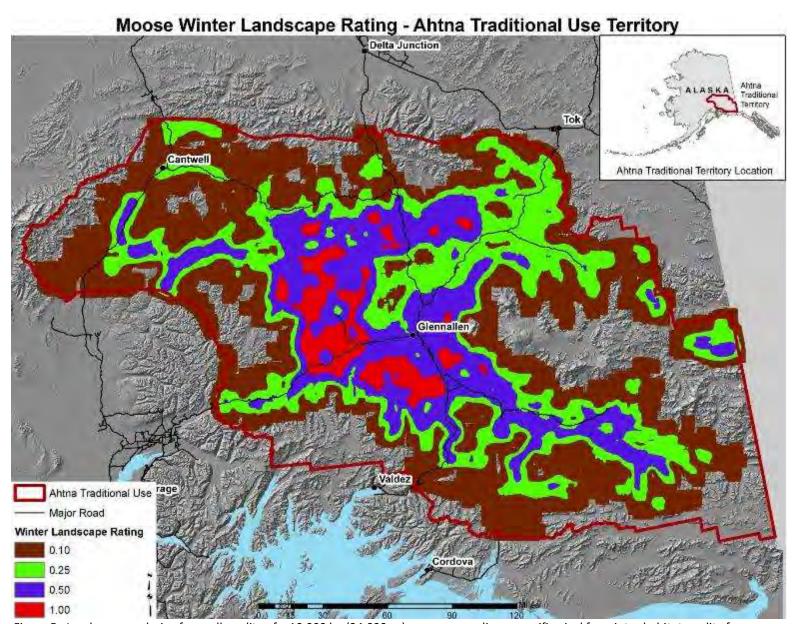


Figure 5. Landscape analysis of overall quality of a 10,000 ha (24,000 ac) area surrounding a specific pixel for winter habitat quality for moose.

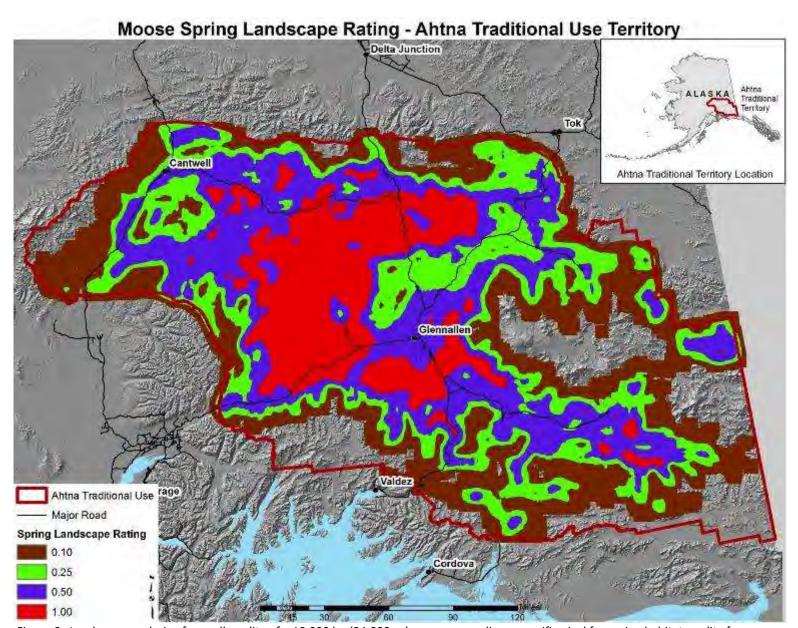


Figure 6. Landscape analysis of overall quality of a 10,000 ha (24,000 ac) area surrounding a specific pixel for spring habitat quality for moose.

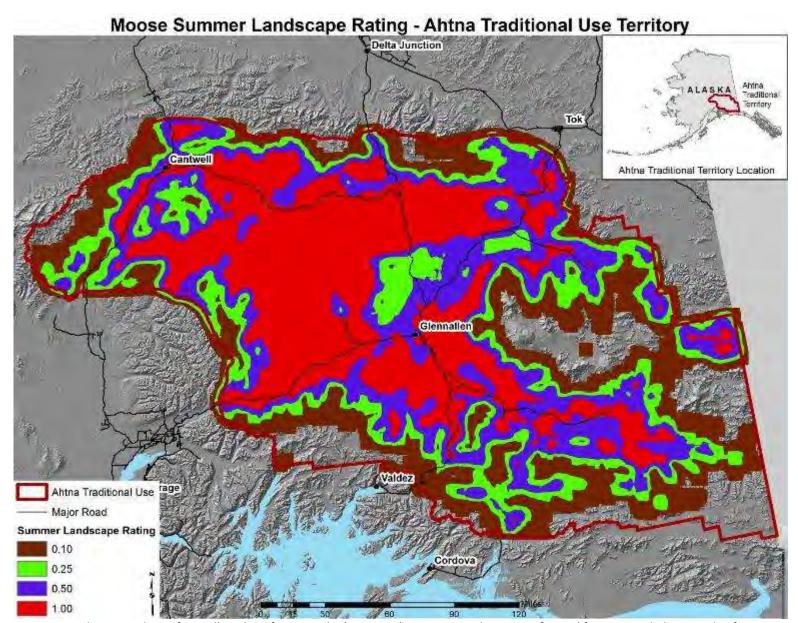


Figure 7. Landscape analysis of overall quality of a 10,000 ha (24,000 ac) area surrounding a specific pixel for summer habitat quality for moose.

Literature Cited

- Allen, A. W., P. A. Jordan, and J. W. Terrell. 1987. Habitat suitability index models: Moose. 82(10.155). Lake Superior Region, U.S. Fish and Wildlife Service Biological Report.
- Ballard, W. B., J. S. Whitman, and D. J. Reed. 1991. Population dynamics of moose in south-central Alaska. Wildlife Monographs 114:1-49.
- Beyer, H. L., R. Ung, D. L. Murray, and M. J. Fortin. 2013. Functional responses, seasonal variation and thresholds in behavioural responses of moose to road density. Journal of Applied Ecology 50:286-294.
- Bjorneraas, K., I. Herfindal, E. J. Solberg, B. E. Sther, B. van Moorter, and C. M. Rolandsen. 2012. Habitat quality influences population distribution, individual space use and functional responses in habitat selection by a large herbivore. Oecologia 168:231-243.
- Boertje, R. D., M. A. Keech, and T. F. Paragi. 2010. Science and values influencing predator control for Alaska moose management. Journal of Wildlife Management 74:917-928.
- Bowyer, R. T., V. Van Ballenberghe, J. G. Kie, and J. A. K. Maier. 1999. Birth-site selection by Alaskan moose: Maternal strategies for coping with a risky environment. Journal of Mammalogy 80:1070-1083.
- Chekchak, T., R. Courtois, J. P. Ouellet, L. Breton, and S. St-Onge. 1998. Characteristics of moose (Alces alces) calving sites. Canadian Journal of Zoology-Revue Canadienne De Zoologie 76:1663-1670.
- Collins, W. B., and D. J. Helm. 1997. Moose, Alces alces, habitat relative to riparian succession in the boreal forest, Susitna River, Alaska. Canadian Field-Naturalist 111:567-574.
- Crete, M., and R. Courtois. 1997. Limiting factors might obscure population regulation of moose (Cervidae: Alces alces) in unproductive boreal forests. Journal of Zoology 242:765-781.
- Demarchi, M. W., and F. L. Bunnell. 1995. Forest cover selection and activity of cow moose in summer. Acta Theriologica 40:23-36.
- Dettki, H., R. Lofstrand, and L. Edenius. 2003. Modeling habitat suitability for moose in coastal northern Sweden: Empirical vs process-oriented approaches. Ambio 32:549-556.
- Doerr, J. G. 1983. Home range size, movements and habitat use in 2 moose, Alces alces, populations in southeastern Alaska. Canadian Field-Naturalist 97:79-88.
- Dussault, C., R. Courtois, and J. P. Ouellet. 2006. A habitat suitability index model to assess moose habitat selection at multiple spatial scales. Canadian Journal of Forest Research-Revue Canadianne De Recherche Forestiere 36:1097-1107.
- Dussault, C., R. Courtois, J. P. Ouellet, and I. Girard. 2005a. Space use of moose in relation to food availability. Canadian Journal of Zoology-Revue Canadienne De Zoologie 83:1431-1437.
- Dussault, C., J. P. Ouellet, R. Courtois, J. Huot, L. Breton, and H. Jolicoeur. 2005b. Linking moose habitat selection to limiting factors. Ecography 28:619-628.
- Dussault, C., J. P. Ouellet, R. Courtois, J. Huot, L. Breton, and J. Larochelle. 2004. Behavioural responses of moose to thermal conditions in the boreal forest. Ecoscience 11:321-328.
- Dussault, C., J. P. Ouellet, C. Laurian, R. Courtois, M. Poulin, and L. Breton. 2007. Moose movement rates along highways and crossing probability models. Journal of Wildlife Management 71:2338-2345.
- Eastman, D. S. 1977. Habitat selection and use in winter by moose in sub-boreal forests of north-central British Columbia, and relationships to forestry. Dissertation, University of British Columbia, Vancouver, Canada.
- Eldegard, K., J. T. Lyngved, and O. Hjeljord. 2012. Coping in a human-dominated landscape: trade-off between foraging and keeping away from roads by moose (Alces alces). European Journal of Wildlife Research 58:969-979.
- Harris, G., R. M. Nielson, T. Rinaldi, and T. Lohuis. 2014. Effects of winter recreation on northern ungulates with focus on moose (Alces alces) and snowmobiles. European Journal of Wildlife Research 60:45-58.
- Hepinstall, J. A., L. P. Queen, and P. A. Jordan. 1996. Application of a modified habitat suitability index model for moose. Photogrammetric Engineering and Remote Sensing 62:1281-1286.
- Herfindal, I., J. P. Tremblay, B. B. Hansen, E. J. Solberg, M. Heim, and B. E. Saether. 2009. Scale dependency and functional response in moose habitat selection. Ecography 32:849-859.
- Laurian, C., C. Dussault, J. P. Ouellet, R. Courtois, and M. Poulin. 2012. Interactions between a large herbivore and a road network. Ecoscience 19:69-79.

- Laurian, C., C. Dussault, J. P. Ouellet, R. Courtois, M. Poulin, and L. Breton. 2008. Behavior of moose relative to a road network. Journal of Wildlife Management 72:1550-1557.
- Leclerc, M., C. Dussault, and M. H. St-Laurent. 2012a. Multiscale assessment of the impacts of roads and cutovers on calving site selection in woodland caribou. Forest Ecology and Management 286:59-65.
- Leclerc, M., J. Lamoureux, and M. H. St-Laurent. 2012b. Influence of young black spruce plantations on moose winter distribution. Journal of Wildlife Management 76:1686-1693.
- LeResche, R. E., and J. L. Davis. 1973. Importance of nonbrowse foods to moose on the Kenai Peninsula, Alaska. Journal of Wildlife Management 37:279-287.
- Lowe, S. J., B. R. Patterson, and J. A. Schaefer. 2010. Lack of behavioral responses of moose (Alces alces) to high ambient temperatures near the southern periphery of their range. Canadian Journal of Zoology-Revue Canadienne De Zoologie 88:1032-1041.
- Mabille, G., C. Dussault, J. P. Ouellet, and C. Laurian. 2012. Linking trade-offs in habitat selection with the occurrence of functional responses for moose living in two nearby study areas. Oecologia 170:965-977.
- MacCracken, J. G., V. V. Ballenberghe, and J. M. Peek. 1997. Habitat relationships of moose on the Copper River Delta in coastal south-central Alaska. Wildlife Monographs 136:3-52.
- MacCracken, J. G., V. Vanballenberghe, and J. M. Peek. 1993. Use of aquatic plants by moose sodium hunger or foraging efficiency. Canadian Journal of Zoology-Revue Canadienne De Zoologie 71:2345-2351.
- Maier, J. A. K., J. M. V. Hoef, A. D. McGuire, R. T. Bowyer, L. Saperstein, and H. A. Maier. 2005. Distribution and density of moose in relation to landscape characteristics: effects of scale. Canadian Journal of Forest Research-Revue Canadienne De Recherche Forestiere 35:2233-2243.
- McCann, N. P., R. A. Moen, and T. R. Harris. 2013. Warm-season heat stress in moose (Alces alces). Canadian Journal of Zoology-Revue Canadienne De Zoologie 91:893-898.
- McLoughlin, P. D., E. Vander Wal, S. J. Lowe, B. R. Patterson, and D. L. Murray. 2011. Seasonal shifts in habitat selection of a large herbivore and the influence of human activity. Basic and Applied Ecology 12:654-663.
- Melin, M., J. Matala, L. Mehtatalo, R. Tiilikainen, O. P. Tikkanen, M. Maltamo, J. Pusenius, and P. Packalen. 2014. Moose (Alces alces) reacts to high summer temperatures by utilizing thermal shelters in boreal forests an analysis based on airborne laser scanning of the canopy structure at moose locations. Global Change Biology 20:1115-1125.
- Miquelle, D. G., J. M. Peek, and V. Vanballenberghe. 1992. Sexual segregation in Alaskan moose. Wildlife Monographs 122:1-57.
- Neitfeld, M., J. Wilk, K. Woolnough and B. Hoskin. 1985. Wildlife habitat requirement summaries for selected wildlife species in Alberta. Wildlife Resource Inventory Unit, Alberta Energy and Natural Resources Fish and Wildlife Division.
- Oehlers, S. A., R. T. Bowyer, F. Huettmann, D. K. Person, and W. B. Kessler. 2011. Sex and scale: implications for habitat selection by Alaskan moose Alces alces gigas. Wildlife Biology 17:67-84.
- Osko, T. J., M. N. Hiltz, R. J. Hudson, and S. M. Wasel. 2004. Moose habitat preferences in response to changing availability. Journal of Wildlife Management 68:576-584.
- Poole, K. G., and K. Stuart-Smith. 2006. Winter habitat selection by female moose in western interior montane forests. Canadian Journal of Zoology-Revue Canadienne De Zoologie 84:1823-1832.
- Puttock, G. D., P. Shakotko, and J. G. Rasaputra. 1996. An empirical habitat model for moose, Alces alces, in Algonquin Park, Ontario. Forest Ecology and Management 81:169-178.
- Schwab, F. E., and M. D. Pitt. 1991. Moose selection of canopy cover types related to operative temperature, forage, and snow depth. Canadian Journal of Zoology-Revue Canadienne De Zoologie 69:3071-3077.
- Shanley, C. S., and S. Pyare. 2011. Evaluating the road-effect zone on wildlife distribution in a rural landscape. Ecosphere 2:1-16.
- van Beest, F. M., B. Van Moorter, and J. M. Milner. 2012. Temperature-mediated habitat use and selection by a heat-sensitive northern ungulate. Animal Behaviour 84:723-735.
- White, K. S., and J. Berger. 2001. Antipredator strategies of Alaskan moose: are maternal trade-offs influenced by offspring activity? Canadian Journal of Zoology-Revue Canadienne De Zoologie 79:2055-2062.

Appendix B. Caribou Habitat Model for South Central Alaska Introduction

The Copper River-Ahtna Intertribal Conservation District (CRITR) has initiated a landscape scale project funded through the Conservation Innovation Grant program of the Natural Resources Conservation Service. The purpose of the project is to develop innovative tools and to increase the technical capacity for planning by CRITR. The specific objectives include:

- Conduct an ecologically based resource assessment of Ahtna lands; develop ecological site classification, and develop site-specific vegetation treatments;
- > Develop moose and caribou habitat models to support habitat management of these species;
- Develop an innovative 10-year management plan for Ahtna's 1.7 million acres to increase moose for food and biomass for energy while maintaining or improving caribou habitat; and
- Train local technicians to conduct habitat treatments and monitor results.

A caribou habitat model is an important tool to help identify sites with the best potential for maintaining caribou habitat and to incorporate management of these sites into an overall plan that considers a landscape context.

Caribou in Alaska

Caribou (Rangifer tarandus) in Alaska are all considered to be of the Grants subspecies (R.t. granti). Weckworth et al. (2012) examined mitochondrial DNA and found that this subspecies was not different from the barren-ground caribou (R.t. groenlandicus) in the Yukon and Northwest Territories, although these subspecies are still generally considered separate. Weckworth et al. (2012) also reported greater similarities in the northern woodland ecotype of the woodland caribou (R.t. caribou) along with Grants and barren-ground caribou when compared to other subspecies of caribou which they attributed to the location of refugia for these subspecies during the ice ages, with the Grants, barren-ground, and northern woodland caribou being of Beringian origin while the other subspecies and ecotypes relied on refugia south of the glacial ice. These differences could result in habitat and behavioral differences between these groupings, although all caribou and reindeer are considered the same species. Genetic analysis of Grants caribou in Alaska (Mager et al. 2014) indicated that while individual herds on the Alaska Peninsula showed considerable genetic variation among discrete herds, caribou on the Mainland did not show the same level of herd distinction. Caribou subspecies and ecotypes in Canada are largely distinguished by differences in their environments and in their habitat use. In Alaska, genetic analyses (Mager et al. 2014) do not support differentiation among the different mainland herds even though considerable differences exist in the types of habitat conditions used across the range of the species. Caribou habitat requirements vary throughout the year with greatest consideration given to winter habitat, calving habitat, and summer habitat. Winter and summer habitats are influenced by the availability of preferred foods as well as avoidance of predation risk and disturbance from human activities. Calving habitat, as discussed below, appears to be most sensitive to predator avoidance, but also requires food availability.

Caribou habitat selection has been identified to have hierarchical considerations meaning that landscape characteristics can determine if caribou will use a particular zone or region, while daily use of sites within the zone will be based on specific stand characteristics (Bradshaw et al. 1995, Stuart-Smith et al. 1997, Anderson 1999, Boan et al. 2014). Both landscape and daily use scales need to be considered in assessing caribou habitat requirements.

Caribou Herds in South Central Alaska

Caribou herds occurring in South Central Alaska discussed by Hemming (1971), Alaska Department of Fish and Game (ADFG) (2011) and Collins et al. (2011) included the Chisana, Mentasta, and Nelchina herds. Population estimates by ADFG (2011) for expected herd sizes in 2011 were 700 for the Chisana herd, 350 for the Mentasta herd, and 46,500 for the Nelchina herd. The U.S. National Park Service (Putera 2015) estimated the Chisana herd size to be 701 in 2013 with a 90% confidence interval of 639-763. They reported 16 calves/100 cows and 49 bulls for 100 cows from their 2013 survey.

Chisana Herd

The Chisana herd was described by Hemming (1971) as originating from the Fortymile Herd and were reported to use the Nutzotin Mountains along tributaries of the Chisana and White rivers. He also reported that no migratory movements occurred, and that calving was reported to occur from the benchlands along Sheep Creek on Mt. Sulzer to the rolling hills north of Ptarmigan Lake. This herd currently occupies the upper Chisana and White River drainages in Wrangell St. Elias National Park and areas of neighboring Yukon, Canada (Bentzen 2011). Alaska considers this part of the Grants subspecies of caribou while Canada considers it part of the woodland subspecies of caribou which has been supported by genetic analysis (Zittlau et al. 2000). This herd has been noted to have habitat selection for calving similar to mountain caribou in that individual cows select higher elevations to disperse their densities (Bentzen 2011). ADFG (2011) identified low calf recruitment as a management concern for the Chisana Herd. They reported that winter range condition appeared to be poor based on a low percentage of lichens and higher amounts of moss in the winter diet in this area, however no recommendations on how to improve winter habitat quality were included in their report.

Mentasta Herd

The Mentasta herd was reported by Hemming (1971) as having originated from the Fortymile Herd and occurring from the Mentata Mountains south to the western slopes of the Wrangell Mountains. He also reported that winter habitat was alpine areas and sparsely covered spruce flats from the Wrangell Mountains north to the Gerstle River, and that calving occurred on the slopes of Mount Sanford and on the Macomb Plateau east of the Johnson River. This herd is included on the map of caribou herds in Alaska (ADFG 2011), but overlaps with the Nelchina herd. It is not individually described in the ADFG caribou management report.

Nelchina Herd

The range of the Nelchina herd, occurring in the Nelchina Basin, was extensively described by Hemming (1971). However, Collins et al. (2011) reported changes to this range starting in the 1990's as a result of heavy foraging on lichens in the original range (Figure 1). Collins et al. (2011:369) provided an excellent description of these dynamics:

"The Nelchina Caribou (Rangifer tarandus) Herd (NCH) declined from a peak population of 70,000 in the mid-1960s to approximately 10,000 in the early 1970s (Siniff and Skoog 1964, Bos 1975, Lieb et al. 1988). From 1977 to 1995, the herd rebounded, reaching about 45,000 caribou and surpassing the Alaska Department of Fish and Game management objective of 30,000.

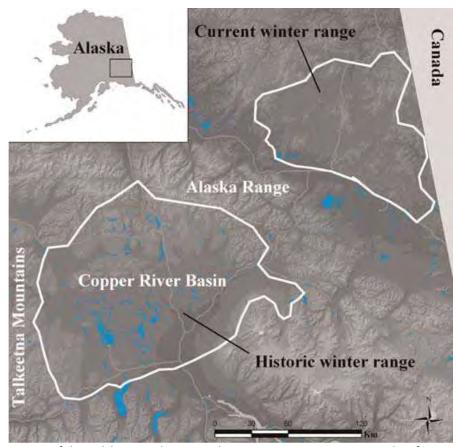


Figure 1. Current range of the Nelchina Herd compared to its previous range. Figure taken from Collins et al. (2011).

During this time, forage lichens only partially recovered from previous overgrazing, raising concerns about negative impacts by caribou on lichen standing crops within the herd's winter range (Lieb 1994). Since the mid-1990s, increased harvests have been used to reduce herd size to the management objective, and numbers have fluctuated between about 30,000 and 35,000 caribou. In the early 1990s as the NCH reached high numbers and lichens were overgrazed, the herd began a dramatic shift from its historic wintering range in the Copper River Basin and southern Alaska Range to its current winter range in the Tanana Hills and western Yukon Territory (Valkenburg et al. 2002). This was the first shift in approximately 100 yr for which records are available, and it required additional migration of 100 km. Prior to arrival of the NCH, the current winter range had not been grazed since collapse of the Fortymile Caribou Herd approximately 40 yr earlier."

This herd has varied considerably in population size from 7,000-70,000 animals over the past 70 years but is now managed with a population goal of between 35-40,000 and an annual harvest of 3-6,000 animals (Schwanke 2011). Calving habitat occurs in the eastern Talkeetna Mountains from the Little Nelchina River to the Fog Lakes (Schwanke 2011) with habitat use of this area continuing into the summer. Winter habitat use occurs from Cantwell east through game management units 13A, 13B, 11, 12 and 20E with apparent use of lichens associated with older burns along the Taylor Highway (Schwanke 2011). Despite a large burn in 2004 in subunit 20E which is avoided, 60-95% of the herd still winters in this subunit. Overlap with wintering caribou from the Fortymile herd in this area could lead to over-utilization of lichen and range deterioration.

Habitat monitoring of areas used by the Nelchina herd (Lieb 1994 as reported in Schwanke 2011) found that at times when herd sizes were larger, amounts of lichens in key winter and calving areas declined. It appeared that heavy use of the East Talkeetna Mountain area in the 1970's heavily impacted lichen amounts so that the area has not been used as a calving or wintering area ever since. Poor range conditions were noted as a contributing factor to the low weight gains of calves in this herd. Following herd reductions through hunting programs initiated in the 1990's, calf weight gains, though still quite variable, have been increasing. However, because of the high elevation of the range of this herd, cool, late springs shorten the summer growing season, reducing food availability in some years, resulting in considerable variation in calf weight gains.

Caribou Habitat Requirements

General and Year-round Caribou Habitat Requirements

Critical habitat needs of woodland caribou were discussed by Racey and Arsenault (2007). They identified critical habitat as high quality winter and summer range, calving areas, other known high use areas, and the connectivity among these areas. They recommended strategies to maintain these high quality areas with minimal disturbance including fire and anthropogenic activities.

Rettie and Messier (2000) reported that caribou in Saskatchewan used seasonal ranges that included peatlands and black spruce dominated stands relative to recently disturbed stands and early seral stage forests. In all populations, caribou preferred peatlands and black spruce forests to all other habitat types at the daily area scale.

Jones et al. (2007) identified the variability in habitat requirements of "ecotypes" of woodland caribou in British Columbia with different herds selecting for different vegetation types to meet their habitat requirements. This study emphasizes the importance of considering the local habitat selection of caribou in specific management areas, as the species differs considerably across its range in what it selects as habitat to meet its various food, reproduction, cover, and other needs. For example, Apps and McLellan (2006) analyzed distributions of subpopulations of mountain caribou in southeastern British Columbia and found that this subspecies required blocks of old forests of cedar and hemlock occurring on wet sites that were away from roads and other human developments, revealing a different habitat selection than other subspecies of caribou even within British Columbia.

Wittmer et al. (2005) reported that various caribou subpopulations are declining as a consequence of increased predation. Recovery of these subpopulations will require a multi-species perspective to address the issues of predator densities in relation to other prey species. Ferguson et al. (1988) studied a herd of woodland caribou that persisted on an island in Ontario and concluded that this small population were not extirpated as were those on the nearby mainland because of the lack of predation by wolves, as food resources were greater on the mainland than on the island, but wolves were not regularly present on the island.

O'Brien et al. (2006) stated "Although predation on calves and adults is the proximate limiting population factor, it is the reduction in the availability of lichen-rich mature conifer stands and increased access by predators and hunters that have led to a decline in woodland caribou populations across Canada (e.g., <u>James and Stuart-Smith, 2000</u>, <u>Smith et al., 2000</u> and <u>Dyer et al., 2001</u>)." They recommended strategies that maintain older conifer forests providing high-quality patches that are connectivity within a matrix that can facilitate movement and foraging away from predators and human

activity, citing supporting evidence from <u>Rettie and Messier (2000)</u>, <u>Smith et al. (2000)</u>, and <u>James et al. (2004)</u>. O'Brien et al. (2006) examined locations and movements of caribou in Manitoba and reported that of 721 late winter location points, "42.4% were located within high-quality patches, 90.8% were located within 500 m and 99.3% were located within 1000 m from high-quality patches." They compared this to random points with had 21.6% located within high-quality patches, 60.9% within 500 m and 81.8% within 1000 m of high-quality patches. They noted that the caribou herds they studied concentrated in areas which they defined as being >5-10,000 ha in size which contained large clusters of high quality habitat.

Wittmer et al. (2007) found that timber harvesting that increased the occurrence of early seral forests in landscapes occupied by woodland caribou in British Columbia altered the predator—prey system by maintaining alternative prey that increased numbers of predators. Survival probabilities for adult females were best explained by the amount of early seral stands within an overall range of a subpopulation. Female survival was higher at the home range scale when higher proportions of old forest and lower amounts of mid-aged forest were present.

Joly et al. (2003) reported that caribou in the Nelchina herd in southcentral Alaska selected areas that had not burned in the past 50 years, but Schwanke (2011) noted that burned areas did provide grasses and forbs that are preferred summer food, so a mosaic of burned and unburned areas may be desirable. Robinson et al. (2012) reported that burned areas decreased quality of caribou habitat while also increasing densities of wolves that then reduced the amount of "safe zones" for caribou. They suggested that management should strive to maintain these safe zones for caribou. Briand et al. (2009) studied woodland caribou in eastern Canada and reported that they avoided areas in summer that had a well-developed shrub layer that would be higher quality habitat for moose.

Summer foods of caribou were reported to not be well documented, but were thought to include sedges, cotton-grass, fungi, grasses, ericaceous shrubs (e.g., Labrador tea, blueberry, bearberry), twinflower, mosses and woody browse such as willows, birch and aspen (Cenovus FCCL Ltd. Narrows Lake Project 2010). Thompson et al. (2015) determined diets of caribou in Ontario and reported that caribou heavily used lichens in the winter, and continued to make use of these foods in the summer as well. The breadth of diet doubled in summer, but lichens still contributed over 60% of food eaten. Three genera of green plants were the predominant additional foods eaten in summer but were not species associated with early successional areas.

Calving Habitat

Calf predation during the neonatal period was reported to occur from wolves (Gasaway et al. 1983, Bergerud and Elliot 1986, Bergerud and Page 1987, Seip 1992), bears (Ballard 1994, Adams et al. 1995, Young and McCabe 1997, Mahoney and Virgl 2003, Pinard et al. 2012, Dussault et al. 2012), golden eagles (Dale et al. 1994, Adams et al. 1995, Schwanke 2011), bald eagles (Schwanke 2011), and wolverine (Gustine et al. 2006). Risk of predation appears to strongly influence selection of calving areas (Bergerud et al. 1984, Bergerud 1996, Barten et al. 2001) as well as the general presence of caribou (Boan et al. 2014). Pinard et al. (2012) found that caribou in their study area in Quebec appeared to select calving areas that minimized densities of wolves, their primary predator. However, in this study area bear populations have been increasing and caribou did not appear to select calving areas that minimized bear densities resulted in nearly 60% calf predation by bears.

Bergerud (2007) reported that the persistence of mountain and boreal woodland caribou depended upon low predation risk in calving areas and suggested that for these subspecies, directly reducing

predation by wolves and coyotes may be needed given the expansion of these predators and their primary prey (moose and deer) due to human activities including climate change. Concern over increasing numbers of predators was also noted by Latham et al. (2013) who identified expanding deer and coyote populations as a concern for woodland caribou in Alberta where increases in deer could support increased numbers of predators including coyotes.

Latham et al. (2011a) agreed with concerns over wolf predation on caribou but also noted that bears are an additional predator that could influence calf survival. While they found that bears generally avoided the bogs and fens selected by woodland caribou in Alberta where they conducted their study, they noted that some bears did select these areas and could be effective predators on calves. Latham et al. (2011b) examined relationships of wolves, moose, deer and woodland caribou in southern Alberta. They found that deer populations have increased dramatically and appear to be supporting an increase in wolf populations. They reported that wolf predation on caribou had also increased and contributed to caribou populations going from being stable to being in decline.

Gustine et al. (2006) reported that woodland caribou selected calving areas that had lower amounts of herbaceous vegetation and more shrub cover than random locations in British Columbia. However, Barten et al. (2001) reported that female caribou with calves avoided shrub areas and selected lichen tundra areas more than female caribou without calves in Wrangell St. Elias National Park in Alaska.

Leclerc et al. (2012) studied woodland caribou calving areas in Quebec. They found that calving areas were located away from roads and cutover areas at three different scales, the annual home range, calving home range, and forest stand. They noted that at the forest stand scale calving areas were located away from cutovers and roads and in areas with a lower basal area of black spruce or balsam fir.

Winter Habitat

Barrier and Johnson (2012) investigated winter foraging sites of barren-ground caribou in the Northwest Territories. They found that caribou selected sites with higher amounts of lichen present. Higher amounts of rock or higher basal area of conifer trees reduced selection of sites for foraging. They suggested that future increases in incidences and severity of fires could reduce available habitat for caribou. Briand et al. (2009) found that woodland caribou in eastern Canada selected wintering areas that had higher amounts of terrestrial lichens or ericaceous shrub cover in older stands, and avoided areas that had a well-developed shrub layer that would be higher quality habitat for moose. Joly et al. (2010) found that caribou in Northwest Alaska selected areas with higher amounts of lichen in winter, and avoided burned areas that were less than 58 years old.

The Cenovus FCCL Ltd. Narrows Lake Project (2010) developed a caribou habitat model for Alberta and provided a review of caribou habitat requirements. They reported that winter habitat selection by woodland caribou is strongly associated with peatland habitats citing studies conducted by Anderson (1999), Bradshaw et al. (1995), Edmonds and Bloomfield (1984), Stuart-Smith et al. (1997), and Schneider et al. (2000). Upland-dominated landscapes were reported to be generally used less by woodland caribou (Bradshaw et al. 1995, Schneider et al. 2000). Schneider et al. (2000) reported that the majority of upland habitat use by caribou was in patches occurring within large peatland complexes. Schneider et al. (2000) also reported that caribou use of pure upland habitat decreased exponentially with distance from peatlands.

The most important winter foods of caribou in boreal are terrestrial lichens (Edmonds and Bloomfield 1984, Manitoba Model Forest 1995) including *Cladina* species, such as *C. mitis*, *C. uncialus and C.*

rangiferina; Centraria islandica and Stereocaulon spp. (Manitoba Model Forest 1995). Bradshaw et al. (1995) reported that Cladina were the most common food species found in snow craters dug by woodland caribou in northeastern Alberta.

Collins et al. (2011) studied winter habitat selection of the Nelchina Herd. They found that this herd seldom occurred above 1500m in elevation in winter. As discussed previously, they found that this herd had shifted the location of its winter range, and attributed this to over-utilization of lichen in the original range and the improved status of lichen in the new winter range. They reported the lichens present in the new range to be Cladonia amaurocraea, Cladina rangiferina, Flavocetraria cucullata, that were used by caribou along with lowbush cranberry (Vaccinium vitis-idaea). Collins et al. (2011:370) described this wintering area as: "Black spruce (Picea mariana) forest was the dominant cover type. Aspen (Populus tremuloides), birch (Betula papyrifera), and white spruce (Picea glauca) were also present as small stands interspersed within the black spruce forest. At elevations >1,100 m, forest gave way to shrubs (Alder [Alnus spp.], birch [Betula spp.]) and alpine communities. Muskegs and tussock tundra were common in poorly drained, low-lying areas." They also reported that preferred species of lichen were Cladina arbuscula-mitis, C. rangiferina, and C. stellaris, with Flavocetraria cucullata, Cetraria islandica, Cladonia uncialis, and Stereocaulon spp. being secondary species. Peltigera apthosa and Peltigera spp. were lichens reported to not be used as forage. Collins et al. (2011:375) found that: "After fires, forage lichens seldom recovered sufficiently to attract grazing until after 60 yr, and as a group, primary forage lichen species did not reach maximum productivity until after 180 yr." Overgrazing of lichens could significantly reduce lichen abundance, but recovery could occur as quickly as 20 years if the grazing was removed.

Boan et al. (2014) reported that presence of wintering woodland caribou in Ontario was negatively influenced by higher probability of wolves, which in turn were influenced by the presence of moose and logging roads. They also found a negative relationship between occurrence of moose habitat in close proximity to caribou and an indirect negative influence of the quality of moose forage habitat and caribou habitat.

Anthropogenic Disturbance and Habitat Changes

Vors et al. (2007) estimated effects of anthropogenic disturbances on caribou extirpation in areas of Ontario and reported that "forest cutovers were the best predictor of caribou occupancy, with a tolerance threshold of 13 km to nearest cutover and a time lag of 2 decades between disturbance by cutting and caribou extirpation."

James and Stuart-Smith (2000) examined woodland caribou and wolf locations and predation sites in relation to linear corridors (roads, trails, seismic lines, and pipelines). They reported that caribou locations were farther from linear corridors than random locations, while wolf locations were closer. They also reported that wolf predation sites on caribou were closer to linear corridors than live locations of caribou, and caribou killed by wolves were killed closer to linear corridors than their live locations prior to being predated. They concluded that adding linear corridors in caribou range will increase caribou risks to predation.

James et al. (2004) studied woodland caribou in Alberta. They reported "selection of fen/bog complexes by caribou and selection of well-drained habitats by moose and wolves resulted in spatial separation. This spatial separation in turn reduced wolf predation pressure on caribou but did not provide a total refuge from wolves. Any management activities that increase the density of moose and wolves or increase access of wolves into fen/bog complexes will likely reduce the refuge effect provided by large

fen/bog complexes." This study supported the contention that increasing moose populations in an area will result in increased predation on caribou, but also noted that any increase in accessibility or numbers of wolves in caribou range can contribute to increased mortality and population risks. Similarly, Adam et al. (2004) found that moose and wolves utilized well-drained areas in Alberta while caribou stayed in bogs/fens where numbers of wolves were lower. They cautioned that any activities that increased moose or wolf access into the wetter areas could increase predation on caribou. Both of these studies supported the contention that increasing moose populations in an area will result in increased predation on caribou, but also noted that any increase in accessibility or numbers of wolves in caribou range can contribute to increased mortality and population risks.

Johnson and Russell (2014) studied the distribution of the Porcupine caribou herd in winter in relation to human disturbances over a 27 year time frame. They reported that caribou avoided human disturbances, particularly settlements followed by main roads. They also noted gradual changes over time in avoidance patterns.

Smith et al. (2000) conducted a long-term study of woodland caribou in Alberta and noted the effects of timber harvesting on this population. They reported "Caribou avoided using recently fragmented areas by an average of 1.2 km. If fragmentation of the winter range continues through timber harvesting and other industrial activities, the 'spacing out' antipredator strategy used by caribou may be compromised. Based on these findings, timber-harvesting strategies are recommended that (i) ensure an adequate area of usable habitat to support the current population, (ii) minimize the amount of fragmented area, and (iii) in the short term avoid presently defined core use areas." Courtois et al. (2007) reported that forest management strategies should be oriented toward the protection of large interconnected blocks of forest to favor caribou spacing away from humans and predators in order to keep direct and indirect sources of caribou mortality at low levels. Leclerc et al. (2012) made recommendations for caribou in Quebec including amalgamating all forestry activities within intensive management zones in order to spatially isolate large patches of suitable calving habitat from anthropogenic disturbances.

Home ranges of caribou vary seasonal and depending upon the migratory status of a herd. Johnson et al. (2003) reported home ranges up to 182 km². O'Brien et al. (2006) in examining caribou habitat use found that caribou responded to overall habitat quality in areas of >5-10,000 ha in size. Courtois et al. (2007) recommended maintaining blocks of caribou habitat of 100-250 km² (10,000-25,000 ha) with minimal amounts of disturbance (burns, logging, or mechanical treatments less than 20-30 years old) for persistence of woodland caribou in Quebec. Lesmerises et al. (2013) also examined effects of patch sizes on use of caribou in managed landscapes in Quebec. They determined that use increased sharply as patch sizes increased up to 100 km² (10,000 ha), with further increases up to 500 km² (50,000 ha) but leveled off after that. These studies indicate that with large enough patch sizes of high quality habitat, caribou can disperse to low densities and through this reduce overall risks of predation (Bergerud and Page 1987) where primary predators (wolves) select other areas where prey densities (moose and deer) may be higher. This strategy breaks down then high quality habitat becomes fragmented due to habitat changes caused by human activities including logging and other mechanical treatments, fires that alter historical patterns of disturbance, and anthropogenic infrastructure that disrupt caribou habitat use or that allow greater access into high quality habitat by predators. In particular, as habitat changes and climate change combine effects in some landscapes, new predators such as coyotes and bears have been reported to increase in densities putting new pressures on caribou herds through increased predation, primarily on calves. Fortin et al. (2013) studied caribou distributions in Quebec in relation to anthropogenic created edges and found that caribou, in avoiding human disturbances, tended to be pushed into concentrations about 4.5 km from edges where they then become more vulnerable to

predation by wolves or other predators. This finding provides further explanation of the increases in predation around anthropogenic edges found by James and Stuart-Smith (2000) discussed previously.

Dyer et al. (2002) examined the effects of roads and seismic lines on movements of woodland caribou in Alberta. They did not find any effects from seismic lines, but reported that roads with moderate traffic were 6 times less likely to be crossed than habitat without a road present. Shindler et al. (2007) examined the effects of a logging road through winter habitat of caribou in Manitoba. They reported that even with the road closed to all but logging traffic, that high quality habitat within 1 km of the road received less use by caribou than in areas farther from the road. Similarly, Leblond et al. (2011) reported that caribou were influenced by the presence of roads within 1.25 km in Quebec.

Woodland caribou have been reported to be sensitive to various human activities including activities that allow human access to wilderness areas, especially on All-Terrain Vehicles (ATVs) and snowmobiles (Cenovus FCCL Ltd. Narrows Lake Project 2010, Manitoba Model Forest 1995). Apps and McLellan (2006) found that persistence of subpopulations of mountain caribou was best explained by the presence of preferred habitat types as well as remoteness from human activities including low road densities and minimal motorized access.

Beauchesne et al. (2013) found that caribou in Quebec expanded their home ranges as the amount of disturbances in their habitat increased, up to a point where further increase in disturbances caused home range contraction. They reported "density of major roads and the proportion of clearcuts had an important impact on space use throughout the whole year, but the impact of roads was particularly important during calving, summer and rut, while the impact of clearcuts prevailed in spring, early and late winter." They also found that a more convoluted shape of cutblocks amplified the effect of clearcuts on caribou space use.

Habitat Model for Caribou in Southcentral Alaska

Landscape Assessment

At a landscape scale, caribou habitat will be considered to be important within mapped caribou range (Figure 2). Within existing caribou range, caribou habitat quality will be rated according to the procedures outlined below. A moving window analysis will be used to evaluate the quality of an area of approximately 10,000 ha (24,000 ac) surrounding each pixel as indicated in Table 1.

Table 1. Landscape scale rating of habitat quality based on aggregate quality of winter, calving, and summer habitat within a 10,000 ha area.

Percentage of 10,000 ha area with HSI > 0.75	Area weighted HSI value For winter, calving habitat and summer habitat
>50%	1.0
25-50%	0.5
10-25%	0.25
<10%	0.1

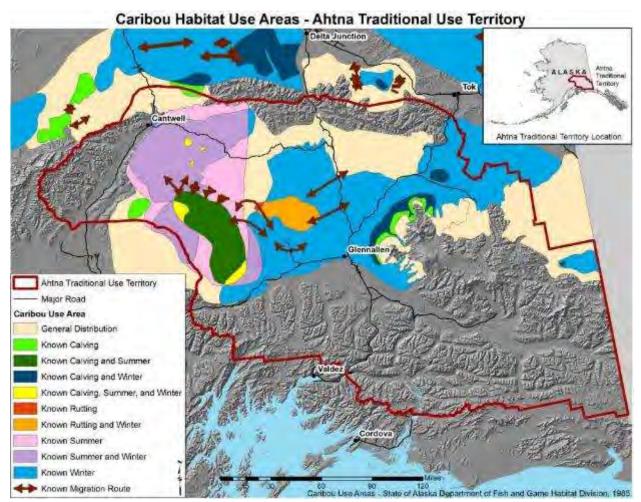


Figure 2. Map of existing caribou distribution in the Ahtna southcentral Alaska project area (from Alaska Department of Fish and Game).

Stand Assessment

The existing caribou range is mapped for each biophysical setting/structural class combination which is then assigned a habitat quality value for both winter and calving/summer habitat (Table 2). This value will then be further modified based on an overlay of anthropogenic disturbances. Table 2 lists the value of each biophysical setting/structural class. Table 3 shows the complete name of each biophysical setting as represented by the BpS code.

Table 2. HSI Scores by biophysical setting (BpS) and Structure class for caribou in the Ahtna Traditional Use Region of Southcentral Alaska.

BpS ¹	Sclass	Winter	Summer/Calving	BpS ¹	Sclass	Winter	Summer/Calving
16030	Α	0.25	0.25	16212	С	0.25	0.25
16030	В	0.25	0.50	16212	D	0.25	0.50
16030	С	0.10	0.10	16220	Α	0.25	0.25
16030	D	0.10	0.25	16220	В	0.25	0.25
16030	Ε	1.00	0.50	16220	С	1.00	0.50
16790	Α	0.25	0.25	16300	Α	0.25	0.25
16790	В	0.10	0.50	16300	В	0.25	0.25
16790	С	0.10	0.25	16300	С	0.25	0.50
16790	D	0.50	0.75	16102	Α	0.25	0.50
16790	Ε	0.75	0.50	16102	В	0.10	1.00
16041	Α	0.25	0.25	16280	Α	0.25	0.25
16041	В	0.25	0.50	16280	В	0.10	0.75
16041	С	0.10	0.10	16280	С	0.10	1.00
16041	D	1.00	0.25	16351	Α	0.50	1.00
16041	Ε	1.00	0.50	16310	Α	0.50	1.00
16042	Α	0.25	0.25	16290	Α	0.25	0.25
16042	В	0.25	0.50	16290	В	0.10	0.50
16042	С	0.10	0.10	16330	Α	0.50	1.00
16042	D	1.00	0.75	16110	Α	0.25	0.25
16042	Е	1.00	0.50	16120	Α	0.25	0.25
16011	Α	0.25	0.25	16080	Α	0.25	0.50
16011	В	0.10	0.50	16080	В	0.10	0.50
16011	С	1.00	1.00	16090	Α	0.25	0.50
16012	Α	0.25	0.25	16090	В	0.10	0.25
16012	В	0.10	0.50	16520	Α	0.25	0.25
16012	С	1.00	1.00	16520	В	0.10	0.10
16050	Α	0.25	0.25	16430	Α	0.50	1.00
16050	В	0.25	0.25	16170	Α	0.10	0.10
16050	С	0.10	0.10	16170	В	0.25	0.25
16050	D	0.10	0.10	16170	С	0.25	0.25
16050	Ε	0.10	0.25	16170	D	0.10	0.50
16070	Α	0.25	0.25	16170	Е	0.10	0.50
16070	В	0.10	0.10	16181	Α	0.25	0.25
16061	Α	0.25	0.25	16181	В	0.25	0.25
16061	В	0.25	0.25	16181	С	0.25	0.25
16061	С	0.10	0.10	16181	D	0.10	0.50
16061	D	0.10	0.10	16372	Α	0.25	0.25
16141	Α	0.25	0.25	16372	В	0.10	0.75
16141	В	0.25	0.25	16372	С	0.25	1.00
16141	С	0.10	0.25	16240	Α	0.10	0.50

BpS ¹	Sclass	Winter	Summer/Calving	BpS ¹	Sclass	Winter	Summer/Calving
16141	D	0.25	0.50	16481	Α	0.25	0.25
16141	Е	0.25	0.50	16481	В	0.10	0.10
16142	Α	0.25	0.25	16481	С	0.10	0.10
16142	В	0.25	0.25	16460	Α	0.25	0.25
16142	С	0.10	0.25	16460	В	0.10	0.10
16142	D	0.25	0.50	16460	С	0.10	0.10
16142	Е	0.25	0.50	16460	D	0.10	0.10
16150	Α	0.25	0.25	16440	Α	0.10	0.10
16150	В	0.25	0.25	16500	Α	0.25	0.25
16150	С	0.50	0.25	16500	В	0.25	0.25
16150	D	1.00	0.75	16500	С	0.10	0.10
16150	Ε	1.00	0.75	16500	D	0.10	0.10
16160	Α	0.25	0.25	16550	Α	0.25	0.25
16160	В	0.10	0.50	16550	В	0.25	0.25
16160	С	0.25	0.50	16550	С	0.10	0.10
16211	Α	0.50	0.25	16550	D	0.10	0.10
16211	В	0.50	0.25	16550	Е	0.10	0.10
16211	С	0.50	0.25	16590	Α	0.10	0.10
16211	D	0.25	0.75	16450	Α	0.25	0.25
16212	Α	0.25	0.25	16800	Α	0.50	0.25
16212	В	0.25	0.25	16800	В	0.10	0.25
				16620	Α	0.25	0.25

¹BpS codes

Table 3. Biophysical setting (BpS) setting names in the Ahtna Traditional Use Region of Southcentral Alaska.

BpS	EDM_Name
16011	Treeline White Spruce - Boreal
16012	Treeline White Spruce - SubBoreal
16030	White Spruce-Hardwood - Boreal
16041	Mesic Black Spruce - Boreal
16042	Mesic Black Spruce - SubBoreal
16050	Mesic Birch-Aspen
16061	Dry Aspen-Steppe Bluff
16070	Subalpine Balsam Poplar-Aspen
16080	Avalanche Slope Shrubland
16090	Mesic Subalpine Alder
16102	Mesic Scrub Birch/Willow
16110	Mesic Bluejoint Meadow
16120	Dry Grassland
16141	Montane Floodplain - Boreal
16142	Montane Floodplain - Subboreal

16150	Large River Floodplain	
16160	Riparian Stringer	
16170	Shrub and Herbaceous Floodplain	
16181	Herbaceous Fen	
16211	Dwarf Black Spruce Peatland - Boreal	
16212	Dwarf Black Spruce Peatland - Subboreal	
16220	Black Spruce Wet-Mesic Slope	
16240	Deciduous Shrub Swamp	
16280	Low Shrub-Tussock Tundra	
16290	Tussock Tundra	
16300	Wet Black Spruce-Tussock	
16310	Alpine Dwarf-Shrub Summit	
16330	Alpine Mesic Herbaceous Meadow	
16351	Alpine Ericaceous Dwarf-Shrubland	
16372	Alpine Floodplain	
16430	Alpine Dwarf Shrubland	
16440	Sitka Spruce	
16450	Alpine Mesic Herbaceous Meadow	
16460	Western Hemlock	
16481	Mountain Hemlock	
16500	Periglacial Woodland-Shrubland	
16520	Mesic Scrub Birch/Willow	
16550	Montane Floodplain - Subboreal	
16590	Mountain Hemlock Peatland	
16620	Shrub and Herbaceous Floodplain	
16790	White Spruce-Hardwood - SubBoreal	
16800	Avalanche Slope Shrubland	

²Structure classes are shown in the attached ecosystem diversity matrices.

Winter

Quality of each pixel or stand based on lichen (*Cladina arbuscula—mitis, C. rangiferina, C. stellaris, Flavocetraria cucullata, Cetraria islandica, Cladonia uncialis,* and *Stereocaulon* spp.) production categorized by Biophysical setting and structure class. Rated as 0.1, 0.25, 0.50, 0.75, or 1, with poor quality habitat ranked as 0.1.

- a. Questions or assumptions:
 - i. Should a variable be added for elevation, with winter habitat occurring below 1500m?
 - ii. Peatlands with moss from either burns or over-utilization of lichens are low quality while peatlands with lichens are high quality. These may not be capable of being mapped with remotely sensed information, so that a potential habitat quality may need a site visit to adjust.
 - iii. Should there be a variable that provides for snow depth adjustments- based on climate/physical settings or terrain features?
 - iv. At low elevations are mature conifer stands used as thermal cover?

v. At high elevations, burns will provide good summer/calving habitat, but will reduce winter habitat- so a mosaic of the two is desirable with greater needs for winter habitat than summer habitat. At lower elevations, burns will improve moose habitat and increase wolf densities, so do not provide for quality summer habitat- large blocks of unburned areas needed for winter and/or summer habitat at lower elevation biophysical settings.

Caribou Calving

BPS/structure class rating of each category for calving with mature conifer preferred at low elevations, as well as lichen tundra areas at higher elevations and large fens/bogs. Areas of lower value with poor quality are ranked at 0.1. Categories will be rated as: 0.1, 0.25, 0.50, 0.75, or 1, with high quality for landscape analyses considered 75 or higher.

Summer Habitat

BPS/structural class rating of each category for summer habitat in terms of providing desired grasses, forbs and lichens at higher elevations. Categories will be rated as 0.1, 0.25, 0.50, 0.75, or 1.

Disturbance Evaluation

Human disturbances will be mapped across the range of caribou. Each human disturbance will be buffered by an assigned width of effect. This width will be divided into 4 bands with the closest band receiving a reduction of 90%, the second band a reduction of 75%, the third band a reduction of 50%, and the 4th band a reduction of 25%. Disturbance effect distances are listed in Table 4.

Vistnes and Nelleman (2008) reported that 87% of studies looking at caribou responses to disturbance from the 1980's found that human features reduced caribou use by 50-95% within 5 km. Leblond et al. (2014) identified effects of mines, paved roads, and forest roads on caribou in Alberta. They presented the following relationships:

Distance to paved road = -3E-06x2 + 0.0343x + 8.2524

Distance to forest road = 2E-09x3 - 2E-05x2 + 0.0632x + 21.886

Distance to mine = -2E-06x2 + 0.0279x + 5.9313

These equate to effects out to 2500 m on a curvilinear relationship for forest roads and out to 5000 m for paved roads.

Losier et al. (2015) reported higher mortality of caribou cows when then in home ranges that contained 6-20 year old clearcuts, and identified a threshold of >7km that reduced these effects in Quebec. They identified an increasing percentage of clearcuts as a key contributor through indirect habitat loss associated with increases in moose and wolf densities.

These findings help to identify effect distances for impacts to caribou habitat shown in Table 5. It appears that 5km is an appropriate maximum effect distance. A linear response is assumed, although a curvilinear response which increases closer to the disturbance may also be occurring. However, without better data to support such a relationship, the relationship used in this model designates impacts in bands of distance with higher impacts assigned closer to the anthropogenic disturbance.

Table 4. Winter and Calving disturbance distances (linear decrease with proximity to disturbance) for caribou.

Disturbance type	Effective distance
Major road	500 m
Forest or minor road or motorized trail	100 m

Development- towns, etc.	500 m
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Additional Habitat Considerations

The Chisana herd is non-migratory, so connectivity between seasonal ranges may not be an issue. The Nelchina herd has seasonal movements. Should an analysis of potential movement barriers that can occur between seasonal ranges be added?

Model Results

Figures 3-6 display the outputs of the caribou habitat model for the Ahtna Traditional Use Territory.

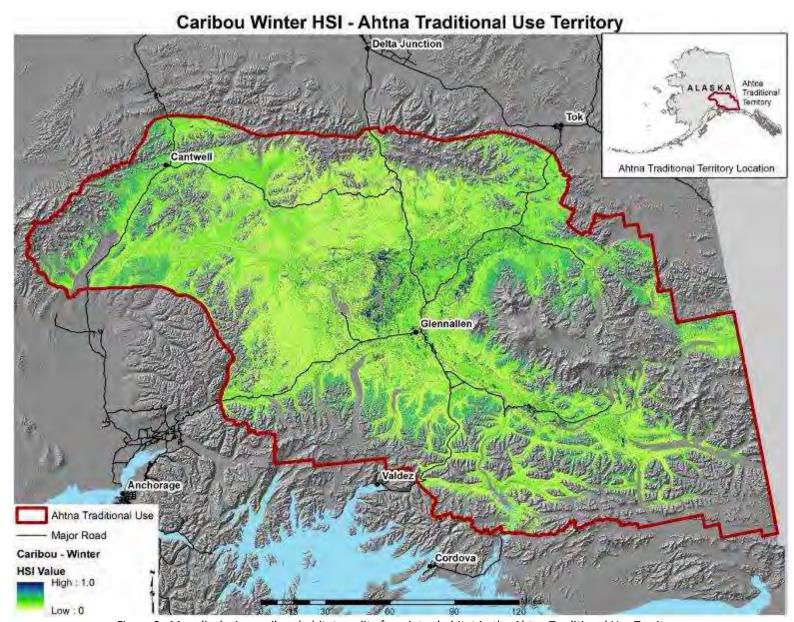


Figure 3. Map displaying caribou habitat quality for winter habitat in the Ahtna Traditional Use Territory.

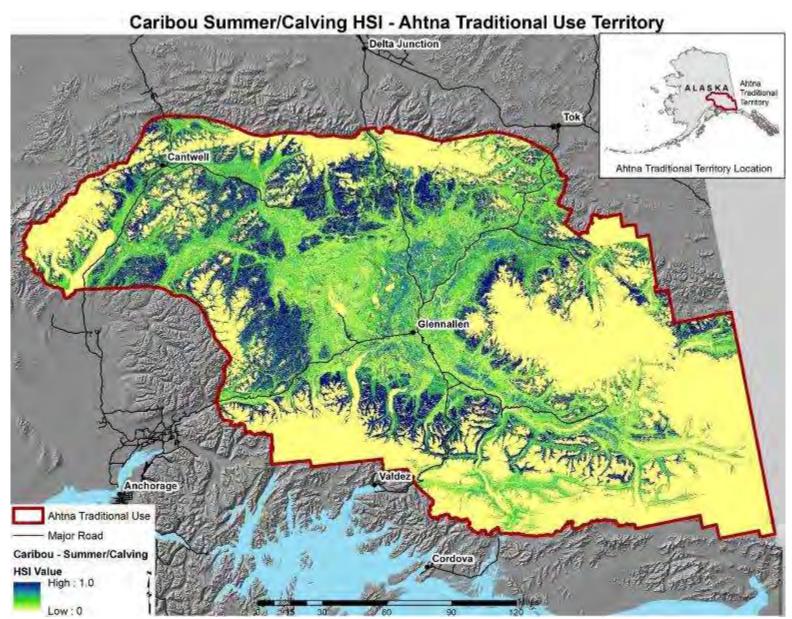


Figure 4. Map displaying caribou habitat quality for summer/calving habitat in the Ahtna Traditional Use Territory.

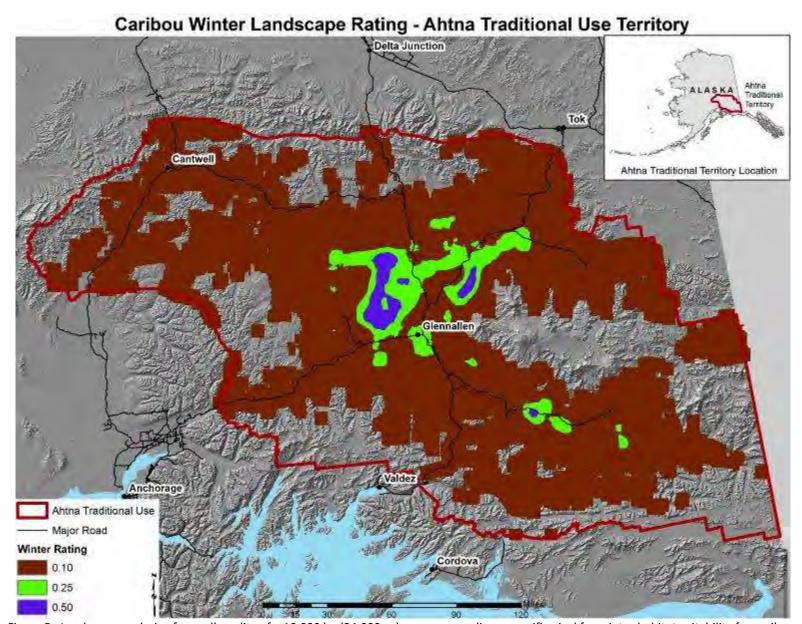


Figure 5. Landscape analysis of overall quality of a 10,000 ha (24,000 ac) area surrounding a specific pixel for winter habitat suitability for caribou.

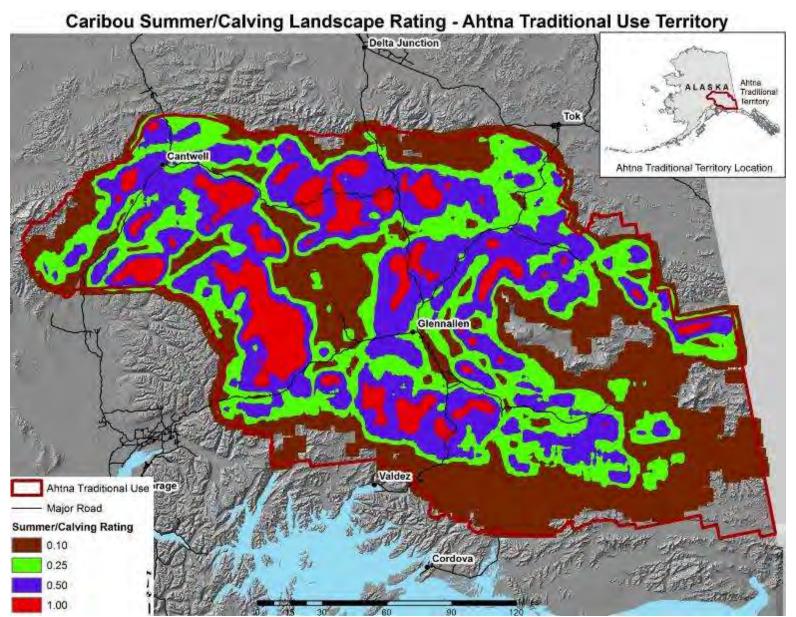


Figure 6. Landscape analysis of overall quality of a 10,000 ha (24,000 ac) area surrounding a specific pixel for summer/calving habitat suitability for caribou.

Literature Cited

- Adam, R. C. J, S. Boutin, and D. M. Hebert. 2004. Spatial separation of caribou from moose and its relation to predation by wolves. Journal of Wildlife Management 68:799-809.
- Adams, L. G., F. J. Singer, and B. W. Dale. 1995. Caribou calf mortality in Denali, National Park, Alaska. Journal of Wildlife Management 59:584–594.
- Alaska Department of Fish and Game. 2011. Caribou management report of survey-inventory activities 1 July 2008- 30 June 2010. P. Harper, editor. Juneau, Alaska, USA.
- Anderson, R. B. 1999. Peatland habitat use and selection by woodland caribou (*Rangifer tarandus tarandus*) in Northern Alberta. Thesis, University of Alberta. Edmonton, Canada.
- Apps, C. D., and B. N. McLellan. 2006. Factors influencing the dispersion and fragmentation of endangered mountain caribou populations. Biological Conservation 130:84-97.
- Ballard, W. B. 1994. Effects of black bear predation on caribou—a review. Alces 30:25–35.
- Barrier, T. A., and C. J. Johnson. 2012. The influence of fire history on selection of foraging sites by barren-ground caribou. Ecoscience 19:177-188.
- Barten, N. L., R. T. Bowyer, K. J. Jenkins. 2001. Habitat use by female caribou: Tradeoffs associated with parturition. Journal of Wildlife Management 65:77-92.
- Beauchesne, D., J. A. G. Jaeger, and M. H. St-Laurent. 2013. Disentangling woodland caribou movements in response to clearcuts and roads across temporal scales. Plos One 8(11).
- Bentzen, T. W. 2011. Unit 12 caribou. Pages 60-73 in: P. Harper, editor. Caribou management report of survey inventory activities 1 July 2008- 30 June 2010. Alaska Department of Fish and Game. Juneau.
- Bergerud, A. T. 1996. Evolving perspectives on caribou population dynamics, have we got it right yet? Rangifer 9:95-116.
- Bergerud, A. T. 2007. The need for the management of wolves- an open letter. Rangifer 17:39-50.
- Bergerud, A. T., H. E. Butler, and D. R. Millar. 1984. Anti-predator tactics of calving caribou: dispersion in mountains. Canadian Journal of Zoology 52:1566-1575.
- Bergerud, A. T., and J. P. Elliot. 1986. Dynamics of caribou and wolves in British Columbia. Canadian Journal of Zoology 64:1515–1529.
- Bergerud, A. T., R. Ferguson, and H. E. Butler. 1990. Spring migration and dispersion of woodland caribou at calving. Animal Behaviour 39:360–368.
- Bergerud, A. T., and R. E. Page. 1987. Displacement and dispersion of parturient caribou at calving as antipredator tactics. Canadian Journal of Zoology 65:1597–1606.
- Boan, J. J., J. R. Malcolm, and B. E. McLaren. 2014. Forest overstory and age as habitat? Detecting the indirect and direct effects of predators in defining habitat in a harvested boreal landscape. Forest Ecology and Management 326:101-108.
- Bos, G. N. 1975. A partial analysis of the current population status of the Nelchina caribou herd.

 Proceedings of the First International Reindeer and Caribou Symposium. University of Alaska, Fairbanks, USA, 9-11 August 1972.
- Bradshaw, C. J. A., D. M. Hebert, B. Rippin, and S. Boutin. 1995. Winter peatland habitat selection by woodland caribou in Northeastern Alberta. Canadian Journal of Zoology 73:1567-1574.
- Briand, Y., J. P. Ouellet, C. Dussault, and M. H. St-Laurent. 2009. Fine-scale habitat selection by female forest-dwelling caribou in managed boreal forest: Empirical evidence of a seasonal shift between foraging opportunities and antipredator strategies. Ecoscience 16:330-340.
- Cenovus FCCL Ltd. 2010. Narrows Lake Project. Appendix 5-V. Wildlife habitat modeling. (accessed June 4, 2015).
- Collins, W. B., B. W. Dale, L. G. Adams, D. E. Mcelwain, and K. Joly. 2011. Fire, grazing history, lichen abundance, and winter distribution of caribou in Alaska's taiga. Journal of Wildlife Management 75:369-377.
- Courtois, R., J. P. Ouellet, L. Breton, A. Gingras, C. Dussault. 2007. Effects of forest disturbance on density, space use, and mortality of woodland caribou. Ecoscience 14:491-498.
- Cumming, H. G, D. B. Beange, and G. Lavoie. 1996. Habitat partitioning between woodland caribou and

- moose in Ontario: the potential role of shared predation risk. Rangifer 9:81–94.
- Dale, B. W., L. G. Adams, and R. T. Bowyer. 1994. Functional response of wolves preying on barrenground caribou in a multiple-prey ecosystem. Journal of Animal Ecology 63:644–652.
- Dyer, S. J., J. P. O'Neill, S. M. Wasel, and S. Boutin. 2001. Avoidance of industrial development by woodland caribou. Journal of Wildlife Management 65:531–542.
- Dyer, S. J., J. P. O'Neill, S. M. Wasel, and S. Boutin. 2002. Quantifying barrier effects of roads and seismic lines on movements of female woodland caribou in northeastern Alberta. Canadian Journal of Zoology 80:839-845.
- Dussault, C., V. Pinard, J. P. Ouellet, R. Courtois, and D. Fortin. 2012. Avoidance of roads and selection for recent cutovers by threatened caribou: fitness-rewarding or maladaptive behaviour? Proceedings of the Royal Society B 279:4481-4488.
- Edmonds, E. J., and M. Bloomfield. 1984. A study of woodland caribou (*Rangifer tarandus caribou*) in west-central Alberta, 1979-1983. Alberta Energy and Natural Resources Fish and Wildlife Division, Edmonton, Alberta, Canada.
- Ferguson, S. H., A. T. Bergerud, and R. Ferguson. 1988. Predation risk and habitat selection in the persistence of a remnant caribou population. Oceologia 76:236-245.
- Fortin, D., P. L. Buono, A. Fortin, N. Courbin, C. T. Gringas, P. R. Moorcroft, R. Courtois, and C. Dussault. 2013. Movement responses of caribou to human-induced habitat edges lead to their aggregation near anthropogenic features. American Naturalist 181:827-836.
- Gasaway, W. C., R. O. Stephenson, J. L. Davis, P. E. K. Shepherd, and O. E. Burris. 1983. Interrelations of wolves, prey, and man in interior Alaska. Wildlife Monographs 84:1-50.
- Gustine, D. D., K. L. Parker, R, J. Lay, M. P. Gillingham, and D. C. Heard. 2006. Calf survival of woodland caribou in a multi-predator ecosystem. Wildlife Monographs 165:1-32.
- Heard, D. C., T. M. Williams, and D. A. Melton. 1996. The relationship between food intake and predation risk in migratory caribou and implications to caribou and wolf population dynamics. Rangifer 9:37–44.
- Hemming, J. E. 1971. The distribution movement patterns of caribou in Alaska. Technical Bulletin 1. Alaska Department of Fish and Game, Juneau.
- James, A. R. C., S. Boutin, D. M. Hebert, and A. B. Rippin. 2004. Spatial separation of caribou from moose and its relation to predation by wolves. Journal of Wildlife Management 68:799-809.
- James, A. R. C., and A. K. Stuart-Smith. 2000. Distribution of caribou and wolves in relation to linear corridors. Journal of Wildlife Management 64:154-159.
- Johnson, C. J., N.D. Alexander, R. D. Wheate, and K. L. Parker. 2003. Characterizing woodland caribou habitat in sub-boreal and boreal forests. Forest Ecology and Management 180:241-280.
- Johnson, C. J. and D. E. Russell. 2014. Long-term distribution responses of a migratory caribou herd to human disturbance. Biological Conservation 177:52-63.
- Joly, K., B. W. Dale, W. B. Collins, and L. G. Adams. 2003. Winter habitat use by female caribou in relation to wildland fires in interior Alaska. Canadian Journal of Zoology 81:1192-1201.
- Joly, K., F. S. Chapin, and D. R. Klein. 2010. Winter habitat selection by caribou in relation to lichen abundance, wildfire, grazing, and landscape characteristics in Northwest Alaska. Ecoscience 17:321-333.
- Jones, E. S., M.P. Gillingham, D. R. Seip, and D. C. Heard. 2007. Comparison of seasonal habitat selection between threatened woodland caribou ecotypes in central British Columbia. Rangifer 17:111-128.
- Latham, A. D. M., M. C. Latham, M. S. Boyce, and S. Boutin. 2011a. Movement responses by wolves to industrial linear features and their effect on woodland caribou in northeastern Alberta. Ecological Applications 21:2854-2865.
- Latham, A. D. M., M. C. Latham, N. A. McCutchen, and S. Boutin. 2011b. Invading white-tailed deer change wolf-caribou dynamics in northeastern Alberta. Journal of Wildlife Management 75:204-212.
- Latham, A. D. M., M. C. Latham, M. S. Boyce, and S. Boutin. 2013. Spatial relationships of sympatric wolves (Canis lupus) and coyotes (C. latrans) with woodland caribou (Rangifer tarandus caribou) during the calving season in a human-modified boreal landscape. Wildlife Research 40:250-260.
- Leblond, M., J. Frair, D. Fortin, C. Dussault, J. P. Ouellet, and R. Courtois. 2011. Assessing the influence

- of resource covariates at multiple spatial scales: an application to forest-dwelling caribou faced with intensive human activity. Landscape Ecology 10:1433-1446.
- Leclerc, M., C. Dussault, and M. H. St-Laurent. 2012. Multiscale assessment of the impacts of roads and cutovers on calving site selection in woodland caribou. Forest Ecology and Management 286:59-65.
- Lesmerises, R., J. P. Ouellet, C, Dussault, and M.H. St Laurent. 2013. The influence of landscape matrix on isolated patch use by wide-ranging animals: conservation lessons for woodland caribou. Ecology and Evolution 3:2880-2891.
- Lieb, J. W. 1994. Analysis of Nelchina Caribou Range III. Alaska Department of Fish and Game, Division of Wildlife Conservation, Research Progress Report 1 July 1989-30 June 1990, Juneau, USA.
- Lieb, J. W., K. W. Pitcher, and R. W. Tobey. 1988. Optimum population size for the Nelchina Caribou Herd? Pages 135–145 in R. D. Cameron and J. L. Davis, editors. Proceedings of the Third North American Caribou Workshop, Chena Hot Springs, USA, 4-6 November, 1987.
- Losier, C. L. S. Courturier, M. H. St.-Laurent, P. Drapeau, C. Dussault, T. Rudolph, V. Broduer, J. A. Merkle, and D. Fortin. 2015. Adjustments in habitat selection to changing availability induce fitness costs for a threatened ungulate. Journal of Applied Ecology 52:496-504.
- Mager, K. H., K. E. Colson, P. Groves, and K. J. Hundertmark. 2014. Population structure over a broad spatial scale driven by nonanthropogenic factors in a wide-ranging migratory mammal, Alaskan caribou. Molecular Ecology 23:6045-6057.
- Mahoney, S. P., and J. A. Virgl. 2003. Habitat selection and demography of a nonmigratory woodland caribou population in Newfoundland. Canadian Journal of Zoology 81:321–334.
- Manitoba Model Forest. 1995. Report on the Manitoba Model Forest integrated forestry/woodland caribou management strategy: Volume 1: Maintaining our options. Manitoba Model Forest and TEAM. Pine Falls, MB, Canada.
- O'Brien, D., M. Manseau, A. Fall, and M. J. Fortin. 2006. Testing the importance of spatial configuration of winter habitat for woodland caribou: An application of graph theory. Biological Conservation 130:70-83.
- Pinard, V., C. Dussault, J. P. Ouellet, D. Fortin, and R. Courtois. 2012. Calving rate, calf survival rate, and habitat selection of forest-dwelling caribou in a highly managed landscape. Journal of Wildlife Management 76:189-199.
- Racey, G. D., and A. A. Arsenault. 2007. In search of a critical habitat concept for woodland caribou, boreal population. Rangifer 17:29-37.
- Rettie, W. J., and F. Messier. 2000. Hierarchical habitat selection by woodland caribou: its relationship to limiting factors. Ecography 23:466-478.
- Robinson, H. S., M. Hebblewhite, N. J. DeCesare, J. Whittington, L. Neufeld, M. Bradley, and M. Musiani. 2012. The effect of fire on spatial separation between wolves and caribou. Rangifer 32:277-294.
- Schindler, D. W., D. W. Walker, T. Davis, and R. Westwood. 2007. Determining effects of an all weather logging road on winter woodland caribou habitat use in south-eastern Manitoba. Rangifer 17:209-217.
- Schneider, R. R., B. Wynes, S. Wasel, E. Dzus, and M. Hiltz. 2000. Habitat use by caribou in northern Alberta, Canada. Rangifer 20:43-50.
- Schwanke, R. A. 2011. Units 13 and 14B caribou management report. Pages 90-108 in P. Harper, editor. Caribou management report of survey-inventory activities 1 July 2008- 30 June 2010. Alaska Department of Fish and Game. Juneau, USA.
- Seip, D. R. 1992. Factors limiting woodland caribou populations and their interrelationships with wolves and moose in Southeastern British Columbia. Canadian Journal of Zoology 70:1494-1503.
- Siniff, D. B., and R. O. Skoog. 1964. Aerial censusing of caribou using stratified random sampling. Journal of Wildlife Management 28:391–401.
- Smith, K. G., E. J. Ficht, D. Hobson, T. C. Sorensen, and D. Hervieux. 2000. Winter distribution of woodland caribou in relation to clear-cut logging in west-central Alberta. Canadian Journal of Zoology 78:1433–1440.
- Stuart-Smith, A. K., C. J. A. Bradshaw, S. Boutin, D. M. Hebert, and A. B. Rippin. 1997. Woodland caribou relative to landscape patterns in northeastern Alberta. Journal of Wildlife Management 61:622-633.
- Thompson, I. D., P. A. Wiebe, E. Mallon, A. R. Rodgers, J. M. Fryxell, J. A. Baker, and D. Reid. 2015.

- Factors influencing the seasonal diet selection by woodland caribou (*Rangifer tarandus tarandus*) in boreal forests in Ontario. Canadian Journal of Zoology 93:87-98.
- Valkenburg, P., M. A. Keech, R. A. Sellers, R. W. Tobey, and B. W. Dale. 2002. Investigation of regulating and limiting factors in the Delta caribou herd. Federal Aid in Wildlife Restoration Final Report, Grant W-27-5, Project 3.42. Alaska Department of Fish and Game, Juneau, USA.
- Vistnes, I., and C. Nellemann. 2008. The matter of spatial and temporal scales: a review of reindeer and caribou response to human activity. Polar Biology 31:399–407.
- Vors, L. S., J. A. Schaefer, B. A. Pond, A. R. Rodgers, and B. R. Patterson. 2007. Woodland caribou extirpation and anthropogenic landscape disturbance in Ontario. Journal of Wildlife Management 71:1249-1256.
- Weckworth, B. V., M. Musiani, A. D. McDevitt, M. Hebblewhite, and S. Mariani. 2012. Reconstruction of caribou evolutionary history in Western North America and its implications for conservation. Molecular Ecology 21:3610-3624.
- Wittmer, H. U., A. R. E. Sinclair, and B. N. McLellan. 2005. The role of predation in the decline and extirpation of woodland caribou. Oecologia 144:257-267.
- Wittmer, H. U., B. N. McLellan, R. Serrouya, and C. D. Apps. 2007. Changes in landscape composition influence the decline of a threatened woodland caribou population. Journal of Animal Ecology 76:568-579.
- Young, D. D., and T. R. McCabe. 1997. Grizzly bear predation rates on caribou calves in northeastern Alaska. Journal of Wildlife Management 61:1056–1066.
- Zittlau, K., J. Coffin, R. Farnell, G. Kuzyk, and C. Strobeck. 2000. Genetic relationships of the Yukon woodland caribou herds determined by DNA typing. Rangifer 12:59-62.

Appendix C. Proposed Improvement Area Descriptions

Cantwell Site Improvements

<u>Carlos Creek:</u> This treatment stand encompasses an old gravel pit that has regenerating willow and poplar which have escaped browsing and are now too tall to be effectively browsed by moose. Cutting and or roller chopping would allow moose to utilize the available browse. The site is also well drained and removing some of the overstory would allow for additional seedling establishment of preferred hardwood species.

<u>Intertie #1 & #2:</u> Good sites along old power line right of way near the Nenana River that has been suggested by the Village Council. There is a white spruce overstory with willow and some cottonwood and poplar in the understory. Wet site which would likely require winter treatment. Primarily a moose browse improvement site but does have the potential to harvest some white spruce.

<u>Jack Canyon:</u> This site is located in the flood plain of Jack Creek. It is heavily covered with large willows that are overgrown and mostly too large to provide moose browse. It is located in important winter range and receives heavy moose use. Access to this parcel is across a Native Allotment which would require permission from the owner.

<u>Slime Creek:</u> Site is located on old gravel pit between Parks Highway and Nenana River. The site has scattered white spruce and poplar with heavy cover of willow and alder. Crushing the willow will allow for the regeneration of moose browse.

<u>Transfer Site:</u> Site is located on old gravel pit between Parks Highway and Nenana River. The site has an overstory of poplar with heavy cover of willow and alder. Crushing the willow and cutting the poplar will allow for the regeneration of moose browse.

Chistochina Site Improvements

<u>Airstrip:</u> Appears to be an old gravel pit that is now used as a community shooting range and dumping area. Appears to receive heavy moose use, particularly in the winter. More heavily disturbed areas have good amounts of willow regeneration and the surrounding stand consists of aspen and poplar. Roller chopping could provide firewood for local community and improve moose browse.

<u>Aspen #1 & #2:</u> White spruce overstory in these stands with scattered popular and occasional willow. Extensive stands of buffaloberry (*Sheperdia canadensis*) in the understory. Stand has good potential for precommerical thinning of spruce and poplar.

Aspen #3, #4, & #5: Increased amounts of willow, aspen, and poplar compared to stands #1 and #2. These sites were previously treated 10+ years ago and regrowth has reached a point where it is escaping moose browsing pressure. Roller chopping would return these stands to productive condition for moose browse.

<u>Mile 26:</u> Stand overstory primarily consists of mature aspen and poplar with scattered white spruce. Stand would be improved for moose browse by overstory removal. Also possible to generate pulp from aspen and poplar.

Roadside: This site has mature white spruce and poplar near the highway and transitions to black spruce and sphagnum moss as the aspect changes. Scattered willow through the stand and present in both timber types. This site has the potential for both timber and moose browse improvements.

Gakona Site Improvements

North Trail #1, #2, & #3: This stand is mostly a timber improvement site with good numbers of white spruce. Moose browse could be enhanced by crushing willows along an existing roadway.

Old Pit #1 & #2: Heavy beetle kill among white spruce, especially in #1. Timber improvement treatment needed to remove beetle kill spruce in #1 and thin white spruce in #2.

<u>Radio Tower #1 & #2:</u> Good willow growth, particularly on #1. Would benefit from crushing to promote regrowth and improve moose browse. These sites are located in potential caribou habitat which should be a factor when considering treatments.

<u>Sanford Trail #1, #2, & #3:</u> These sites all have good stands of white spruce and would benefit from timber improvement. The stands are merchantable and would continue to be productive for years if managed properly.

<u>Swimming Hole:</u> Site is located around an active gravel pit. It contains got amounts of willow and poplar with some alder. Site would benefit from crushing to bring browse back to a level that would allow moose to utilize it.

Mentasta Site Improvements

<u>Carlson Lake:</u> This site is located along a 17B easement. Overstory of white spruce with scattered shrubs. There is long term potential to manage the white spruce and possibly improve the amount of willow over time.

<u>Little Tok #1:</u> This is an excellent site with large amounts of willow that would benefit from being crushed in order to improve the quality of moose browse. This site is adjacent to a shareholder moose camp which further increases its value for treatment.

<u>Little Tok #2 & #3:</u> White spruce stand that has been previously harvested. These stands would benefit from continued timber management with periodic entry to thin spruce and then eventually harvest the overstory.

<u>Mile 100 #1:</u> Site has heavy cover of young white spruce. It would benefit from a precommercial thin and continued management to insure good timber production. There are some willow and aspen in the understory as well.

<u>Mile 100 #2 & #3:</u> These stands are primarily aspen and willow with some young white spruce in the understory. They would benefit from crushing of the willow and aspen to improve moose browse quality. In addition there is a shareholder moose camp near #3.

<u>Mile 85:</u> Unable to access stand, but appears to be a good candidate for a moose browse improvement. Located around an old gravel pit.

<u>Old Mentasta:</u> Excellent site for timber improvement. Very good white spruce stand with mature trees in overstory. Located along 17B easement. Contact adjacent landowners before beginning work.

Gulkana Site Improvements

Beaver Dam: Excellent site that is set back from the highway about 400 meters. It consists of islands of aspen and white spruce among black spruce wetlands. There is heavy willow use by moose. It would benefit from treatments to enhance browse quality.

<u>Gulkana Gravel Pit:</u> Very good site for treatments behind the townsite of Gulkana. Great access for shareholders to the Copper riverbottom. High willow density in places with occasional dense stands of white spruce and scattered poplar. Excellent place to enhance moose browse and provide biomass for Gulkana.

Copper Pit #1 & #2: Good sites behind locked gate located along active gravel pit in Copper River floodplain. Site would benefit greatly from overstory removal of white spruce and crushing of hardwoods to provide biomass and increase the quality of moose browse.

<u>Gulkana TAPS #1, #2, & #3:</u> Good sites with mix of white spruce and aspen. Also have willow and poplar in the understory. Treatment would benefit moose browse and TAPS provides access to stand. Hauling distance might be a little far to take advantage of biomass.

<u>Gulkana TAPS #4, #5, & #6:</u> Good sites with mix of white spruce and aspen. Also has willow and poplar in the understory. Treatment would benefit moose browse and TAPS provides access to stand. Hauling distance might be a little far to take advantage of biomass.

Highway East #1 & #2: Access to these stands is controlled by Ray Ewan (but they are located on Ahtna Land). We were not able to survey the stands but they have a high potential to be good treatment sites. Pipeline Access #1 & #2: These are both ideal stands for both browse treatments and biomass production. They are close to Gulkana which reduces trucking times and also allows shareholder access. In addition, they are located behind a locked gate which reduces trespass. Consist mostly of hardwoods with white spruce in understory and some mature white spruce. Willow scattered throughout.

Tower Road #1 & #2: These are both ideal stands for both browse treatments and biomass production. They are close to Glennallen which reduces trucking times and also allows shareholder access. In addition, they are located behind a locked gate which reduces trespass. There is a possible access issue due to the road leading to FAA equipment. The stand is a mix of white spruce, aspen, and poplar. Willow scattered throughout.

Chitina Site Improvements

<u>Mile 3 #1 & #2:</u> Decent sites for moose browse improvement. Site currently consists of tall willows and alders with young white spruce in understory. Crushing would revive browse and improve quality and accessability.

Kluti-Kaah Site Improvements

<u>CC Airstrip:</u> Good stand for browse enhancement and biomass production. Access is good along the Copper Center airstrip. Stand consists of mature white spruce and aspen with a willow understory.

<u>Mile 92 Pit:</u> Nice site located behind a gravel pit. Access is restricted due to a locked gate which makes it ideal for discouraging trespass. Site consists of mixed white spruce and aspen overstory with willow understory. Good site for biomass production, timber harvest, and browse enhancement.

<u>Mile 98.5:</u> Nice site located along a pipeline access road and the TAPS. Access is restricted due to a locked gate which makes it ideal for discouraging trespass. Site consists of young white spruce with some aspen and a lot of willow in the understory. Great site for browse enhancement with some biomass.

<u>Old Edgerton #2:</u> Really nice site that is heavy to aspen. There are some white spruce and poplar in the overstory as well. Scattered willow in the understory. Good stand for browse enhancement and some timber harvest possible.

<u>Willow Lake:</u> Excellent site on the other side of TAPS from the highway. Provides locked access, but there is a passage under TAPS to allow equipment to access site. Old gravel pit with good density of felt-leaf willow along with poplar, aspen, and young spruce. Perfect location for browse enhancement.

Tazlina Site Improvements

<u>Ahtna Office #1</u>: Access ends at the beginning of this unit. The site is almost too wet, but would be fine for winter treatments. There is 20-30% cover of spruce with a lot of willow in the understory. Recommend harvesting spruce and knocking down willows to improve moose browse.

<u>Airport #1 & #2:</u> Good mix of species with very mature aspen, medium sized white spruce, some balsam poplar and a variety of willow species. Excellent access with good road. Recommend harvesting overstory and treating willows to improve moose browse.

<u>Fisher's Pit #1 & #2:</u> Mix of harvested and unharvested timber. Pockets of 100% white spruce with pockets of aspen and white spruce understory. Extremely variable stand. Recommend harvesting mature spruce and aspen, thinning younger spruce, and treating willows to improve moose browse.

North Fireline #1, #2, & #3: These were recommended as good sites for moose browse treatments by Sarah Daszkiewicz.

<u>Taz West Trails #5:</u> This may be a good site. Due to the soft ground it is only suitable for a winter treatment. There is a mix of stunted black spruce and taller black spruce with mixed willow in the understory. Recommend crushing or cutting spruce and encouraging willow regeneration/regrowth for improving moose browse.

<u>Taz West Trails #6:</u> Good access off well pad road. Mix of spruce with a few hardwoods. There is also some beetle killed spruce. Variety of willow species in the understory. Recommend harvesting overstory and treating willows to improve moose browse.

<u>Tazlina Fireline #1, #4, #5, & #6:</u> These stands are all adjacent to areas that have been previously harvested. Portions of them may have been harvested as well. They consist of medium white spruce and mature aspen. They all have willow present in the understory. Suitable treatments would include harvesting the aspen and some spruce and then treating the willow to improve moose browse.

<u>Tazlina Fireline #2:</u> This stand consists of dense medium white spruce. It would benefit from a timber improvement treatment to thin the spruce and allow the remaining trees to increase growth rates.

<u>Tazlina Fireline #3:</u> This stand has been harvested and seen significant regrowth of aspen, spruce, and willow. It would benefit from roller chopping to improve the moose browse.

<u>Tazlina Logging Road #1:</u> Stand consists of aspen and white spruce. A precommercial thinning would be an excellent treatment to improve the production of this stand. It would also increase the quality of moose browse.

<u>Tazlina Logging Road #2:</u> This is a closed stand and is overstocked with spruce. Thinning would allow the remaining spruce to increase production. It is an excellent candidate for a timber stand improvement.

<u>Tazlina Logging Road #3:</u> Previously harvested stand in places with a lot of aspen regeneration. More white spruce could be removed and the aspen could be treated to increase moose browse.

<u>Tazlina Pit:</u> Treatment area surrounds a gravel pit with gated access road. Excellent location to treat the willows surrounding the gravel pit and provide a harvest location for shareholders.

<u>Tazlina TAPS North #1:</u> Beetle killed white spruce in overstory that should be removed. A lot of willow in the understory that could be treated to improve moose browse. Some poplar and aspen. Along pipeline so there could be access difficulties.

<u>Tazlina TAPS North #2 & #3:</u> White spruce overstory with some poplar and aspen. Decent amount of willow in the understory that could be treated to improve moose browse. Along pipeline so there could be access difficulties.

<u>Terrace Drive:</u> This would make an excellent demonstration area for a timber stand improvement due to proximity to town. It is mostly mature spruce with 40% canopy cover. Potentially thin now, then return and harvest in approximately 35 years.